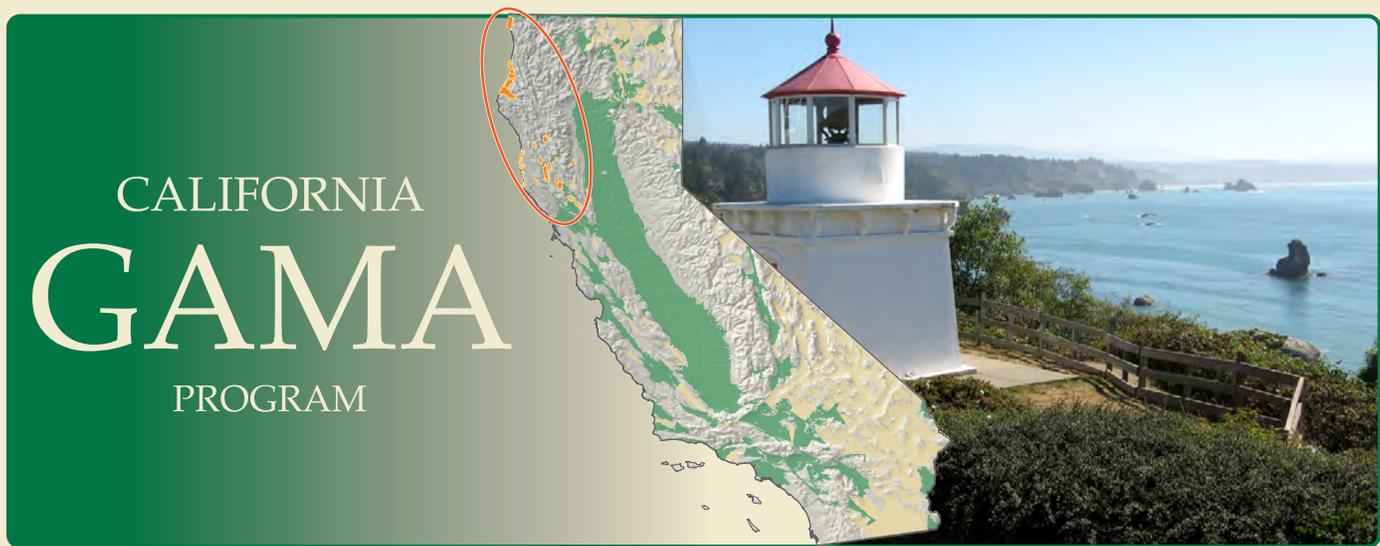


Prepared in cooperation with the California State Water Resources Control Board and the U.S. Geological Survey National Water-Quality Assessment Program

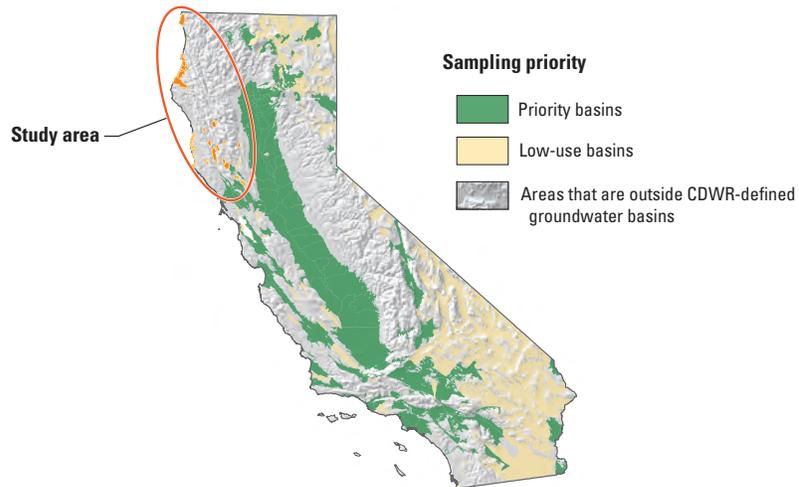
A product of the California Groundwater Ambient Monitoring and Assessment (GAMA) Program

Status and Understanding of Groundwater Quality in the Northern Coast Ranges Study Unit, 2009: California GAMA Priority Basin Project



Scientific Investigations Report 2014–5215

Front Cover Map: Groundwater basins categorized by sampling priority. Location of groundwater basin boundaries from California Department of Water Resources (CDWR, 2003).



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Status and Understanding of Groundwater Quality in the Northern Coast Ranges Study Unit, 2009: California GAMA Priority Basin Project

By Timothy M. Mathany and Kenneth Belitz

A product of the California Groundwater Ambient Monitoring and Assessment (GAMA) Program

Prepared in cooperation with the California State Water Resources Control Board and the U.S. Geological Survey National Water-Quality Assessment Program

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Contents

Abstract.....	1
Introduction	2
Purpose and Scope	4
Hydrogeologic Setting of the Northern Coast Ranges Study Unit.....	5
Northern Coast Ranges—Interior Basins Study Area	5
Northern Coast Ranges—Coastal Basins Study Area	9
Methods.....	10
Status Assessment.....	10
Relative-Concentrations and Water-Quality Benchmarks	10
Datasets Used for Status Assessment.....	11
USGS-GAMA Grid Sites	11
Additional Data Used for Spatially Weighted Calculation	11
Selection of Constituents for Additional Evaluation in the Status Assessment.....	13
Calculation of Aquifer-Scale Proportions.....	15
Understanding-Assessment Methods	16
Statistical Analysis	17
Potential Explanatory Factors	17
Land Use.....	17
Well Construction Information.....	20
Hydrologic Conditions	20
Groundwater Age.....	20
Geochemical Conditions.....	23
Correlations Between Explanatory Factors	23
Status and Understanding of Water Quality.....	28
Inorganic Constituents.....	28
Inorganic Constituents with Health-Based Benchmarks	28
Trace Elements.....	28
Understanding Assessment for Boron	32
Understanding Assessment for Arsenic	39
Nutrients.....	42
Uranium and Other Radioactive Constituents.....	42
Inorganic Constituents with Aesthetic-Based Benchmarks	42
Understanding Assessment for Iron and Manganese	42
Organic Constituents with Health-Based Benchmarks	45
Volatile Organic Compounds.....	47
Understanding Assessment for Chloroform	49
Pesticides and Pesticide Degradates	51
Understanding Assessment for Simazine.....	51
Special-Interest Constituent with Health-Based Benchmark.....	53
Understanding Assessment for Perchlorate.....	53

Contents—Continued

Summary.....	54
Acknowledgments.....	55
References.....	55
Appendix A. Map of the Northern Coast Ranges Study Unit	65
Appendix B. Ancillary Datasets	71
Appendix C. Calculation of Aquifer-Scale Proportions.....	81
Appendix D. Comparison of CDPH and GAMA Priority Basin Data	83
Appendix E. Additional Water-Quality Data	85

Figures

1. Map showing location of the Northern Coast Ranges study unit and the California hydrogeologic provinces, California GAMA Priority Basin Project.....	3
2. Map showing location of the Coastal Basins and Interior Basins study areas, major cities and towns, major roads, and hydrologic features in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	6
3. Maps showing geologic formations and locations of the Coastal Basins study area and Interior Basins study area, locations of hot springs and geothermal wells, and approximate boundaries of the Napa and Sonoma Valleys and the Geysers Steam Field, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	7
4. Map showing locations of study area grid cells, USGS-GAMA grid sites, and CDPH sites, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	12
5. Map showing land use and the location of leaking or formerly leaking underground fuel tanks in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	18
6. Ternary diagram showing percentage of urban, agricultural, and natural land uses for the study unit and study areas and the area surrounding each USGS-GAMA grid site in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	19
7. Boxplots showing construction characteristics for USGS-GAMA grid sites in the Interior Basins study area and Coastal Basins study area, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	21
8. Boxplots showing relation of groundwater age classification to well depth and depth to top-of-perforations for wells in the Interior Basins study area, well depth and depth to top-of-perforations for wells in the Coastal Basins study area, and bar chart showing the depth class of all wells in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	22
9. Graphs showing groundwater age classification and well depth in relation to pH and dissolved oxygen concentrations, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project	27
10. Graph showing maximum relative-concentration of constituents detected in USGS-GAMA grid sites, by constituent class, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	30

Figures—Continued

11. Graphs showing relative-concentrations of selected constituents with health-based benchmarks and constituents with aesthetic-based benchmarks in USGS-GAMA grid sites, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project31
12. Maps showing relative-concentrations of boron in USGS-GAMA grid sites and geologic features in the Coastal Basins study area and Interior Basins study area, and geologic formations, locations of hot springs and geothermal wells, and approximate boundaries of the Napa and Sonoma Valleys and the Geysers Steam Field, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project34
13. Maps showing relative-concentrations of arsenic in USGS-GAMA grid sites, CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database, and geologic features in the Coastal Basins study area and Interior Basins study area and geologic formations, locations of hot springs and geothermal wells, and approximate boundaries of the Napa and Sonoma Valleys and the Geysers Steam Field Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project36
14. Plots showing boron concentrations relative to field water temperature, field water temperature and distance to nearest hot spring, field water temperature and distance to nearest geothermal well, and redox classification in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project38
15. Plots showing arsenic concentrations relative to groundwater age classification and redox classification in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project41
16. Maps showing relative-concentrations of iron in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database in the Coastal Basins study area and Interior Basins study area, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project43
17. Maps showing relative-concentrations of manganese in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database in the Coastal Basins study area and Interior Basins study area, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project44
18. Graph showing maximum relative-concentration of organic and special-interest constituents detected in USGS-GAMA grid sites, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project45
19. Graphs showing detection frequency and maximum relative-concentration of organic and special-interest constituents detected in USGS-GAMA grid sites, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project46
20. Maps showing relative-concentrations of chloroform detected in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database, and locations of leaking or formerly leaking underground fuel tanks, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project48
21. Plots showing trihalomethane chloroform concentrations relative to groundwater age classification and depth to top-of-perforations in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project49

Figures—Continued

22. Map showing relative-concentrations of simazine detected in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database, and locations of leaking or formerly leaking underground fuel tanks, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	50
23. Plot showing the herbicide simazine concentrations relative to groundwater age classification and depth to top-of-perforations in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	51
24. Maps showing relative-concentrations of perchlorate in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database, and land use, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	52
25. Plot showing the special-interest constituent perchlorate concentrations relative to geochemical condition and depths to top-of-perforations in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	53

Tables

1. Constituent classes and numbers of constituents analyzed for the 58 sites sampled, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	11
2. Numbers of constituents analyzed and detected in each constituent class with each type of benchmark, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	13
3. Constituents selected for additional evaluation in the status assessment, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	14
4. Constituents reported at concentrations greater than benchmarks historically in the California Department of Public Health database, but not during the 3-year time period used in the status assessment, Northern Coast Ranges study unit, California GAMA Priority Basin Project.....	15
5. Results of non-parametric analysis of correlations between selected potential explanatory factors, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	24
6. Results of Wilcoxon rank-sum tests on USGS-GAMA grid site data used to determine significant differences between constituent values grouped by potential explanatory factors, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	25
7. Aquifer-scale proportions calculated using grid-based and spatially weighted methods for those constituents that met criteria for additional evaluation in the status assessment, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	29
8A. Summary of aquifer-scale proportions for inorganic constituent classes, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	33
8B. Summary of aquifer-scale proportions for organic and special-interest constituent classes, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	33
9. Results of non-parametric analysis for correlations between selected water-quality constituents and potential explanatory factors, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.....	40

Conversion Factors, Datums, and Abbreviations and Acronyms

Conversion Factors

Inch/Pound to SI

Multiply	By	To obtain
	Length	
inch (in.)	2.54	centimeter (cm)
foot (ft)	0.3048	meter (m)
mile (mi)	1.609	kilometer (km)
	Area	
square mile (mi ²)	2.590	square kilometer (km ²)

SI to Inch/Pound

Multiply	By	To obtain
	Length	
centimeter (cm)	0.3937	inch (in.)
meter (m)	3.281	foot (ft)
kilometer (km)	0.6214	mile (mi)
	Area	
square kilometer (km ²)	0.3861	square mile (mi ²)

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as follows:

$$^{\circ}\text{F}=(1.8\times^{\circ}\text{C})+32$$

Temperature in degrees Fahrenheit (°F) may be converted to degrees Celsius (°C) as follows:

$$^{\circ}\text{C}=(^{\circ}\text{F}-32)/1.8$$

Concentrations of chemical constituents in water are given either in milligrams per liter (mg/L), or micrograms per liter (µg/L). One milligram per liter is equivalent to 1 part per million (ppm); 1 microgram per liter is equivalent to 1 part per billion (ppb).

Concentrations of dissolved noble gases in water are given in standard pressure per gram of water (g⁻¹ H₂O).

Datums

Vertical coordinate information is referenced to the North American Vertical Datum of 1988 (NAVD 88).

Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Land-surface altitude, as used in this report, refers to distance above the vertical datum and is reported as feet above mean sea level (ft above msl).

Depth, as used in this report, refers to distance below the land-surface datum and is reported as feet below land surface (ft bls).

Selected terms and symbols

=	equal to
>	greater than
≥	greater than or equal to
<	less than
≤	less than or equal to

Abbreviations and Acronyms

AB	Assembly Bill (through the California State Assembly)
AL-US	action level (USEPA)
DO	dissolved oxygen
DOGGR	California Department of Conservation Division of Oil, Gas, and Geothermal Resources
DOM	domestic site
GWSI	Groundwater Site Inventory database (USGS)
HAL	lifetime health advisory level
HAL-US	lifetime health advisory level (USEPA)
HBSL	health-based screening level
IND	industrial site
INS	institutional site
IRR	irrigation site
LRL	laboratory reporting level
LT-MDL	long-term method detection level
LUFT	leaking or formerly leaking underground fuel tank
MCL	maximum contaminant level
MCL-CA	maximum contaminant level (CDPH)
MCL-US	maximum contaminant level (USEPA)
MDL	method detection limit
MRDS	Mineral Resources Data System (USGS)
MRL	minimum reporting level
MTBE	methyl <i>tert</i> -butyl ether
NLCD	National Land Cover Dataset (USGS)
NL-CA	notification level (CDPH)
NOCO	Northern Coast Ranges study unit
NOCO-CO	Coastal Basins study area of the Northern Coast Ranges study unit

Abbreviations and Acronyms—Continued

NOCO-IN	Interior Basins study area of the Northern Coast Ranges study unit
ns	no significant difference
NWIS	National Water Information System (USGS)
p	significance level
pmc	percent modern carbon
RSD5-US	risk-specific dose at a risk factor of 10^{-5} (USEPA)
SMCL	secondary maximum contaminant level
SMCL-CA	secondary maximum contaminant level (CDPH)
SMCL-US	secondary maximum contaminant level (USEPA)
TDS	total dissolved solids
TEAP	terminal electron acceptor process
THM	trihalomethane
TT-US	treatment technique level (USEPA)
TU	tritium unit
VOC	volatile organic compound
Z	test statistic for Wilcoxon test

Organizations

CDPH	California Department of Public Health (California Department of Health Services <i>prior to July 1, 2007</i>)
CDPR	California Department of Pesticide Regulation
CDWR	California Department of Water Resources
GAMA	Groundwater Ambient Monitoring and Assessment Program (USGS)
LLNL	Lawrence Livermore National Laboratory, Livermore, California
NAWQA	National Water-Quality Assessment Program (USGS)
SWRCB	California State Water Resources Control Board
USEPA	U.S. Environmental Protection Agency
USGS	U.S. Geological Survey

Status and Understanding of Groundwater Quality in the Northern Coast Ranges Study Unit, 2009: California GAMA Priority Basin Project

By Timothy M. Mathany and Kenneth Belitz

Abstract

Groundwater quality in the 633-square-mile (1,639-square-kilometer) Northern Coast Ranges (NOCO) study unit was investigated as part of the Priority Basin Project (PBP) of the Groundwater Ambient Monitoring and Assessment (GAMA) Program and the U.S. Geological Survey (USGS) National Water-Quality Assessment Program. The study unit is composed of two study areas (Interior Basins and Coastal Basins) and is located in northern California in Napa, Sonoma, Lake, Colusa, Mendocino, Glenn, Humboldt, and Del Norte Counties. The GAMA-PBP is being conducted by the California State Water Resources Control Board in collaboration with the USGS and the Lawrence Livermore National Laboratory.

The GAMA NOCO study was designed to provide a spatially unbiased assessment of the quality of untreated (ambient) groundwater in the primary aquifer system within the study unit. The assessment is based on water-quality and ancillary data collected in 2009 by the USGS from 58 sites and on water-quality data from the California Department of Public Health (CDPH) database. The primary aquifer system is defined by the perforation intervals of sites listed in the CDPH water-quality database for the NOCO study unit. Groundwater quality in the primary aquifer system may differ from the quality in the shallow or deep water-bearing zones.

The first component of this study, the *status assessment* of the current quality of the groundwater resource, was performed by using data from samples analyzed for inorganic constituents (such as trace elements and major and minor ions), organic constituents (volatile organic compounds and pesticides and pesticide degradates), the special-interest constituent perchlorate, and microbial indicators. This status assessment is intended to characterize the quality of groundwater resources in the primary aquifer system of the NOCO study unit, not the quality of treated drinking water delivered to consumers by water purveyors.

Relative-concentrations (sample concentration divided by the health- or aesthetic-based benchmark concentration) were used for evaluating groundwater quality for those constituents

that have Federal or California regulatory or nonregulatory benchmarks for drinking-water quality. A relative-concentration greater than ($>$) 1.0 indicates a concentration greater than a benchmark, and a relative-concentration less than or equal to (\leq) 1.0 indicates a concentration less than or equal to a benchmark. Relative-concentrations of organic constituents and perchlorate were classified as “high” (relative-concentration >1.0), “moderate” ($0.1 < \text{relative-concentration} \leq 1.0$), or “low” (relative-concentration ≤ 0.1). Relative-concentrations of inorganic constituents were classified as “high” (relative-concentration >1.0), “moderate” ($0.5 < \text{relative-concentration} \leq 1.0$), or “low” (relative-concentration ≤ 0.5).

Aquifer-scale proportion was used as the primary metric in the status assessment for evaluating regional-scale groundwater quality. High aquifer-scale proportion was defined as the percentage of the area of the primary aquifer system with a relative-concentration >1.0 for a particular constituent or class of constituents; the percentage is based on an aerial rather than a volumetric basis. Moderate and low aquifer-scale proportions were defined as the percentage of the primary aquifer system with moderate and low relative-concentrations, respectively. Two statistical approaches—grid-based and spatially weighted—were used to evaluate aquifer-scale proportions for individual constituents and classes of constituents. Grid-based and spatially weighted estimates were comparable in the NOCO study unit (within 90 percent confidence intervals).

Inorganic constituents (one or more) with health-based benchmarks were detected at high relative-concentrations in 10.3 percent and at moderate relative-concentrations in 13.8 percent of the primary aquifer system. The high aquifer-scale proportion of inorganic constituents primarily reflected high aquifer-scale proportions of boron (in 8.6 percent of the primary aquifer system), arsenic (in 3.4 percent), and barium (in 1.7 percent). Inorganic constituents with aesthetic-based benchmarks were detected at high relative-concentrations in 39.7 percent and at moderate relative-concentrations in 10.3 percent of the primary aquifer system. The constituents present at high relative-concentrations were iron (25.9 percent) and manganese (39.7 percent).

2 Status and Understanding of Groundwater Quality in the Northern Coast Ranges Study Unit, 2009

Relative-concentrations of organic constituents with health-based benchmarks (one or more) were high in 0.2 percent, moderate in 1.7 percent, and low in 39.7 percent of the primary aquifer system. Organic constituents were not detected in 58.4 percent of the primary aquifer system. Of the 168 organic constituents analyzed, 11 constituents were detected. Two organic constituents had detection frequencies >10 percent: the trihalomethane chloroform and the herbicide simazine. For the 10 detected organic constituents that had health-based benchmarks, nearly all detections had low relative-concentrations. The special-interest constituent perchlorate was detected at moderate relative-concentrations in 1.7 percent and at low relative-concentrations in 22.4 percent of the primary aquifer system. Perchlorate was not detected in 75.9 percent of the primary aquifer system.

The second component of this study, the *understanding assessment*, evaluated relations between constituent concentrations and values of selected potential explanatory factors to identify the factors potentially affecting the concentrations and occurrences of constituents found at high relative-concentrations or, for organic constituents, with detection frequencies >10 percent. The potential explanatory factors evaluated were land use (including density of septic tanks and leaking or formerly leaking underground fuel tanks), well construction (well depth and depth to the top of the perforated interval in the well), hydrologic conditions (aridity index, field water temperature, and distance to nearest hot spring and geothermal well), pH, dissolved oxygen concentration, study area, groundwater age distribution, and geochemical conditions.

High and moderate relative-concentrations of boron primarily occurred in the Interior Basins study area and may be attributed to groundwater interacting with hydrothermal systems. High and moderate relative-concentrations of boron were associated with elevated groundwater temperatures, groundwater chemistry characteristics similar to those of geothermal waters, and distance to known geothermal areas. Boron concentrations generally were higher where low dissolved oxygen concentrations or anoxic conditions exist. High and moderate relative-concentrations of arsenic predominantly occur in the Interior Basins study area under reducing conditions. Arsenic concentrations also may be influenced by hydrothermal systems (when present).

Chloroform, simazine, and perchlorate were observed in the Interior Basins and Coastal Basins study areas, predominantly at shallow sites with top-of-perforation depths ≤70 feet below land surface, with modern water (post-1950s), and with oxic groundwater conditions.

Introduction

To assess the quality of ambient groundwater in aquifers used for drinking-water supply and to establish a baseline groundwater-quality monitoring program, the California State Water Resources Control Board (SWRCB), in collaboration with the U.S. Geological Survey (USGS) and Lawrence

Livermore National Laboratory (LLNL), implemented the Groundwater Ambient Monitoring and Assessment (GAMA) Program (California State Water Resources Control Board, 2011, website at http://www.waterboards.ca.gov/water_issues/programs/gama/). The statewide GAMA Program was initiated in 2000 in response to Legislative mandates (State of California, 1999, 2001a). The program currently consists of four projects: (1) the GAMA Priority Basin Project (GAMA-PBP), conducted by the USGS (U.S. Geological Survey, 2011, California Water Science Center website at <http://ca.water.usgs.gov/gama/>); (2) the GAMA Domestic Well Project, conducted by the SWRCB; (3) the GAMA Special Studies, conducted by LLNL; and (4) the GeoTracker GAMA web-based groundwater information system, developed by the SWRCB. On a statewide basis, the GAMA-PBP focused on the primary aquifer system, typically the deep portion of the groundwater resource, and the SWRCB Domestic Well Project generally focused on the shallow aquifer systems.

In response to the Groundwater Quality Monitoring Act of 2001, the GAMA-PBP was initiated in 2003 to assess and monitor the quality of groundwater in California (State of California, 2001b). The USGS, in collaboration with the SWRCB, designed the GAMA-PBP as a means to assess groundwater basins through direct sampling of groundwater and other statistically reliable sampling approaches (Belitz and others, 2003; California State Water Resources Control Board, 2003). The GAMA-PBP is a comprehensive assessment of statewide groundwater quality designed to improve understanding and identification of risks to groundwater resources and to increase the availability of information about groundwater quality to the public. Additional partners in the GAMA-PBP include the California Department of Public Health (CDPH), the California Department of Pesticide Regulation (CDPR), the California Department of Water Resources (CDWR), and local water agencies and well owners (Kulongoski and Belitz, 2004).

The ranges of hydrologic, geologic, and climatic conditions in California were considered in this statewide assessment of groundwater quality. Belitz and others (2003) partitioned the State into 10 hydrogeologic provinces, each with distinctive hydrologic, geologic, and climatic characteristics (fig. 1). All of these hydrogeologic provinces include groundwater basins designated by the CDWR (California Department of Water Resources, 2003). The CDWR-defined groundwater basins generally consist of relatively permeable, unconsolidated deposits of alluvial origin. Eighty percent of California's approximately 16,000 active or standby public-supply wells or springs listed in the statewide water-quality database maintained by the CDPH (hereinafter referred to as CDPH sites) are located within CDWR-designated groundwater basins (Belitz and others, 2003). The CDPH Drinking Water Program, which regulates water quality in public-supply wells, was transferred to the SWRCB Division of Drinking Water on July 1, 2014; however, the term 'CDPH sites' is retained in this report for consistency with other GAMA-PBP publications and because the CDPH had jurisdiction over public-supply wells at the time that samples were collected for this study. These basins



Figure 1. Location of the Northern Coast Ranges study unit and the California hydrogeologic provinces (modified from Belitz and others, 2003), California GAMA Priority Basin Project.

4 Status and Understanding of Groundwater Quality in the Northern Coast Ranges Study Unit, 2009

were prioritized for sampling on the basis of the number of CDPH sites in the basin, with secondary consideration given to municipal groundwater use, agricultural pumping, the number of historically leaking underground fuel tanks, and the number of square-mile sections having registered pesticide applications (Belitz and others, 2003). Of the 472 CDWR-designated basins, 116 basins contain approximately 95 percent of CDPH sites located in CDWR-designated groundwater basins and were defined as priority basins (Belitz and others, 2003). The remaining 356 basins were defined as low-use basins. All of the priority basins, selected low-use basins, and selected areas outside of basins were grouped into 35 GAMA-PBP study units that together represent approximately 95 percent of all CDPH sites. The Northern Coast Ranges (NOCO) study unit is composed of 34 groundwater basins and subbasins in the Northern Coast Ranges hydrogeologic province ([fig. 1](#)).

The goal of the GAMA-PBP is to produce three types of water-quality assessments for each study unit: (1) *Status*: assessment of the current quality of the groundwater resource, (2) *Understanding*: identification of the natural and human factors affecting groundwater quality and explanation of the relations between water quality and selected explanatory factors, and (3) *Trends*: detection of changes in groundwater quality over time (Kulongoski and Belitz, 2004). The assessments are intended to characterize the quality of groundwater in the primary aquifer system of the study unit, not the treated drinking water delivered to consumers by water purveyors. The primary aquifer system for a study unit is defined by the depths of the screened or open intervals of the wells listed in the CDPH water-quality database for the study unit. The CDPH water-quality database lists wells and springs used for public drinking-water supplies and includes wells and springs from systems classified as community (such as those in cities, towns, and mobile-home parks), non-transient, non-community (such as those in schools, workplaces, and restaurants), and transient, non-community (such as those in campgrounds and parks). Groundwater quality in the primary aquifer system may differ from that in shallower or deeper parts of the aquifer system. In particular, shallower groundwater may be more vulnerable to contamination from the land surface.

In addition to being selected for study as part of the GAMA-PBP, the Northern Coast Ranges (NOCO) study unit ([fig. 1](#)) was considered a high priority for sampling to complete an assessment of groundwater quality in the California Coastal Basins Principal Aquifer of the United States (U.S. Geological Survey, 2003) by the USGS National Water-Quality Assessment (NAWQA) Program (Lapham and others, 2005). As a result, the NAWQA Program collaborated with the GAMA-PBP to assess groundwater quality in the NOCO study unit.

Purpose and Scope

The purposes of this report are to provide a (1) *study unit description*: description of the hydrogeologic setting of the NOCO study unit, (2) *status assessment*: assessment

of the status of the current (2009) quality of groundwater in the primary aquifer system in the NOCO study unit, and (3) *understanding assessment*: identification of the natural and human factors affecting groundwater quality and explanation of the relations between water quality and selected explanatory factors. Assessments are made for chemical constituents only; microbiological indicators of groundwater quality are not discussed in this report.

The status assessment includes analyses of water-quality data for 58 sites selected by the USGS for spatial coverage of 1 site per grid cell (hereinafter referred to as USGS-GAMA grid sites) across the NOCO study unit and water-quality data from the CDPH database for samples collected between June 1, 2006, and June 1, 2009, at 276 sites. The details of sample collection, analysis, and quality-assurance procedures for the NOCO study unit and all of the water-quality data collected by the USGS are reported by Mathany and others (2011). GAMA-PBP status assessments are designed to provide a statistically robust characterization of groundwater quality in the primary aquifer system at the study-unit scale (Belitz and others, 2003). The statistically robust design also allows basins to be compared and results to be synthesized regionally and statewide. This report describes methods used in designing the sampling network, identifying CDPH data for use in the status assessments, estimating aquifer-scale proportions of relative-concentrations, and assessing the status of groundwater quality by statistical and graphical approaches.

To provide context, the water-quality data discussed in this report are compared to California and Federal regulatory and nonregulatory benchmarks for drinking water by using relative-concentrations (the ratio of the concentration of a constituent in groundwater to the benchmark concentration). The assessments in this report are intended to characterize the quality of untreated groundwater resources in the primary aquifer system within the study unit, not the drinking water delivered to consumers by water purveyors. This study does not attempt to evaluate the quality of water delivered to consumers; after withdrawal from the ground, water typically is treated, disinfected, or blended with water from other sources to maintain acceptable water quality. Regulatory benchmarks apply to drinking water that is delivered to the consumer, not to untreated groundwater.

The understanding assessment is based on water-quality data and data for potential explanatory factors from the 58 sites sampled by the USGS for the GAMA Program. The potential explanatory factors affecting water quality in the primary aquifer system evaluated are land use (including density of septic tanks and leaking or formerly leaking underground fuel tanks), well construction (well depth and depth to the top of the perforated interval in the well), hydrologic conditions (aridity index, field water temperature, and distance to nearest known geothermal reservoir), pH, dissolved oxygen concentration, groundwater age distribution, and geochemical conditions. Connections between potential explanatory factors and water quality were evaluated by using statistical tests for correlations and by analysis of graphical relations.

Hydrogeologic Setting of the Northern Coast Ranges Study Unit

The NOCO study unit covers approximately 633 square miles (mi²) (1,639 square kilometers; km²) in Napa, Sonoma, Lake, Colusa, Mendocino, Glenn, Humboldt, and Del Norte Counties in northern California. The NOCO study unit lies within the Northern Coast Ranges hydrogeologic province (fig. 1; Belitz and others, 2003) and includes 35 groundwater basins and subbasins (California Department of Water Resources, 2003). For the purpose of this study, these 35 groundwater basins and subbasins were grouped into 2 study areas based on location. The groundwater basins and subbasins located inland, not adjacent to the Pacific Ocean, were aggregated into the Interior Basins study area. The groundwater basins and subbasins adjacent to the Pacific Ocean were aggregated into the Coastal Basins study area. As part of the GAMA-PBP, untreated groundwater samples were collected from 58 sites in the NOCO study unit from June 1, 2009, to October 8, 2009 (Mathany and others, 2011).

Northern Coast Ranges—Interior Basins Study Area

The Interior Basins study area (hereinafter referred to as the NOCO-IN study area) is 256 mi² in area and contains 24 CDWR-defined basins (California Department of Water Resources, 2004a–w, ii). The NOCO-IN study area is located in the Coast Ranges Mountains region of northern California stretching approximately 25 to 60 miles (mi) inland of the Pacific Ocean and 60 to 130 mi north of San Francisco. The study area consists primarily of noncontiguous inland valleys whose boundaries are the surrounding hills and (or) mountains. In the central part of the study area, some of the inland valleys share Clear Lake or other lakes as a border (fig. 2B). Land-surface altitudes in the study area range from about 475 feet above mean sea level (ft above msl) near the town of Hopland, to over 4,200 ft above msl on Mount Konocti, located just east of the town of Kelseyville (fig. A1B).

In the NOCO-IN study area, the climate is classified as Mediterranean, with warm to hot, dry summers and cold, wet winters (National Oceanic and Atmospheric Administration, 2011). Most precipitation in the study area occurs in the fall and winter months, with average annual precipitation ranging from about 22 inches (in.) near the town of Stonyford to more than 65 in. near the town of Laytonville (fig. A1A) (California Department of Water Resources, 2004i, r; U.S. Department of Commerce, 2011).

The main surface drainage features of the study area are the Eel, Russian, and Navarro Rivers and their tributaries, all of which have their headwaters outside of the study area and terminate in the Pacific Ocean (figs. A1A–A1C). Large creeks (Outlet, Mill, Stoney, Scotts, St. Helena, and Pope) and numerous small creeks drain the valleys of the study area.

The primary aquifer system in the study area is in Quaternary alluvial groundwater basins made up of sand, silt, gravel, and clay eroded from the surrounding hills. These deposits interfinger with and grade into alluvial fan and terrace deposits along the sides of the valleys, and into older, more consolidated alluvium at depth, and in some valleys, these deposits grade into finer-grained lake deposits towards the center of the basins. Groundwater conditions are mostly unconfined, with some confined areas towards the center of valleys and at depth. The major rock types surrounding the alluvial valleys are the Jurassic-Cretaceous Franciscan Formation (mudstone, greywacke sandstones, and chert) and ophiolitic and metamorphosed volcanic rocks (Muir and Webster, 1977; Farrar, 1986; California Department of Water Resources, 2004e–w). In groundwater basins near Clear Lake, groundwater is also found in Quaternary volcanic flow rocks and thin volcanic ash layers/lenses interbedded with low-permeability sediments (Soil Mechanics and Foundation Engineers, Inc., 1967; Earth Sciences Associates, 1978) (fig. 3B).

The general direction of groundwater flow in the study area is from areas of high topographic relief to the center of the valleys (areas of low topographic relief), following the direction of surface-water features. The study area has several northwest trending faults and fault zones, which act as hydrologic barriers to groundwater movement (California Department of Water Resources, 2004a–w; fig. 3C).

Groundwater recharge in the study area occurs from a mixture of ambient sources, including direct percolation of precipitation and irrigation waters, infiltration of runoff from surrounding hills/areas, and seepage from rivers and creeks (California Department of Water Resources, 2004a–w).

Most of the central and southern parts of the NOCO-IN study area are located within the boundaries of the Geysers–Clear Lake area (fig. 3C), known to be an active geothermal region of northern California (Goff and others, 1993; Hodgson, 2003; California Department of Water Resources, 2004a–d; U.S. Geological Survey, 2004; California Department of Conservation, 2013). The Geysers–Clear Lake area is characterized by recent volcanism (late-Pliocene to early Holocene), high heat flux, and high geothermal gradients that are enhanced locally by fault-controlled zones of convective heat (geothermal fluid) transport (Wood and Kienle, 1992; Stimac and others, 1997; Erkan and others, 2005; Smithsonian Institution, National Museum of Natural History, Global Volcanism Program, 2013). Hydrothermal systems (rather than one large underlying geothermal reservoir) are known to influence the groundwater within the Geysers–Clear Lake area (Goff and others, 1993).

Three geothermal water types have been identified within the Geysers–Clear Lake area: thermal meteoric, steam condensate, and connate/metamorphic (Donnelly-Nolan and others, 1993; Goff and others, 1993; Peters, 1993). Thermal meteoric water is most commonly found near the main mass of the Clear Lake volcanic field and issues from Quaternary volcanic rocks. These waters appear to be a mixture of higher-temperature geothermal water that has resided at depth for

6 Status and Understanding of Groundwater Quality in the Northern Coast Ranges Study Unit, 2009

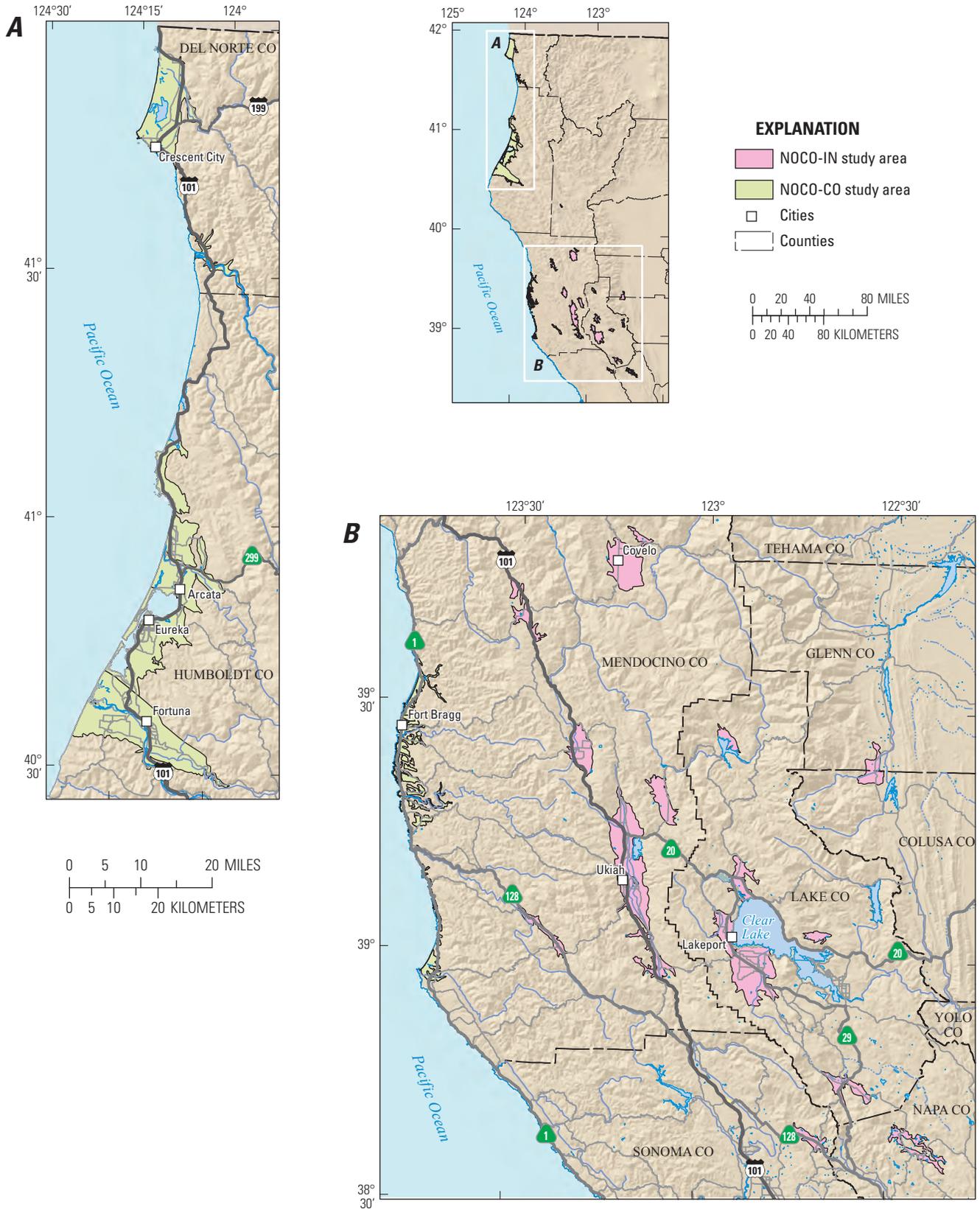


Figure 2. Location of the (A) Coastal Basins (NOCO-CO) and (B) Interior Basins (NOCO-IN) study areas, major cities and towns, major roads, and hydrologic features in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

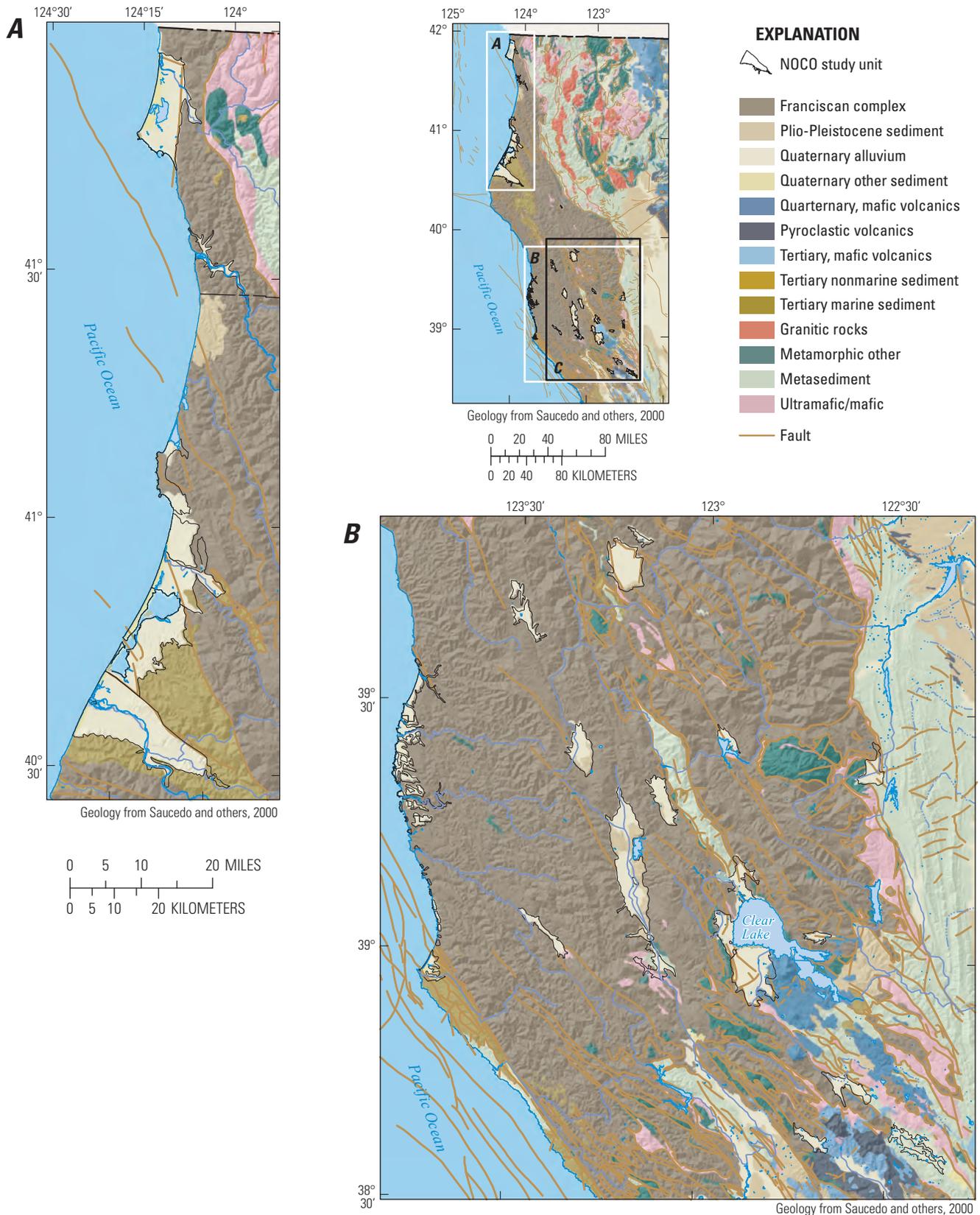


Figure 3. Geologic formations and locations of the (A) Coastal Basins study area, (B) and (C) Interior Basins study area, locations of hot springs and geothermal wells, and approximate boundaries of the Napa and Sonoma Valleys and the Geysers Steam Field, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

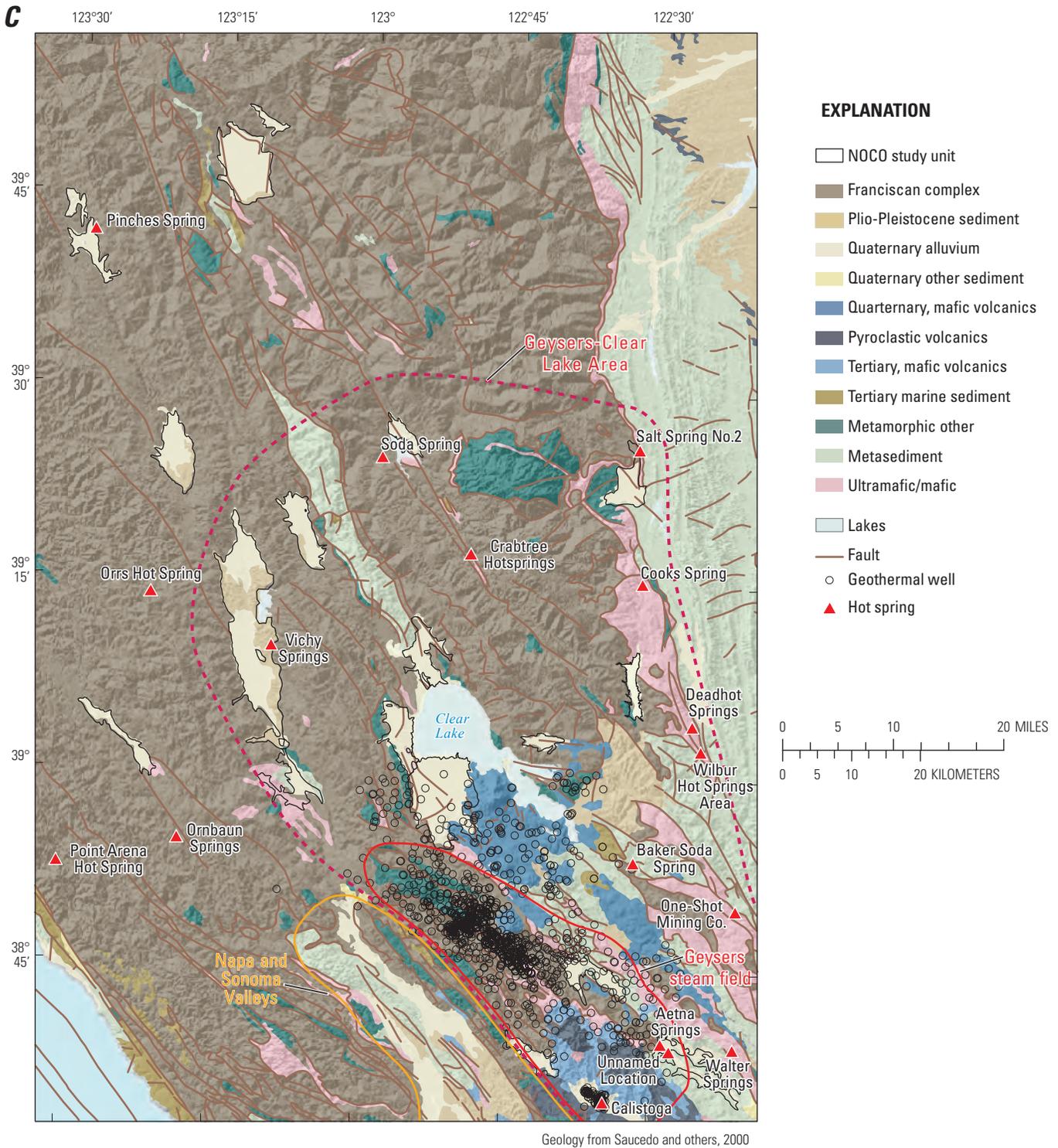


Figure 3. —Continued

relatively short time periods and a cooler, near-surface meteoric water. Steam condensate waters are most commonly found near the Geysers steam field along the southwestern portion of the Geysers–Clear Lake area (fig. 3C). These waters are a mixture of groundwater, condensed steam from an underlying hydrothermal system, and heated groundwater. Connate and metamorphic (a subclass of connate) waters are commonly found in the eastern reaches of the Geysers–Clear Lake area. Connate waters typically issue from rocks of the Great Valley sequence and are most common in the eastern Clear Lake region, whereas metamorphic waters typically issue from rocks of the Franciscan Formation near the main Clear Lake volcanic field. These three geothermal water types are unique in distribution and origin, but they share common chemical signatures of hydrothermal activity—near-neutral pH and elevated concentrations of boron, bromide, arsenic, and lithium (Donnelly-Nolan and others, 1993; Goff and others, 1993; Peters, 1993).

The NOCO-IN study area is just north of the Napa and Sonoma Valleys (fig. 3C), a hydrothermally active area with similar Tertiary to Holocene age volcanic processes as in the Geysers–Clear Lake area (U.S. Geological Survey, 2004). Similar to the Geysers–Clear Lake area, the Napa and Sonoma Valleys are characterized by localized hydrothermal systems associated with fault zones, high geothermal gradients, and high heat flux (Kulongoski and others, 2010). A study by Forrest and others (2013) used USGS-GAMA data from the Napa and Sonoma Valleys (North San Francisco Bay study unit: Kulongoski and others, 2006) to determine which constituents were most indicative of hydrothermal contamination of public supply wells. The study identified water temperature, concentrations of boron, arsenic, lithium, and chloride, and helium isotope ratios as the constituents that were most useful in evaluating the contribution of the hydrothermal system to the public supply wells.

Northern Coast Ranges—Coastal Basins Study Area

The Coastal Basins study area (hereinafter referred to as the NOCO-CO study area) is 377 mi² in area and contains 11 CDWR-defined basins and subbasins (California Department of Water Resources, 2004x–hh). The study area is located along the northern California coast beginning north of Point Arena and extending to the Oregon border (figs. A1D–A1F). The study area has three distinct sections (northern, central, and southern), each bordered on the west by the Pacific Ocean and on the north, east, and south by the hills surrounding the coastal valleys (fig. 2A). Land-surface altitudes in the study area range from sea level where the study area boundary touches the Pacific Ocean, to approximately 500 ft above msl in the hills above the city of Fortuna (fig. 2).

In the NOCO-CO study area, the climate is classified as coastal (National Oceanic and Atmospheric Administration, 2011), with cool to mild summers and cold, wet winters. Coastal fog and low clouds are common throughout the year.

Average annual precipitation in the study area ranges from about 38 in. near the town of Trinidad to about 79 in. near the town of Klamath (figs. A1D, A1E; U.S. Department of Commerce, 2011).

The major surface drainage features of the study area are the Smith, Klamath, Mad, Eel, Noyo, and Big Rivers and their tributaries, all of which have their headwaters outside of the study area and terminate in the Pacific Ocean (figs. A1D–A1F).

The primary aquifer system in the northern part of the study area principally occurs in the Smith River Plain, a marine terrace made up of Holocene alluvial fan and floodplain deposits; in the Pleistocene terrace deposits; and in the Pleistocene Battery Formation (California Department of Water Resources, 1987, 2004x). The Smith River Plain and the Pleistocene terrace deposits are included in Quaternary alluvium and Pleistocene-Pliocene nonmarine deposits, and the Pleistocene Battery Formation is mapped as Tertiary marine sediment (fig. 3A). In the central part of the study area, the primary aquifer system occurs in groundwater basins consisting of alluvial deposits associated with drowned river mouths of the Klamath, Mad, and Eel Rivers. The alluvial deposits consist of Holocene alluvium, dune sand, the Pleistocene Hookton Formation, and the deeper Pliocene-Pleistocene Carlotta Formation (Johnson, 1978; California Department of Water Resources, 2004y–dd). The primary aquifer system in the southern part of the study area is in the Fort Bragg Terrace deposits. These deposits are discontinuous, uplifted, and dissected Pleistocene marine terraces along the Pacific Ocean. The terrace deposits consist of Tertiary marine sediment, dune sand, and semi-consolidated clay, silt, sand, and gravel derived from adjacent formations (largely the Franciscan complex) that are up to 150 feet (ft) thick (California Department of Water Resources, Northern District, 1982; California Department of Water Resources, 2004ff–gg).

The general direction of groundwater-flow in the study area is from east to west towards the Pacific Ocean. In the northern part of the study area, dune sand accumulation has created Lake Earl, which acts as a restrictive structure to groundwater movement (fig. A1D). In the central and southern parts of the study area, several northwest trending faults and fault zones act as hydrologic barriers (California Department of Water Resources, 2004x–gg; fig. 3A).

Groundwater recharge in the study area comes from a mixture of ambient sources including direct percolation of precipitation and irrigation waters, infiltration of runoff from surrounding hills/areas, seepage from rivers and creeks, and subsurface inflow (from non-alluvial geologic units that underlie the alluvial basins) (California Department of Water Resources, 2004x–gg).

The groundwater basins in the northern and central parts of the study area are in hydraulic connection with the Pacific Ocean, and seawater extends landward as a wedge underneath fresh groundwater in some of the basins. In the southern part of the study area, the uplifted terrace deposits prevent hydraulic connection between groundwater basins and the Pacific Ocean (California Department of Water Resources, 2004x–gg).

Methods

This section describes the methods used for the status assessment and understanding assessment for water quality in the NOCO study unit. Methods used for compiling data for the potential explanatory factors are described in [appendix B](#).

Status Assessment

The status assessment provides a spatially unbiased assessment of groundwater quality in the primary aquifer system of the NOCO study unit. Methods used for the status assessment included (1) assembling water-quality benchmarks and calculating relative-concentrations, (2) assembling datasets for use in the status assessment, (3) selecting constituents for additional evaluation, and (4) calculating aquifer-scale proportions for these constituents.

Relative-Concentrations and Water-Quality Benchmarks

To provide context for the groundwater-quality data, concentrations of constituents measured in the untreated groundwater were compared to benchmarks established by the U.S. Environmental Protection Agency (USEPA) and CDPH for evaluation of the quality of drinking water (U.S. Environmental Protection Agency, 2008, 2009a, b, 2012; California Department of Public Health, 2006, 2008, 2010). The benchmarks used for each constituent were selected in the following order of priority:

1. Regulatory, health-based CDPH and USEPA maximum contaminant levels (MCL-CA and MCL-US) and action levels (AL-US).
2. Nonregulatory, aesthetic-based CDPH and USEPA secondary maximum contaminant levels (SMCL-CA and SMCL-US). For constituents with recommended and upper SMCL-CA levels, the values for the upper levels were used.
3. Nonregulatory, health-based CDPH notification levels (NL-CA), USEPA lifetime health advisory levels (HAL-US), and USEPA risk-specific doses for a risk of 1 in 100,000 (RSD5-US). Risk-specific doses for 1 in 100,000 ($1 \text{ in } 10^5$) were calculated by dividing USEPA values for estimated excess lifetime cancer risk from drinking water of 1 in 10^4 by 10.

For constituents with multiple types of benchmarks, this hierarchy may not result in selection of the benchmark with the lowest concentration. Additional information on the types of benchmarks and listings of the benchmarks for all constituents analyzed is provided by Mathany and others (2011).

Groundwater-quality data are presented as relative-concentrations, the concentrations of constituents measured

in groundwater relative to regulatory and nonregulatory benchmarks used to evaluate drinking-water quality:

$$\text{Relative-concentration} = \frac{\text{Sample concentration}}{\text{Benchmark concentration}} \quad (1)$$

Relative-concentrations less than 1.0 indicate a sample concentration less than the benchmark, and relative-concentrations greater than 1.0 indicate a sample concentration greater than the benchmark. The use of relative-concentrations also permits comparison on a single scale of constituents present at a wide range of concentrations. Relative-concentrations can only be computed for constituents with water-quality benchmarks; therefore, constituents without water-quality benchmarks are not included in the status assessment.

The four microbial indicators analyzed in samples from the NOCO study unit, total coliforms, *Escherichia coli* (*E. coli*), and coliphage (somatic and f-specific), have water-quality benchmarks, but are not included in the status and understanding assessments because the results will be presented in 1 report for all 25 GAMA-PBP public-supply aquifer study units (Carmen Burton, U.S. Geological Survey, written commun., 2014).

Toccalino and others (2004), Toccalino and Norman (2006), and Rowe and others (2007) previously used the ratio of measured sample concentration to the benchmark concentration [either MCL-US or health-based screening levels (HBSLs)] and defined this ratio as the benchmark quotient. HBSLs were not used in this report because HBSLs are not currently used as benchmarks by California drinking-water regulatory agencies. Because different water-quality benchmarks may be used to calculate relative-concentrations and benchmark quotients, the terms are not interchangeable.

For ease of discussion, relative-concentrations of constituents were classified into low, moderate, and high categories:

[**Abbreviations:** >, less than; ≤, greater than or equal to]

Category	Relative-concentrations for organic and special-interest constituents	Relative-concentrations for inorganic constituents
High	>1	>1
Moderate	>0.1 and ≤1	>0.5 and ≤1
Low	≤0.1	≤0.5

For organic and special-interest constituents, a relative-concentration of 0.1 was used as a threshold to distinguish between low and moderate relative-concentrations for consistency with other studies and reporting requirements (U.S. Environmental Protection Agency, 1998; Toccalino and others, 2004). For inorganic constituents, which tend to be more prevalent than organic constituents in groundwater (Toccalino and others, 2010), a relative-concentration of 0.5 was used as a threshold to distinguish between low and moderate relative-concentrations. The primary reason for using a higher threshold

for inorganic constituents was to focus attention on those inorganic constituents that are of most immediate concern (Fram and Belitz, 2012). Although more complex classifications could be devised based on the properties and sources of individual constituents, use of a single moderate/low threshold value for each of the two major groups of constituents provided a consistent and objective criteria for distinguishing constituents at moderate rather than low concentrations.

Datasets Used for Status Assessment

Two datasets were used in the status assessments: (1) data from USGS-GAMA grid sites and (2) data combined from USGS-GAMA grid sites and CDPH sites. This section explains how each dataset was assembled. Comparisons of USGS-GAMA and CDPH data are presented in [appendix D](#).

USGS-GAMA Grid Sites

The grid-based calculations of aquifer-scale proportions used data from 58 sites sampled by the USGS for spatial coverage of one well per grid cell across the study unit (grid sites). Detailed descriptions of the methods used to identify sites for sampling are given in Mathany and others (2011). Briefly, each study area was divided into equal-area grid cells (Interior Basins, 8.5-mi² [22.0-km²] cells and Coastal Basins, 12.5-mi² [32.4-km²] cells), and in each cell, one site was randomly selected to represent the cell (Scott, 1990) ([figs. 4, A1A–A1F](#)). Sites were selected from those in the statewide database of public-supply wells maintained by the CDPH. If a cell had no accessible sites listed in the CDPH water-quality database, then appropriate sites were selected from the USGS Groundwater Site Inventory (GWSI) database. The NOCO study unit contained 60 grid cells, and the USGS sampled sites in 58 of those cells. Of the 58 USGS-GAMA grid sites, 31 were listed in the CDPH water-quality database, and 27 were non-CDPH sites perforated at depths similar to the depths of CDPH sites in their respective cells. One of the USGS-GAMA grid sites that was listed in the CDPH database was a spring. USGS-GAMA grid sites were named with an alphanumeric GAMA-ID consisting of a prefix identifying the study area and a number indicating the order of sample collection ([figs. A1A–A1F](#)). The following prefixes were used to identify the study area: NOCO-IN (Interior Basins study area) and NOCO-CO (Coastal Basins study area).

Samples collected from USGS-GAMA grid sites were analyzed for 234 constituents ([table 1](#)). The collection, analysis, and quality-control data for the constituents listed in [table 1](#) are described by Mathany and others (2011). Because samples from all USGS-GAMA grid sites were analyzed for the full suite of constituents, it was not necessary to supplement the grid site dataset with data from the CDPH database, as was done in the status assessments in many other GAMA-PBP study units (Landon and others, 2010; Burton and others, 2011; Kulongoski and Belitz, 2011). The two grid cells without USGS-GAMA grid sites had no CDPH sites with water-quality data for the 3-year interval June 1, 2006, through June 1, 2009.

Table 1. Constituent classes and numbers of constituents analyzed for the 58 sites sampled, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[GAMA, Groundwater Ambient Monitoring and Assessment Program]

Constituent class	Number of constituents ¹
Inorganic constituents	
Trace elements	23
Major and minor ions, alkalinity, specific conductance, and total dissolved solids	13
Nutrients and dissolved organic carbon	6
Uranium and other radioactive constituents ²	7
Organic constituents	
Volatile organic compounds	85
Pesticides and pesticide degradates	83
Special-interest constituent	
Perchlorate	1
Geochemical and age-dating tracers	
Dissolved oxygen, pH, temperature, and turbidity	4
Tritium	1
Noble gases (helium, neon, argon, krypton, xenon), ³ He/ ⁴ He of helium, and tritium	7
δ ² H and δ ¹⁸ O of water	2
Carbon-14 and δ ¹³ C of dissolved carbonates	2
Sum	234

¹All sites also were analyzed for 12 pharmaceutical compounds and 4 microbial indicators. Assessment results for these constituents are not being presented for individual study units, therefore, they are not included in the count of constituents. A statewide synthesis of GAMA Priority Basin Project results for pharmaceuticals was presented by Fram and Belitz (2011), and a synthesis of results for microbial indicators will be presented at a later date (Carmen Burton, U.S. Geological Survey California Water Science Center, written commun., 2013).

²Activities of radon-222, radium, gross alpha and gross beta particles, lead-210, and polonium-210 and uranium concentrations were measured. Gross alpha and gross beta particle activities were measured after 72-hour and 30-day holding times; data from the 30-day measurement are used in this report. Radium activity equals the sum of the two isotopes measured: radium-226 and radium-228.

Additional Data Used for Spatially Weighted Calculation

The spatially weighted calculations of aquifer-scale proportions of relative-concentrations used data from the 58 USGS-GAMA grid sites and from the 276 sites in the CDPH water-quality database having water-quality data during the 3-year interval June 1, 2006, through June 1, 2009. For the 27 sites with USGS-GAMA and CDPH data, only the USGS-GAMA data were used. Many of the 249 sites having only CDPH data had data for a limited number of constituents. For example, 162 sites had data for nitrate, but only 72 sites had data for at least 1 trace element. Water-quality data collected by the CDPH are available from the SWRCB's publically-accessible Internet database GeoTracker GAMA (California State Water Resources Control Board, 2009, website at <https://geotracker.waterboards.ca.gov/gama/>).

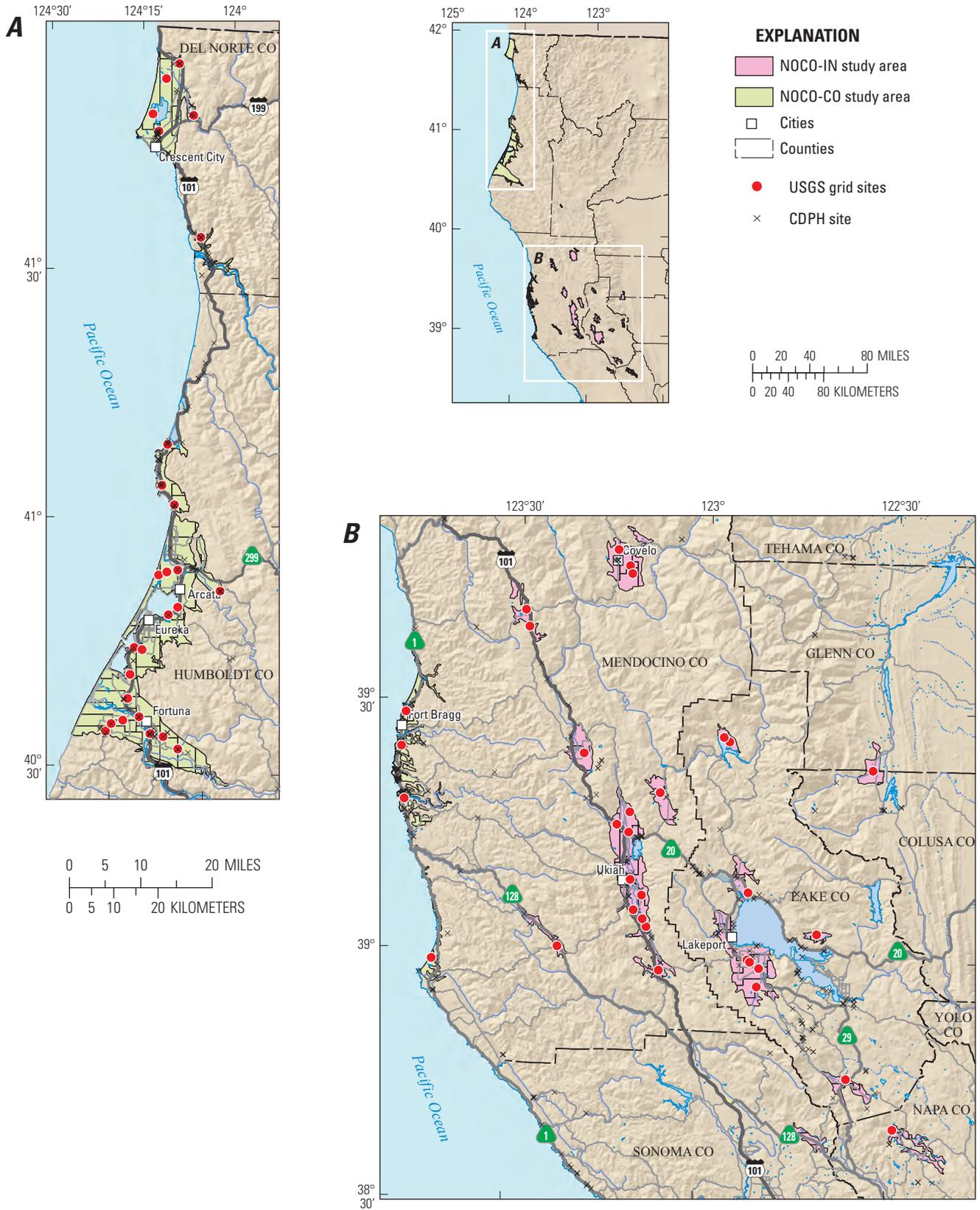


Figure 4. Locations of study area grid cells, USGS-GAMA grid sites, and CDPH sites, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area]

Selection of Constituents for Additional Evaluation in the Status Assessment

As many as 234 constituents were analyzed by USGS-GAMA in samples from sites in the NOCO study unit as part of the status assessment (table 1); however, only subsets of

these constituents were identified for additional evaluation in this report. Of the 238 constituents analyzed, 114 constituents did not have benchmarks (table 2). Because relative-concentrations cannot be calculated for constituents without benchmarks, these 114 constituents were not evaluated in this report. The 126 constituents having benchmarks were assessed,

Table 2. Numbers of constituents analyzed and detected in each constituent class with each type of benchmark, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[GAMA, Groundwater Ambient Monitoring and Assessment Program]

Benchmark type	Groups of inorganic constituents						Sum of inorganic constituents	
	Trace elements and major and minor ions		Nutrients		Uranium and other radioactive constituents			
	Number of constituents							
	Analyzed	Detected	Analyzed	Detected	Analyzed	Detected	Analyzed	Detected
Regulatory, health-based	14	14	2	2	5	5	21	21
Nonregulatory, health-based	4	4	1	1	0	0	5	5
Nonregulatory, aesthetic-based	8	8	0	0	0	0	8	8
None	10	10	3	3	2	2	15	15
Total:	36	36	6	6	7	7	49	49

Benchmark type	Groups of organic constituents				Sum of organic constituents	
	Volatile organic compounds		Pesticides and pesticide degradates			
	Number of constituents					
	Analyzed	Detected	Analyzed	Detected	Analyzed	Detected
Regulatory, health-based	33	7	6	2	39	9
Nonregulatory, health-based	26	1	18	0	44	1
Nonregulatory, aesthetic-based	0	0	0	0	0	0
None	26	0	59	1	85	1
Total:	85	8	83	3	168	11

Benchmark type	Other constituent groups			
	Geochemical and age-dating tracers		Special-interest constituent	
	Number of constituents			
	Analyzed	Detected	Analyzed	Detected
Regulatory, health-based	1	1	1	1
Nonregulatory, health-based	0	0	0	0
Nonregulatory, aesthetic-based	1	1	0	0
None	14	14	0	0
Total:	16	16	1	1

Benchmark type	Sum of all constituents	
	Number of constituents	
	Analyzed	Detected
Regulatory, health-based	62	32
Nonregulatory, health-based	49	6
Nonregulatory, aesthetic-based	9	9
None	114	30
Total:	234	77

and a subset of these constituents was selected for additional evaluation (table 3) on the basis of the following two criteria:

- Constituents present at high or moderate relative-concentrations at the USGS-GAMA grid sites or in the CDPH water-quality database for any sample collected between June 1, 2006, and June 1, 2009;
- Organic constituents with detection frequencies of greater than 10 percent in the USGS-GAMA grid site dataset for the study unit.

The USGS conducted a review of the CDPH water-quality database (January 1, 1976, to May 31, 2006) to identify constituents with high relative-concentrations

historically, but not currently. Constituent concentrations may be historically high, but not currently high, because of improvement of groundwater quality with time or abandonment of sites with high concentrations. Historically high concentrations of constituents that did not otherwise meet the criteria for additional evaluation in the status assessment are not considered representative of potential groundwater-quality concerns in the study unit from 2006 to 2009.

For the NOCO study unit, nine constituents had high relative-concentrations reported in the CDPH database during the historical period, but did not also have high concentrations reported during the current period or in the USGS-GAMA dataset (table 4).

Table 3. Constituents selected for additional evaluation in the status assessment, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[**Benchmark type:** *Regulatory, health-based benchmarks:* MCL-US, USEPA maximum contaminant level; MCL-CA, CDPH maximum contaminant level. *Nonregulatory, health-based benchmarks:* NL-CA, CDPH notification level. *Nonregulatory, aesthetic-based benchmarks:* SMCL-CA, CDPH secondary maximum contaminant level. **Benchmark units:** µg/L, micrograms per liter; mg/L, milligrams per liter. **Abbreviations:** USGS, U.S. Geological Survey; GAMA, Groundwater Ambient Monitoring and Assessment Program; USEPA, U.S. Environmental Protection Agency; CDPH, California Department of Public Health; TDS, total dissolved solids; MTBE, methyl *tert*-butyl ether]

Constituent class	Typical use or source	Benchmark			Criteria for selection for status assessment	Selected for understanding assessment
		Type ¹	Value	Units		
Inorganic constituents with health-based benchmarks						
Trace elements						
Arsenic	Naturally occurring	MCL-US	10	µg/L	Concentration	Yes
Barium	Naturally occurring	MCL-CA	1,000	µg/L	Concentration	No
Boron	Naturally occurring	NL-CA	1,000	µg/L	Concentration	Yes
Nickel	Naturally occurring	MCL-CA	100	µg/L	Concentration	No
Nutrients						
Nitrate (as nitrogen)	Fertilizer, human and animal waste, naturally occurring	MCL-US	10	mg/L	Concentration	No
Inorganic constituents with aesthetic-based benchmarks						
Iron	Naturally occurring	SMCL-CA	300	µg/L	Concentration	Yes
Manganese	Naturally occurring	SMCL-CA	50	µg/L	Concentration	Yes
TDS	Naturally occurring	SMCL-CA	1,000	mg/L	Concentration	No
Organic constituents with health-based benchmarks						
Volatile organic compounds						
Chloroform (Trichloromethane)	Disinfection byproduct, trihalomethane	MCL-US	80	µg/L	Concentration, frequency	Yes
Vinyl chloride (Chloroethene)	Organic synthesis reagent	MCL-CA	0.5	µg/L	Concentration	No
MTBE	Gasoline oxygenate	MCL-CA	13	µg/L	Concentration ³	No
Pesticides and pesticide degradates						
Simazine	Herbicide	MCL-US	4	µg/L	Frequency	Yes
Special-interest constituent with health-based benchmark						
Perchlorate	Naturally occurring, rocket fuel, fireworks, fertilizer	MCL-CA	6	µg/L	Concentration, frequency	Yes

¹Maximum contaminant level benchmarks are listed as MCL-US when the MCL-US and MCL-CA are identical and as MCL-CA when the MCL-CA is lower than the MCL-US or no MCL-US exists. Sources of benchmarks:

MCL-CA, NL-CA, and SMCL-CA: California Department of Public Health (2006, 2008, 2010)

MCL-US: U.S. Environmental Protection Agency (2008, 2009a, 2009b, 2012).

²Benchmark value is the sum of four trihalomethanes (chloroform, bromodichloromethane, dibromochloromethane, and bromoform).

³High values were found in the CDPH database during the current period of study. MTBE was detected only at low relative-concentrations by USGS-GAMA.

Table 4. Constituents reported at concentrations greater than benchmarks historically (January 1, 1976 to May 31, 2006) in the California Department of Public Health database, but not during the 3-year time period used in the status assessment, Northern Coast Ranges study unit, California GAMA Priority Basin Project.

[**Benchmark type:** *Regulatory, health-based benchmarks:* MCL-US, USEPA maximum contaminant level; AL-US, USEPA action level; RSD5-US, USEPA risk-specific dose at a risk factor of 10^{-5} ; MCL-CA, CDPH maximum contaminant level. *Nonregulatory, aesthetic-based benchmark:* SMCL-CA, CDPH secondary maximum contaminant level. **Benchmark units:** $\mu\text{g/L}$, micrograms per liter; mg/L , milligrams per liter. **Abbreviations:** USGS, U.S. Geological Survey; GAMA, Groundwater Ambient Monitoring and Assessment Program; USEPA, U.S. Environmental Protection Agency; CDPH, California Department of Public Health; TTHMs, total trihalomethanes]

Constituent class	Typical use or source	Benchmark			Number of sites with historical data	Number of sites with a high value	Date of most recent high value
		Type ¹	Value	Units			
Inorganic constituents							
Aluminum	Naturally occurring	MCL-CA	1,000	$\mu\text{g/L}$	348	2	07/22/2002
Copper	Naturally occurring	AL-US	1,300	$\mu\text{g/L}$	358	1	10/22/2002
Chloride	Naturally occurring	SMCL-CA	500	mg/L	363	1	02/18/2004
Fluoride	Naturally occurring	MCL-CA	2	$\mu\text{g/L}$	462	1	11/25/2003
Lead	Naturally occurring	AL-US	15	$\mu\text{g/L}$	330	10	10/08/2003
Organic constituents							
Aldrin ²	Insecticide	RSD5-US	0.02	$\mu\text{g/L}$	104	1	04/04/1978
Di(2-ethylhexyl)phthalate ²	Plasticizer	MCL-CA	4	$\mu\text{g/L}$	49	2	09/04/1992
Lindane ²	Insecticide	MCL-US	0.2	$\mu\text{g/L}$	185	1	04/03/1990
TTHMs	Disinfection byproduct	MCL-US ³	80	$\mu\text{g/L}$	373	1	06/01/1988

¹Maximum contaminant level benchmarks are listed as MCL-US when the MCL-US and MCL-CA are identical and as MCL-CA when the MCL-CA is lower than the MCL-US or no MCL-US exists. Sources of benchmarks:

MCL-CA and SMCL-CA: California Department of Public Health (2006, 2008, 2010)

MCL-US, AL-US, and RSD5-US: U.S. Environmental Protection Agency (2008, 2009a, 2009b, 2012).

²Constituent not analyzed by USGS-GAMA in the Northern Coast Ranges study unit.

³MCL-US benchmark for TTHMs is for the sum of chloroform, bromodichloromethane, dibromochloromethane, and bromoform.

Calculation of Aquifer-Scale Proportions

The status assessment is intended to characterize the quality of groundwater resources in the primary aquifer system of the NOCO study unit. The primary aquifer system is defined by the depth intervals over which wells listed in the CDPH database are screened or perforated; these wells are primarily classified as municipal and community drinking-water supply wells. The use of the term “primary aquifer system” does not imply that a discrete aquifer unit exists. In most groundwater basins, municipal and community supply wells generally are perforated at greater depths than domestic wells. However, to the extent that domestic wells are perforated over the same depth intervals as the CDPH wells, the assessments presented in this report also may be applicable to the portions of the aquifer system used for domestic drinking-water supplies.

Two statistical approaches, grid-based and spatially weighted (Belitz and others, 2010), were selected to evaluate the proportions of the primary aquifer system in the NOCO study unit with high, moderate, and low relative-concentrations of constituents relative to benchmarks. Aquifer-scale proportions are defined as the percentage of the area (rather than the volume) of the primary aquifer systems with high, moderate, and low relative-concentrations. For ease

of discussion, these proportions are referred to as “high,” “moderate,” and “low” aquifer-scale proportions. Calculations of aquifer-scale proportions were made for individual constituents, as well as for classes of constituents. The classes consisted of groups of related individual constituents. For constituents with human-health benchmarks, the classes included trace elements, nutrients, radioactive constituents, volatile organic compounds, and pesticides.

Both statistical approaches use equal-area grids. The grid-cell sizes in the two study areas of the NOCO study unit were not the same; therefore, aquifer-scale proportions were calculated separately for the two study areas and then combined on an area-weighted basis to obtain aquifer-scale proportions for the study unit as a whole (Bennett and others, 2010; Fram and Belitz, 2012):

$$P_{SU} = \sum P_{SA} F_{SA} \quad (2)$$

where:

- P_{SU} is the aquifer-scale proportion for the study unit,
- P_{SA} is the aquifer-scale proportion for a study area, and
- F_{SA} is the fraction of the total study unit area occupied by the study area.

Aquifer-scale proportions for individual constituents and constituent classes in the NOCO-CO and NOCO-IN study areas are presented in [appendix C \(tables C1, C2\)](#); results for the study unit as a whole are presented in the main body of the report.

The grid-based calculation uses the USGS-GAMA grid-well dataset. For each constituent, the high aquifer-scale proportion was calculated by dividing the number of cells represented by a high relative-concentration for that constituent by the total number of grid cells with data for that constituent. The moderate and low aquifer-scale proportions were calculated similarly. Confidence intervals for the high aquifer-scale proportions for individual constituents were computed using the Jeffrey's interval for the binomial distribution (Brown and others, 2001; Belitz and others, 2010). For calculation of high aquifer-scale proportion for a class of constituents, cells were considered high if relative-concentrations for any of the constituents in that class were high. Cells were considered moderate if values for any of the constituents were moderate, but no values were high. The grid-based estimate is spatially unbiased. However, the grid-based approach may not detect constituents that are present at high concentrations in small proportions of the primary aquifer systems.

The spatially weighted calculation uses the dataset assembled from the USGS-GAMA grid sites and data compiled in the CDPH database for samples collected between June 1, 2006, and June 1, 2009 (the most recent analysis was used for each constituent at each well). For each constituent, the high aquifer-scale proportion was calculated by computing the proportion of sites with high relative-concentrations in each cell and then averaging the proportions for all cells (Isaaks and Srivastava, 1989; Belitz and others, 2010). The moderate aquifer-scale proportion was calculated similarly. Confidence intervals for spatially weighted detection frequencies of high concentrations are not described in this report. For calculation of high aquifer-scale proportion for a class of constituents, the aquifer-scale proportions for sites were considered high if the relative-concentrations for any of the constituents in that class were high. Aquifer-scale proportions for sites were considered moderate if the relative-concentrations for any of the constituents were moderate, but none were high.

In addition, for each constituent, the raw detection frequencies of high and moderate values for individual constituents were calculated using the same dataset as used for the spatially weighted calculations. Raw detection frequencies are not spatially unbiased, however, because the sites in the CDPH database are not uniformly distributed throughout the NOCO study unit ([fig. 4](#)). For example, if a constituent were present at high concentrations in a small region of the aquifer with a high density of sites, the raw detection frequency of high values would be greater than the high aquifer-scale proportion. Raw detection frequencies are provided for reference but were not used to assess aquifer-scale proportions.

Aquifer-scale proportions discussed in this report were estimated primarily by using the grid-based approach and secondarily by using the spatially weighted approach. The grid-based aquifer-scale proportions were used unless the spatially weighted proportions were significantly different. Significantly different results were defined as follows:

- If the grid-based high aquifer-scale proportion was zero and the spatially weighted proportion was non-zero, then the spatially weighted result was used. This situation can happen when the relative-concentration of a constituent is high in a small fraction of the primary aquifer system.
- If the grid-based high aquifer-scale proportion was non-zero and the spatially weighted proportion was outside the 90 percent confidence interval, then the spatially weighted proportion was used.

The grid-based moderate and low proportions were used in most cases because the reporting levels for many organic constituents and some inorganic constituents in the CDPH database were higher than the threshold between moderate and low categories. However, if the grid-based moderate proportion was zero and the spatially weighted proportion non-zero, then the spatially weighted value was used as a minimum estimate for the moderate proportion.

Understanding-Assessment Methods

The understanding assessment is designed to evaluate natural and anthropogenic factors that may affect groundwater quality. A finite set of potential explanatory factors was considered: study area, land use, well depth, depth to the top of well casing perforation (top of perforation), density of septic tanks, density of leaking (or formerly leaking) underground fuel tanks (LUFTs), aridity index, groundwater age, oxidation-reduction condition (redox), field water temperature, distance to nearest hot spring listed in the USGS Mineral Resources Data System (MRDS), distance to nearest geothermal well listed in the California Department of Conservation Division of Oil, Gas, and Geothermal Resources (DOGGR) database, dissolved oxygen concentration, and pH. Statistical tests were used to identify significant correlations between the constituents of interest and these potential explanatory factors.

A subset of the constituents examined in the status assessment and selected classes of constituents were examined in the understanding assessment:

- Constituents with high aquifer-scale proportions of greater than 2 percent. These constituents were selected to focus the understanding assessment on those constituents that may have the greatest effect on groundwater quality.
- Classes of organic constituents and special-interest constituents detected in 10 percent or more of the USGS-GAMA grid sites, regardless of concentration.

The understanding assessment was based on the 58 grid sites sampled by USGS-GAMA. CDPH sites were not used because data for many of the potential explanatory factors were not available. In particular, data for age-dating tracers, dissolved oxygen, well depth, and depth to the top of screened interval are not maintained in the CDPH database.

Statistical Analysis

Nonparametric statistical methods were used to test the significance of correlations among water-quality variables and potential explanatory variables. Nonparametric statistics are robust techniques that are generally not affected by outliers and do not require that the data follow any particular distribution (Helsel and Hirsch, 2002). The significance level (p) used for hypothesis testing for this report was compared to a threshold value (α) of 5 percent ($\alpha = 0.05$) to evaluate whether the relation was statistically significant ($p < \alpha$).

Two different statistical tests were used because the set of potential explanatory factors included categorical and continuous variables. Correlations among continuous variables (land use, septic tank density, LUFT density, well depth, depth to top-of-perforations, dissolved oxygen concentration, pH, field water temperature, aridity index, and distance to the nearest hot spring listed in the USGS MRDS and geothermal well listed in the DOGGR database) were investigated by using Spearman's method to calculate the rank-order correlation coefficients (ρ) among continuous variables. The values of ρ can range from +1.0 (perfect positive correlation), through 0.0 (no correlation), to -1.0 (perfect negative correlation). For potential explanatory factors that were classified into categories (study area, groundwater age, geochemical conditions, and well depth class), the values of water-quality parameters between the categories were compared by using the Wilcoxon rank-sum test. The Wilcoxon rank-sum test is a median test statistic that compares two independent data groups (categories) to determine whether one group contains larger values than the other (Helsel and Hirsch, 2002). The null hypothesis for the Wilcoxon rank-sum test is that there is no significant difference between the observations of the two independent data groups being tested. All statistical analyses were done using TIBCO Spotfire S+® 8.1 for Windows (TIBCO Software Inc., 2008).

Potential Explanatory Factors

Brief descriptions of potential explanatory factors (land use, well construction, hydrologic conditions, groundwater age, and geochemical conditions) are given in this section. Correlations among these potential explanatory factors that could affect apparent relations between potential explanatory factors and water quality also are described. The data sources and methods used for assigning values for potential explanatory factors are described in [appendix B](#).

Land Use

Land use was quantified as the percentages of three land-use types: agricultural, urban, and natural. Land-use percentages were calculated for the study unit, the study areas, and within a radius of 500 meters (m) (500-m buffers) around sites (Johnson and Belitz, 2009).

In the NOCO study unit, natural lands are mostly grasslands, wetlands, beaches, and forests, and the primary use of agricultural land is for pasture (cattle, sheep, and poultry), row crops, hay, vineyards, and timberlands (Nakagaki and others, 2007). The largest urban areas are the cities of Crescent City, Arcata, Eureka, Fort Bragg, Willits, Ukiah, and Lakeport ([fig. 5](#)).

Land use based on all of the land within the NOCO study unit was 60 percent natural, 29 percent agricultural, and 11 percent urban ([fig. 6A](#)). Compared to the land use in the entire study unit, the average land use in 500-m buffer areas around the CDPH sites was less natural (56 percent), less agricultural (19 percent), and more urban (25 percent). The average land use in the 500-m buffer areas around the USGS-GAMA grid sites was 46 percent natural, 37 percent agricultural, and 17 percent urban. The difference between overall land use in the NOCO study unit and land use around the CDPH sites and USGS-GAMA grid sites reflects the fact that CDPH sites and USGS-GAMA grid sites are preferentially located where people are living and working. The difference between the average land use around the CDPH sites and around the USGS-GAMA grid sites reflects the spatially distributed nature of the USGS-GAMA grid sites. The CDPH sites are more biased towards urban land use because more urban areas typically have a higher density of CDPH sites.

The 58 USGS-GAMA grid sites showed a wide variation in land use ([fig. 6B](#)). In the NOCO-IN study area, approximately two-thirds of the sites were primarily mixtures of agricultural and natural land use, and most of the other sites were primarily mixtures of urban and agricultural land use. In the NOCO-CO study area, approximately one-half of the sites were primarily mixtures of urban and natural land use, and most of the other sites were dominated by agricultural land use. The median amount of percent urban land use was significantly greater in the areas surrounding the NOCO-CO sites than for the NOCO-IN sites ($p = 0.022$; [table B2](#)).

Septic tanks and LUFTs within the 500-m buffers around the USGS-GAMA grid sites also were used as markers of land-use patterns. Generally, areas with greater percentages of urban land use have more septic tanks and LUFTs on average, and areas with natural land use have fewer septic tanks and LUFTs. The density of septic tanks the NOCO-IN study area ranged from 0.1 to 30.8 tanks per square kilometer (tanks/km²), with a median of 4.8 tanks/km² ([table B2](#)). In the NOCO-CO study area, the density of septic tanks ranged from 0.4 to 44.9 tanks/km², with a median of 6.6 tanks/km². There were no significant differences in the density of septic tanks ($p = 0.54$) between the NOCO-IN and NOCO-CO study

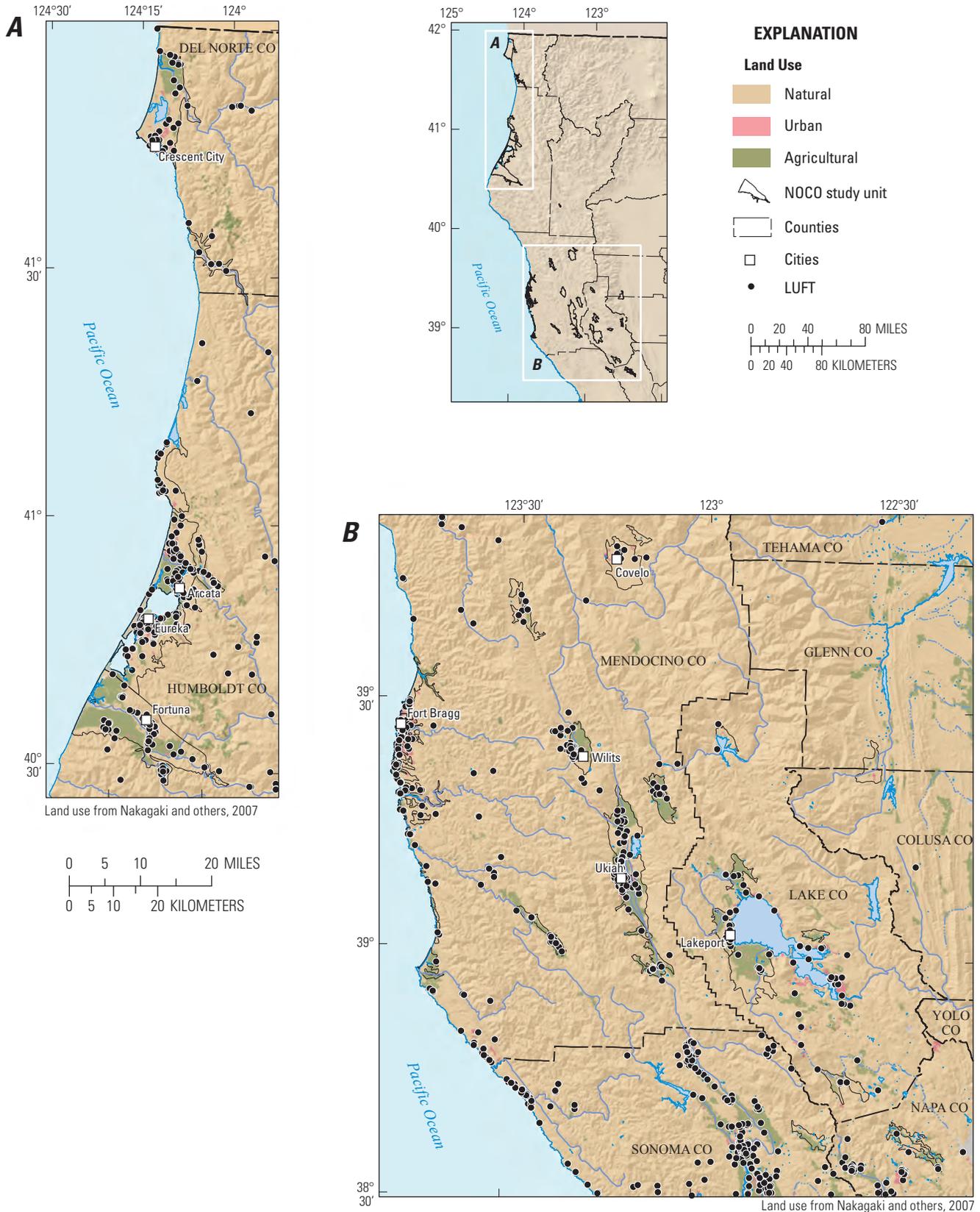


Figure 5. Land use and the location of leaking or formerly leaking underground fuel tanks (LUFTs) in the Northern Coast Ranges (NOCO) study unit, 2009, California GAMA Priority Basin Project.

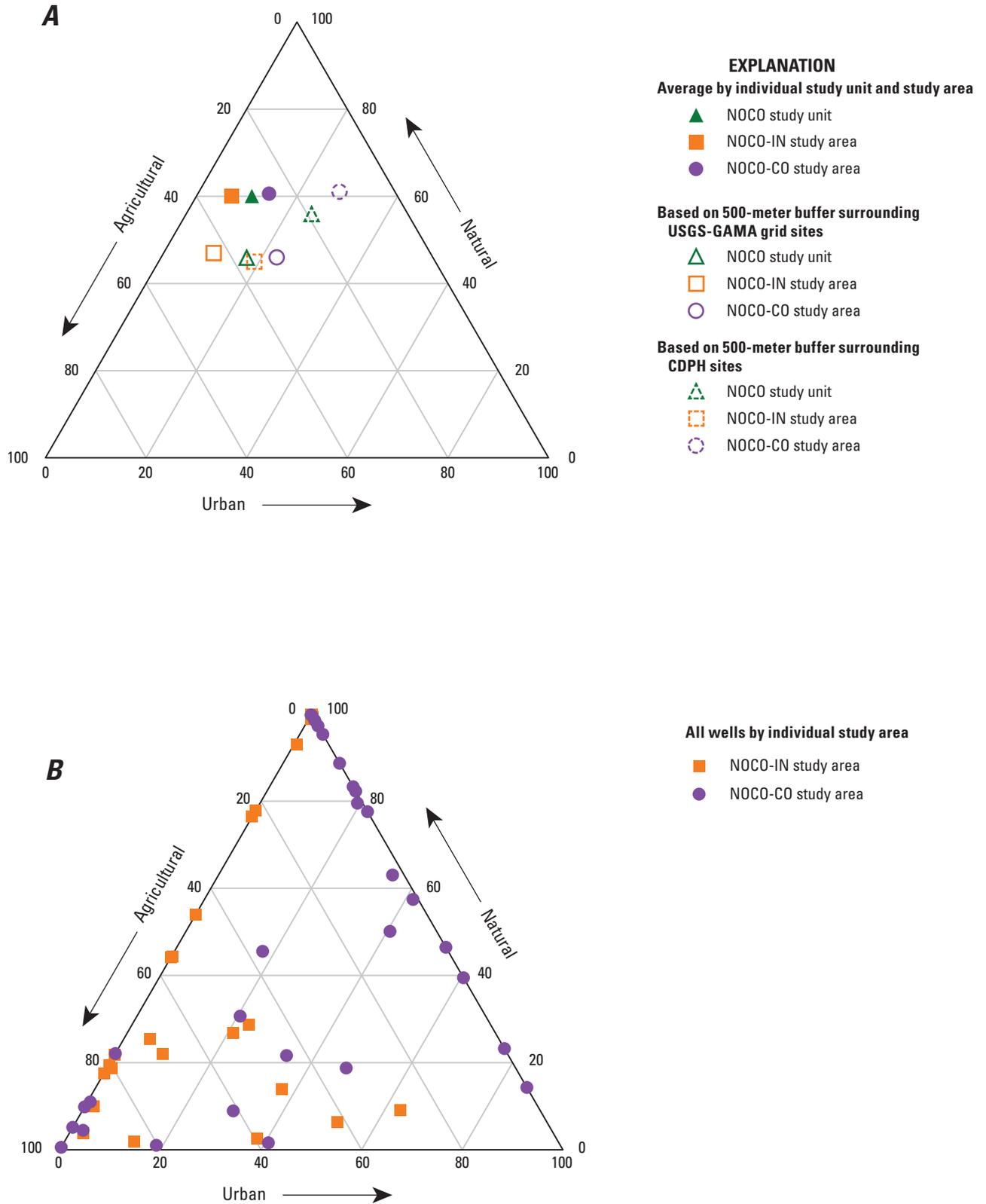


Figure 6. Percentage of urban, agricultural, and natural land uses for the (A) study unit and study areas and (B) the area surrounding each USGS-GAMA grid site in the Northern Coast Ranges (NOCO) study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area]

areas. The density of LUFTs was significantly greater in the areas surrounding the NOCO-CO sites than for the NOCO-IN sites ($p = 0.003$). In the NOCO-IN study area, the density of LUFTs ranged from 0 to 1.77 tanks/km², with a median of 0.02 tanks/km². In the more urban NOCO-CO study area, the density of LUFTs ranged from 0.01 to 1.49 tanks/km², with a median of 0.160 tanks/km².

Well Construction Information

In the NOCO study unit, depths of the USGS-GAMA grid wells (not including the spring) ranged from 15 to 400 ft below land surface (bls), with a median of 100 ft bls (figs. 7A, B). Depths to the top-of-perforations ranged from 10 to 356 ft bls, with a median of 40 ft bls. The perforation length was as much as 237 ft (median: 35 ft; table B3).

Well depths in the NOCO-IN study area ranged from 36 to 400 ft bls, with a median of 116 ft bls. Depths to top-of-perforations ranged from 15 to 148 ft bls, with a median of 40 ft bls and perforation lengths as much as 237 ft (median: 60 ft). In the NOCO-CO study area, wells were shallower and had shorter perforation lengths than the wells in the NOCO-IN study area. Well depths ranged from 15 to 400 ft bls, with a median of 85 ft bls. Depths to top-of-perforations ranged from 10 to 356 ft bls, with a median of 35 ft bls and perforation lengths up to 120 ft (median: 20 ft; table B3).

Hydrologic Conditions

Hydrologic conditions are represented by aridity index, field water temperature, and distance to nearest hot spring listed in the USGS MRDS and geothermal well listed in the DOGGR database (appendix B). The aridity index was used as an indicator of climate and is the average annual precipitation (PRISM Group, Oregon State University, 2012) divided by average annual evapotranspiration (Flint and Flint, 2007). Higher values indicate wetter conditions: values <0.05 are defined as hyper-arid, 0.05–0.20 as arid, 0.20–0.50 as semi-arid, 0.50–0.65 as dry sub-humid, 0.65–1.00 as humid, and >1.00 as wet. In the NOCO-IN study area, the aridity index ranged from 0.48 to 1.73, with a median of 0.83 (table B3). The aridity index in the NOCO-CO study area ranged from 0.86 to 2.17, with a median of 1.26.

Field water temperature was used as a potential indicator of groundwater interacting with hydrothermal systems (Forrest and others, 2013). In the NOCO-IN study area, field water temperatures ranged from 12.5 to 25.0 degrees Celsius (°C), with a median of 18.0 °C (table B3). Field water temperatures in the NOCO-CO study area ranged from 10.5 to 18.0 °C, with a median of 13.0 °C.

Distance to the nearest hot spring listed in the USGS MRDS and geothermal well listed in the DOGGR database was also used as a potential indicator of groundwater interacting with hydrothermal systems (fig. 3C). Previous investigations in the Coast Range Mountains of northern California by the USGS and others have shown that water

discharged from thermal/mineral springs are commonly enriched with dissolved trace elements such as boron, arsenic, bromide, lithium, silver, and gold, among other trace elements and major and minor ions (Berkstresser, 1968; Thompson and others, 1978; Goff and others, 1993; Peters, 1993).

While being near a hot spring or geothermal well alone does not prove direct mixing with thermal spring waters, it may indicate a greater probability that groundwater at a site may be influenced to some degree by the underlying hydrothermal system supplying the hot spring or geothermal well. In the NOCO-IN study area, distance to the nearest hot spring listed in the USGS MRDS ranged from 0.61 to 30.9 kilometers (km), with a median of 13.5 km. Distance to the nearest geothermal well listed in the DOGGR database ranged from 0.37 to 93.7 km, with a median of 26.5 km. In the NOCO-CO study area, the distance to the nearest hot spring ranged from 20.2 to 142 km, with a median of 103 km, and the distance to the nearest geothermal well ranged from 51.1 to 341 km, with a median of 223 km (table B3; U.S. Geological Survey, 2005; California Department of Conservation, 2013).

Hydrologic condition data indicate that the climate was drier, the field water temperatures were warmer, and the hot springs and geothermal wells were closer to the USGS-GAMA sites in the NOCO-IN study area (inland) compared to sites in the NOCO-CO study area (coastal) (tables 6, B3).

Groundwater Age

Data for the age-dating tracers tritium and carbon-14 were used to classify groundwater age distributions into three categories: modern, mixed, and pre-modern (appendix B). Groundwater with tritium activity >0.5 tritium unit (TU) was defined as “modern,” and groundwater with tritium activity <0.5 TU was defined as “pre-modern.” Modern groundwater contains a substantial component of water recharged since 1952 (Michel, 1989). The presence of pre-modern groundwater also was identified using the carbon-14 data: samples with percentage of modern carbon <88 percent were considered to contain a substantial component of pre-modern groundwater. Samples with tritium activity >0.5 TU and modern carbon percentage <88 percent were classified as “mixed-age.” Samples with tritium activity >0.5 TU but no carbon-14 data were classified as “modern or mixed-age.” Of the 58 samples collected by USGS-GAMA, 24 were classified as modern, 20 were mixed-age (evidence of modern and pre-modern groundwater in the same sample), 13 were pre-modern, and 1 was modern or mixed-age (table B4).

Classified groundwater ages generally increased with well depth and depth to the top of the well perforations. Groundwater with pre-modern-age distributions came from wells with significantly greater depths to top of the perforation than groundwater with modern ($p = <0.001$) or mixed-age distributions ($p = 0.008$) (figs. 8B, C). Groundwater with modern-age distributions came from wells with significantly shallower depths than groundwater with mixed-age ($p = 0.006$) or pre-modern-age distributions ($p = 0$) (figs. 8A, C).

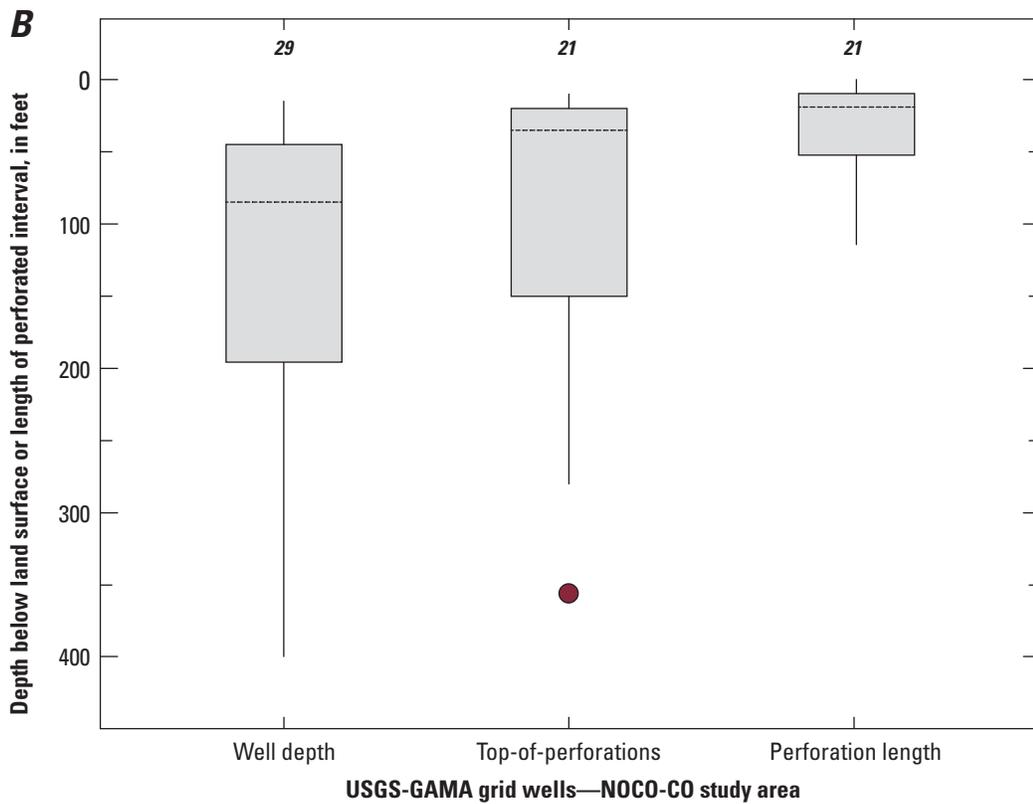
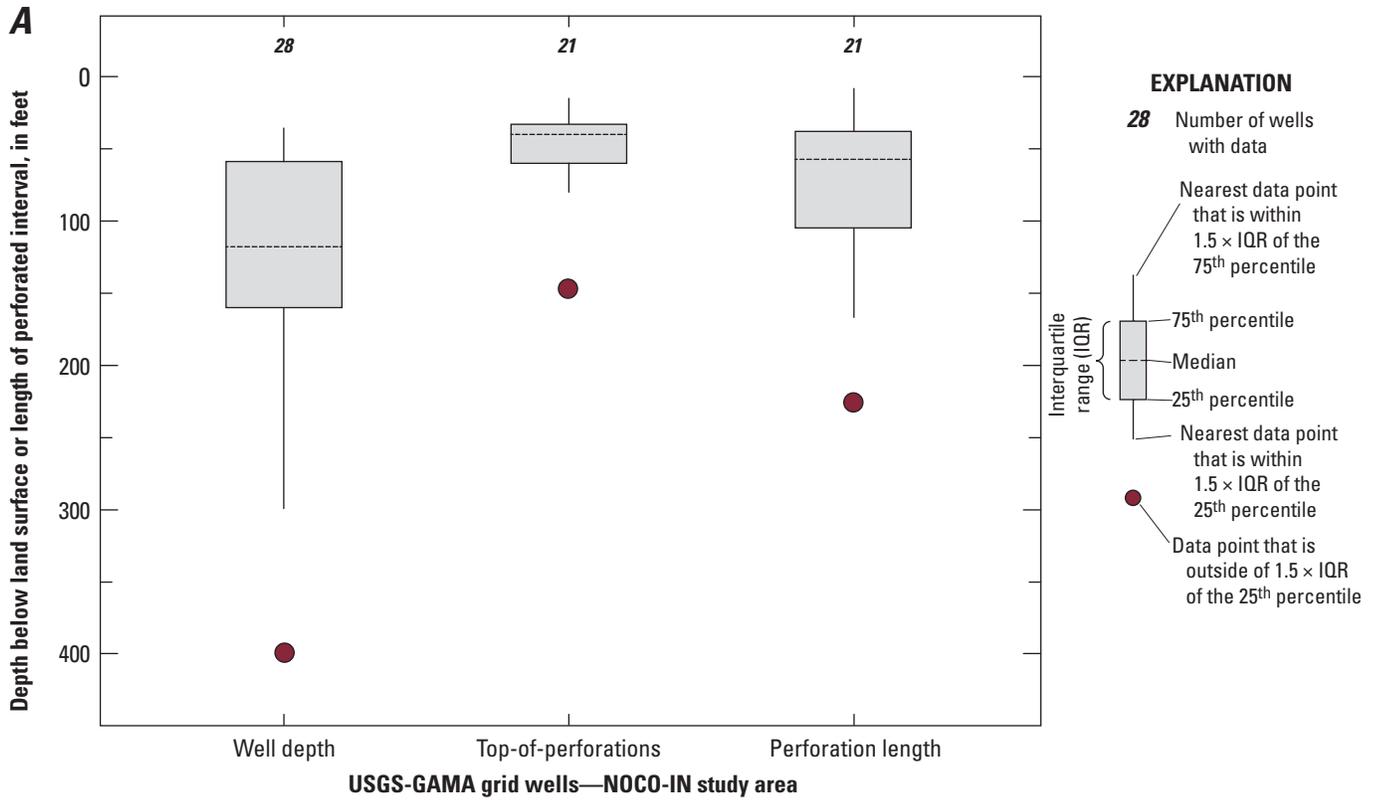


Figure 7. Construction characteristics for USGS-GAMA grid sites in the (A) Interior Basins study area and (B) Coastal Basins study area, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

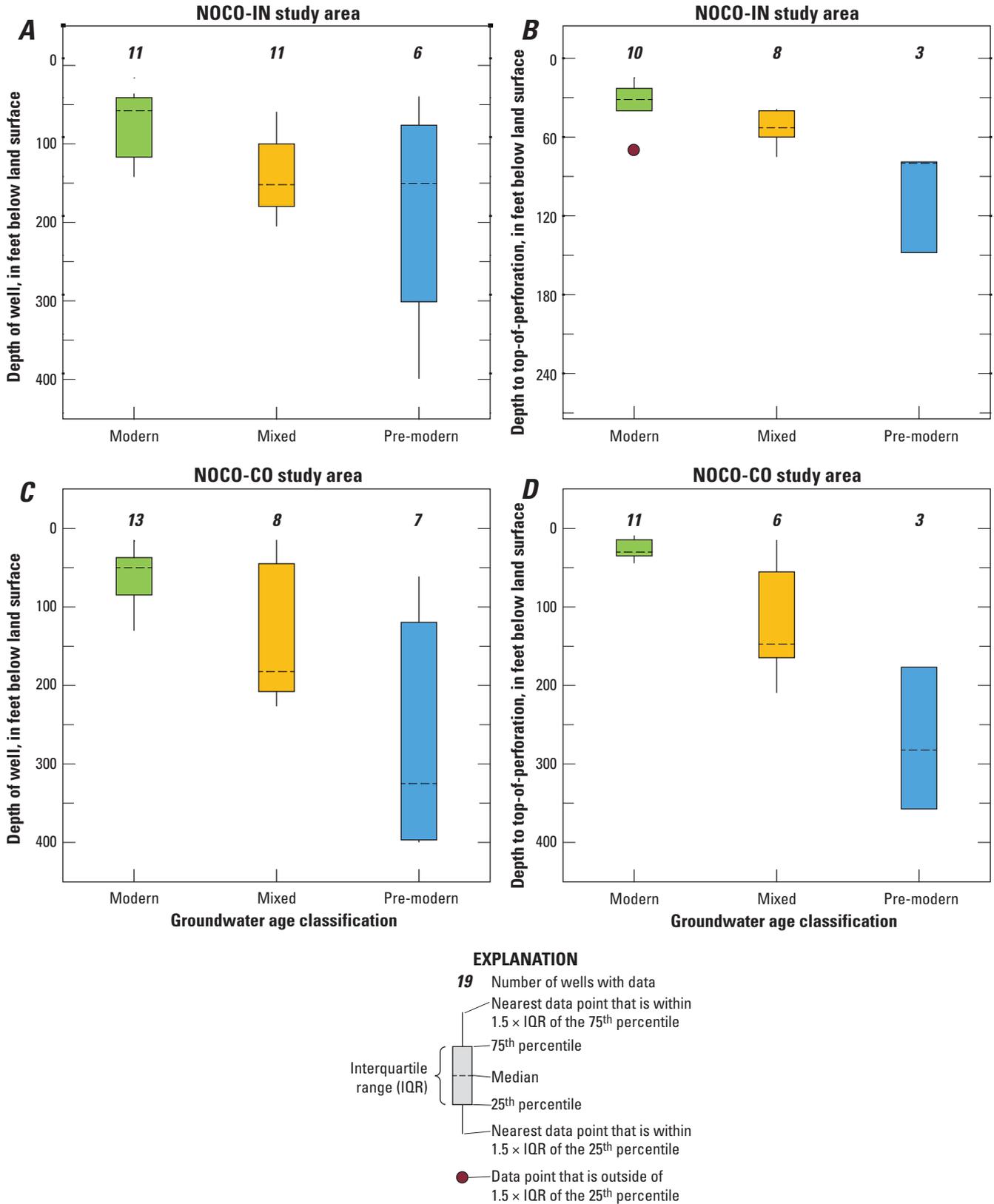


Figure 8. Relation of groundwater age classification to (A) well depth and (B) depth to top-of-perforations for wells in the Interior Basins study area, (C) well depth and (D) depth to top-of-perforations for wells in the Coastal Basins study area, and (E) bar chart showing the depth class of all wells in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

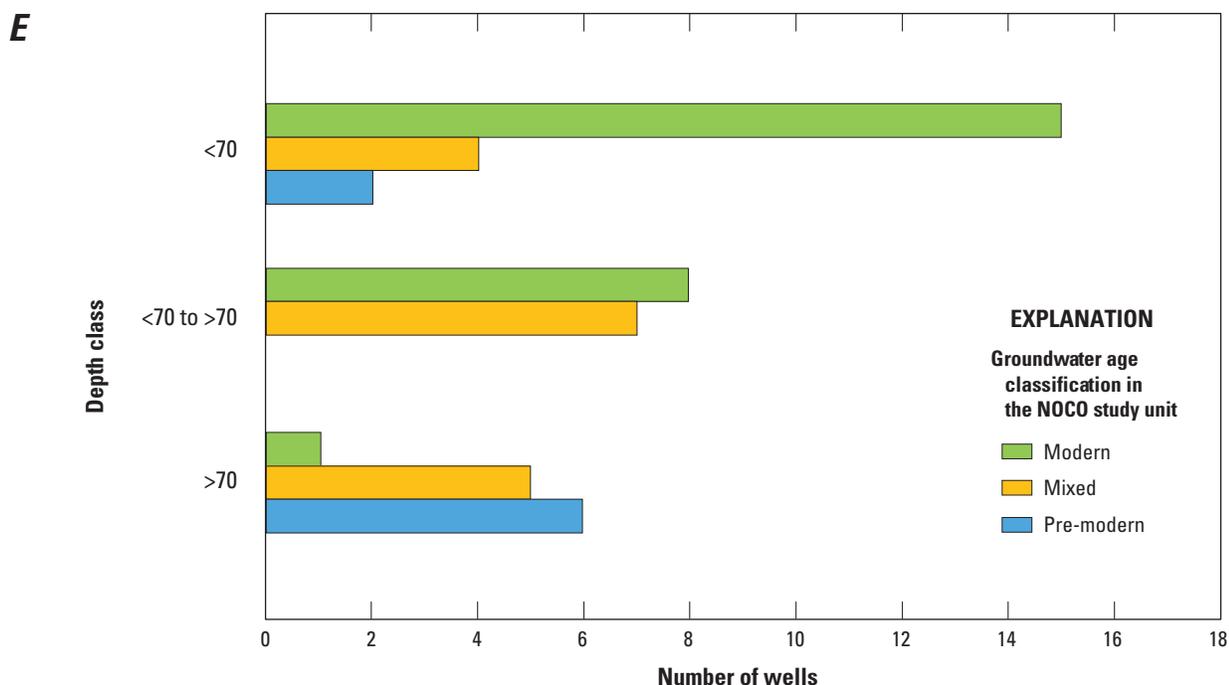


Figure 8. —Continued

Classified groundwater ages were used to create a classification system for well depth. Wells of depth less than a critical depth were defined as shallow; wells with perforations beginning above the critical depth and ending below the critical depth were defined as mixed-depth; wells with perforations entirely below the critical depth were defined as deep. Wells with total well depth below the critical depth and unknown perforation information were defined as mixed-depth or deep. The critical depth of 70 ft bls (depth to top of the perforation interval) was selected by optimizing the segregation of modern-age groundwater into shallow wells, mixed-age groundwater into mixed-depth wells, and pre-modern-age groundwater into deep wells.

Classification system for well depth

[**Abbreviations:** ft bls, feet below land surface; ≥, greater than or equal to; <, less than]

Well depth classification	Well depth	Depth to top of the perforation
Shallow	<70 ft bls	<70 ft bls
Mixed-depth	≥70 ft bls	<70 ft bls
Deep	≥70 ft bls	≥70 ft bls
Mixed-depth or deep	≥70 ft bls	Unknown

In the NOCO study unit, 57 percent of the shallow wells had modern-age groundwater, 41 percent of the mixed-depth wells had mixed-age groundwater, and 55 percent of the deep wells had pre-modern-age groundwater ([fig. 8E](#); [table B4](#)).

Eight wells were classified as mixed-depth or deep; three had mixed-age groundwater, and five had pre-modern-age groundwater. The spring had mixed-age groundwater and is not represented in [figure 8](#).

Geochemical Conditions

Sufficient chemical data, including dissolved oxygen (DO) concentration, were available and used to classify groundwater oxidation-reduction (redox) conditions into three categories: oxic, mixed, and anoxic ([appendix B](#)). Groundwater was oxic at 31 of the sites, mixed at 1 of the sites, and anoxic at 26 of the sites. Anoxic conditions were further subdivided into suboxic; manganese-reducing; iron-reducing; manganese- and iron-reducing; nitrate- and manganese-reducing; and nitrate-, manganese-, and iron-reducing conditions ([table B5](#)).

Correlations Between Explanatory Factors

Apparent correlations between potential explanatory factors and water-quality constituents could result from correlations among potential explanatory factors; therefore, identification of statistically significant correlations between potential explanatory factors is important ([tables 5, 6](#)). Implications of correlations between explanatory variables are discussed later in the report as part of analysis of factors affecting individual water-quality constituents.

Table 5. Results of non-parametric (Spearman’s *rho* method) analysis of correlations between selected potential explanatory factors, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[**Explanation:** Spearman’s *rho* test used and *rho* values shown. Significance level ($p \leq 0.05$); Correlations with $rho > 0.2$ and (or) p -values ≤ 0.05 are shown in **bold**. **Abbreviations:** GAMA, Groundwater Ambient Monitoring and Assessment Program; USGS, U.S. Geological Survey; \leq , less than or equal to; $>$, greater than; ns, no significant correlation; **black text**, significant positive correlation; **red text**, significant negative correlation; DOGGR, Division of Oil, Gas, and Geothermal Resources (California Department of Conservation)]

<i>rho</i> : Spearman’s correlation statistic	Percent urban land use	Percent agricultural land use	Percent natural land use	Density of septic tanks	Density of leaking underground fuel tanks	Well depth	Depth to top of perforations	Dissolved oxygen concentration	pH	Field water temperature	Aridity index ¹	Distance to nearest hot spring listed in the USGS Mineral Resources Data System	Distance to nearest geothermal well listed in the DOGGR database
Percent urban land use		ns	-0.36	0.50	0.51	ns	ns	ns	ns	ns	0.30	0.26	0.26
Percent agricultural land use			-0.80	ns	ns	ns	ns	ns	ns	0.39	-0.29	ns	-0.29
Percent natural land use				-0.36	-0.41	ns	ns	ns	ns	ns	ns	ns	ns
Density of septic tanks					0.37	ns	ns	ns	ns	ns	ns	ns	ns
Density of leaking underground fuel tanks						ns	ns	ns	ns	ns	ns	0.32	ns
Well depth							0.82	ns	0.56	0.34	ns	ns	ns
Depth to top of perforations								ns	0.68	0.39	ns	ns	ns
Dissolved oxygen concentration									ns	ns	ns	ns	ns
pH										ns	ns	ns	ns
Field water temperature											ns	-0.66	-0.72
Aridity index ¹												0.64	0.90
Distance to nearest hot spring listed in the USGS Mineral Resources Data System													0.79
Distance to nearest geothermal well listed in the DOGGR database													

¹Aridity index is average annual precipitation (PRISM Group, Oregon State University, 2012) divided by average annual evapotranspiration (Flint and Flint, 2007).

Table 6. Results of Wilcoxon rank-sum tests on USGS-GAMA grid site data used to determine significant differences between constituent values grouped by potential explanatory factors, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[**Explanation:** Potential explanatory factors without significant correlations are not shown on this table. Only results with p-values ≤ 0.05 are considered significant in this study; Wilcoxon rank-sum tests with exact distribution and continuity correction; Z, test statistic for Wilcoxon test; a significantly positive Z value in a column to the right indicates that the entry in the leftmost column has larger values in the first classification in the column to the right compared to the second classification; a significantly negative Z value in a column to the right indicates that the entry in the leftmost column has larger values in the second classification in the column to the right compared to the first classification. **USGS-GAMA site identification numbers:** NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Abbreviations:** USGS, U.S. Geological Survey; GAMA, Groundwater Ambient Monitoring and Assessment Program; <, less than; >, greater than; ≤, less than or equal to; ns, no significant difference (p-value > 0.05); DOGGR, Division of Oil, Gas, and Geothermal Resources (California Department of Conservation); %, percent]

Selected potential explanatory factors and water-quality constituents	Potential explanatory factors				
	Study area	Groundwater age classifications			Redox classifications
	NOCO-IN compared with NOCO-CO	Modern age compared with mixed-age	Mixed-age compared with pre-modern age	Modern age compared with pre-modern age	Oxic conditions compared with anoxic conditions
	Z: Test statistic for Wilcoxon test	Z: Test statistic for Wilcoxon test			Z: Test statistic for Wilcoxon test
Potential explanatory factors					
Percent urban land use	-2.29	ns	ns	ns	ns
Density of leaking underground fuel tanks	-2.92	ns	ns	ns	ns
Well depth	5.77	-3.00	-2.46	-3.98	-2.05
Depth to top of perforations	4.93	-3.49	-2.53	-3.65	-2.98
Dissolved oxygen	ns	3.18	ns	3.83	6.47
pH	ns	-3.75	ns	-4.37	-2.32
Aridity index ¹	-5.36	ns	ns	ns	-6.14
Field water temperature	5.36	-2.04	ns	ns	ns
Distance to the nearest hot spring listed in the USGS Mineral Resources Data System	-6.40	ns	ns	ns	ns
Distance to nearest geothermal well listed in the DOGGR database	-6.23	ns	ns	ns	ns
Water-quality constituents					
Boron ²	4.23	ns	ns	ns	-3.53
Arsenic ²	ns	-4.65	ns	-3.50	-3.36
Iron ²	ns	-2.79	ns	-2.32	-5.26
Manganese ²	ns	-2.89	ns	-3.48	-6.24
Chloroform ³	ns	2.00	ns	ns	3.36
Simazine ³	ns	2.35	ns	ns	2.11
Perchlorate ³	ns	ns	ns	ns	2.63

¹Aridity index is average annual precipitation (PRISM Group, Oregon State University, 2012) divided by average annual evapotranspiration (Flint and Flint, 2007).

²Constituents with ≥ 2% high aquifer-scale proportion.

³Organic constituents (special-interest constituent) with detection frequencies ≥ 10%, including detections below reporting levels.

The NOCO-IN study area generally had deeper wells and deeper depth to top-of-perforations than the NOCO-CO study area. Samples from the NOCO-IN study area also had higher groundwater temperatures, and the climate was drier in the NOCO-IN study area as indicated by a lower aridity index ([table 6](#)). The NOCO-IN study area sites generally were located closer to hot springs or geothermal wells. The NOCO-CO study area had a higher percentage of urban land use and density of LUFTs than the NOCO-IN study area ([table 6](#)).

In the NOCO study unit, relations were observed between well depth and groundwater age. The median depth of wells with groundwater classified as pre-modern was deeper than the depths of wells with groundwater classified as modern or mixed-ages ([table 6; figs. 8A, C](#)). In addition, groundwater classified as mixed-age was deeper than groundwater classified as modern. The median depth to the top-of-perforations of wells with groundwater classified as pre-modern also was deeper than the depth to the top-of-perforations of wells with groundwater classified as modern or mixed-age ([table 6; figs. 8B, D](#)). Well depths or depths to top-of-perforations in wells with groundwater samples classified as modern and mixed-age were not significantly different.

Groundwater ages for most wells perforated entirely at depths <70 ft bls (15 of 19 wells, 79 percent) were modern age ([figs. 8B, D](#)). Groundwater ages for wells with the top-of-perforations <70 ft but with the bottom-of-perforations ≥ 70 ft were modern or mixed-age. Groundwater ages for most wells perforated entirely at depths ≥ 70 ft were pre-modern or mixed-age.

The densities of septic tanks and LUFTs were positively correlated with percent urban land use and negatively correlated with natural land use ([table 5](#)).

Field water temperature was significantly positively correlated with well depth, depth to top-of-perforations, and the percentage of agricultural land use ([table 5](#)). Field water temperature was negatively correlated with the aridity index and distances to nearest hot spring and geothermal well. Groundwater temperatures also were significantly lower in sites with modern ages than in sites with mixed ages, indicating that groundwater temperatures generally increase with depth ([table 6](#)). The aridity index was positively correlated with distances to the nearest hot spring and geothermal well ([table 5](#)).

pH was positively correlated with well depth and depth to top-of-perforations ([table 5](#)), indicating increasing pH with increasing depth ([fig. 9A](#)). These correlations were expected based on previous studies in other aquifers in California (Jurgens and others, 2008) and reflect dissolution of primary aquifer minerals, causing the pH of groundwater to increase with depth and continued contact of groundwater with aquifer materials. pH values also were higher for groundwater with mixed or pre-modern ages and for groundwater with anoxic conditions ([table 6; fig. 9A](#)). DO concentrations were significantly greater in groundwater with modern or mixed ages than in groundwater with pre-modern ages ([table 6; fig. 9B](#)), indicating that DO has been consumed in older waters that have had more time to interact with oxidizable material in the groundwater flow system.

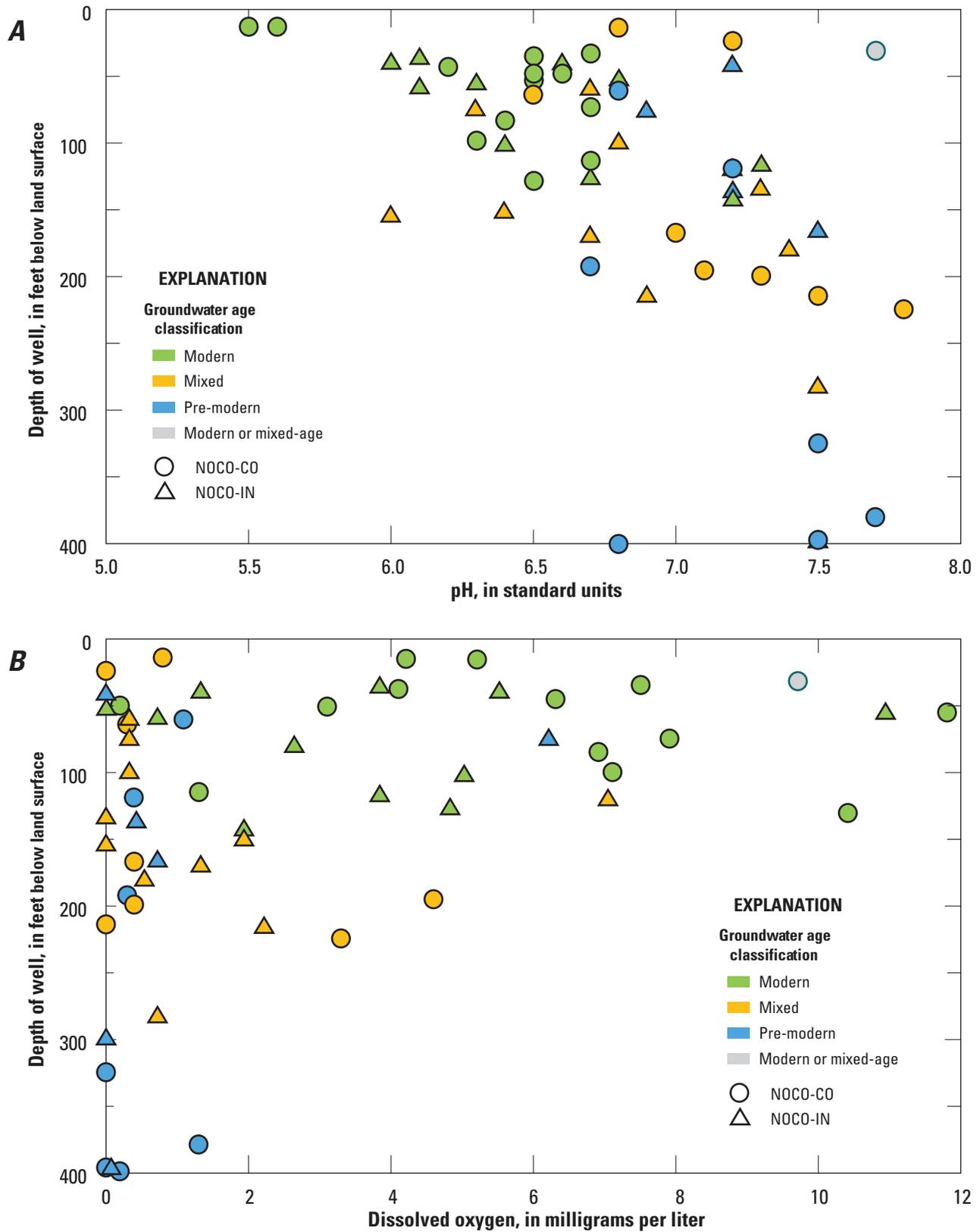


Figure 9. Groundwater age classification and well depth in relation to (A) pH and (B) dissolved oxygen concentrations, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area]

Status and Understanding of Water Quality

The status assessment was designed to identify the constituents or classes of constituents most likely to be of water-quality concern because of their high relative-concentrations or their prevalence. The assessment applies only to constituents having regulatory or nonregulatory health-based or aesthetic-based benchmarks established by the USEPA or the CDPH (California Department of Public Health, 2006, 2008, 2010; U.S. Environmental Protection Agency, 2008, 2009a, b, 2012). The spatially distributed, randomized approach to site selection and data analysis yields a view of groundwater quality in which all areas of the primary aquifer system are weighted equally; regions with a high density of groundwater use or with high density of potential contaminants were not preferentially sampled (Belitz and others, 2010).

The understanding assessment was designed to help answer the question of why selected constituents are, or are not, detected in the groundwater, and may improve understanding of how natural and anthropogenic factors affect groundwater quality. The assessment addresses a subset of the constituents selected for additional evaluation in the status assessment and is based on statistical correlations between water quality and a set of potential explanatory factors. The assessment was not designed to identify specific sources of specific constituents to specific sites.

The following discussion of the status and understanding assessment results is divided into two parts—inorganic constituents and organic constituents—and each part has a tiered structure. Each part begins with a survey of how many constituents were detected at any concentration in USGS-GAMA samples compared to the number analyzed and a graphical summary of the relative-concentrations of constituents detected in the grid sites. Aquifer-scale proportions are presented for the subset of constituents that met criteria for additional evaluation based on relative-concentration, or for organic (special-interest) constituents, prevalence. For constituents that have understanding assessment results, those results are presented immediately following the status assessment results for that constituent.

The aquifer-scale proportions calculated by using the spatially weighted approach were within the 90 percent confidence intervals for their respective grid-based aquifer proportions for all 13 of the constituents listed in [tables 7](#), [C1](#), and [C2](#), providing evidence that the grid-based approach yields statistically equivalent results to the spatially weighted approach.

Inorganic Constituents

Inorganic constituents typically occur naturally in groundwater, although their concentrations may be influenced by human activities as well as by natural factors. All 49 inorganic constituents analyzed by USGS-GAMA were detected at least once in the NOCO study unit. Of these 49 constituents, 26 had regulatory or nonregulatory health-based benchmarks, 8 had nonregulatory aesthetic-based benchmarks, and 15 had no established benchmarks ([table 2](#)).

Eight inorganic constituents were identified for additional evaluation in the status assessment because they were detected at moderate or high relative-concentrations in the USGS-GAMA grid sites. The constituents with health-based benchmarks are arsenic, barium, boron, nickel, and nitrate, and the constituents with aesthetic-based benchmarks are iron, manganese, and total dissolved solids (TDS) ([figs. 10](#), [11](#); [tables 3](#), [7](#)). Four of these inorganic constituents were selected for further evaluation in the understanding assessment because they were present at high relative-concentrations in >2 percent of the primary aquifer system: boron, arsenic, iron, and manganese.

Inorganic Constituents with Health-Based Benchmarks

Inorganic constituents with health-based benchmarks, as a group, had high relative-concentrations in 10.3 percent, moderate relative-concentrations in 13.8 percent, and low relative-concentrations (or non-detections) in 75.9 percent of the primary aquifer system ([table 8A](#)).

Trace Elements

Trace elements, as a class, were detected at high relative-concentrations (for one or more constituents) in 10.3 percent of the primary aquifer system, moderate relative-concentrations in 8.6 percent, and low relative-concentrations in 81.1 percent ([table 8A](#)). Boron and arsenic were detected at high relative-concentrations in more than 2 percent of the primary aquifer system, and barium was detected at high relative-concentrations in 1.7 percent of the primary aquifer system ([table 7](#)). The spatial distributions of selected trace elements for USGS-GAMA grid sites and the CDPH sites are shown in [figures 12 and 13](#). Note that boron was not sampled by the CDPH from June 1, 2006, to June 1, 2009.

Table 7. Aquifer-scale proportions calculated using grid-based and spatially weighted methods for those constituents that met criteria for additional evaluation in the status assessment, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[**Relative-concentration categories:** *high*, concentration greater than water-quality benchmark; *moderate*, concentration is less than the benchmark and is greater than or equal to 0.1 of the benchmark (for organic and special-interest constituents) or 0.5 of the benchmark (for inorganic constituents); *low*, concentration less than 0.1 of benchmark (for organic constituents and special-interest constituents) or 0.5 of the benchmark (for inorganic constituents). **Abbreviations:** GAMA, Groundwater Ambient Monitoring and Assessment Program; CDPH, California Department of Public Health; TDS, total dissolved solids; MTBE, methyl *tert*-butyl ether; µg/L, micrograms per liter]

Constituent	Raw detection frequency, in percent ¹			Spatially weighted aquifer-scale proportion, in percent ¹			Grid-based aquifer-scale proportion, in percent			90 percent confidence interval for grid-based high proportion, in percent ²	
	Number of sites	Moderate	High	Number of cells	Moderate aquifer proportion	High aquifer proportion	Number of sites	Moderate aquifer proportion	High aquifer proportion	Lower limit	Upper limit
Inorganic constituents with health-based benchmarks											
Trace Elements											
Arsenic	130	3.1	3.1	58	6.0	3.5	58	6.9	3.4	1.0	9.2
Barium	131	0	0.8	58	0	1.7	58	0	1.7	0.3	6.5
Boron ³	58	5.2	8.6	58	5.2	8.6	58	5.2	8.6	4.0	16.2
Nickel	133	0.8	0	58	1.7	0	58	1.7	0	0	3.2
Nutrients											
Nitrate	220	1.8	0	58	1.4	0	57	3.5	0	0	3.3
Inorganic constituents with aesthetic-based benchmarks											
Iron	121	6.6	21.5	58	6.4	25.7	58	5.2	25.9	17.4	33.7
Manganese	123	4.1	30.9	58	7.8	41.2	58	8.6	39.7	29.6	48.0
TDS	118	5.1	0	58	6.9	0	58	6.9	0	0	3.2
Organic constituents with health-based benchmarks											
Volatile organic compounds											
Chloroform (Trichloromethane)	118	1.7	0	58	2.2	0	58	1.7	0	0	3.2
Vinyl chloride (Chloroethene)	115	0.9	0	58	1.7	0	58	1.7	0	0	3.2
MTBE ⁴	134	0	0.7	58	0	0.2	58	0	0	0	3.2
Pesticides and pesticide degradates											
Simazine	123	0	0	58	0	0	58	0	0	0	3.2
Special-interest constituent with health-based benchmark											
Perchlorate	162	0.6	0	58	0.6	0	58	1.7	0	0	3.2

¹Based on most recent analyses for each CDPH site during June 1, 2006–June 1, 2009, combined with data from USGS-GAMA grid sites.

²Based on the Jeffrey's interval for the binomial distribution (Brown and others, 2001).

³Boron was not sampled by the CDPH during June 1, 2006–June 1, 2009.

⁴High relative-concentrations of MTBE were reported multiple times in the CDPH water-quality database from one site in the NOCO-IN study area during the current period of study. Sixteen groundwater samples were collected by the CDPH between August 16, 2006, and May 20, 2009, from this site for analysis of MTBE; 15 had concentrations of MTBE between 13 and 21 µg/L (MCL-CA = 13 µg/L), and 1 was a non-detection (most recent value, May 20, 2009). On that basis, it was determined that the non-detection of MTBE reported as the most recent value in the CDPH water-quality database does not represent present-day conditions. The MTBE sample collected on April 30, 2009 (17 µg/L) was substituted in order to more accurately reflect present-day conditions.

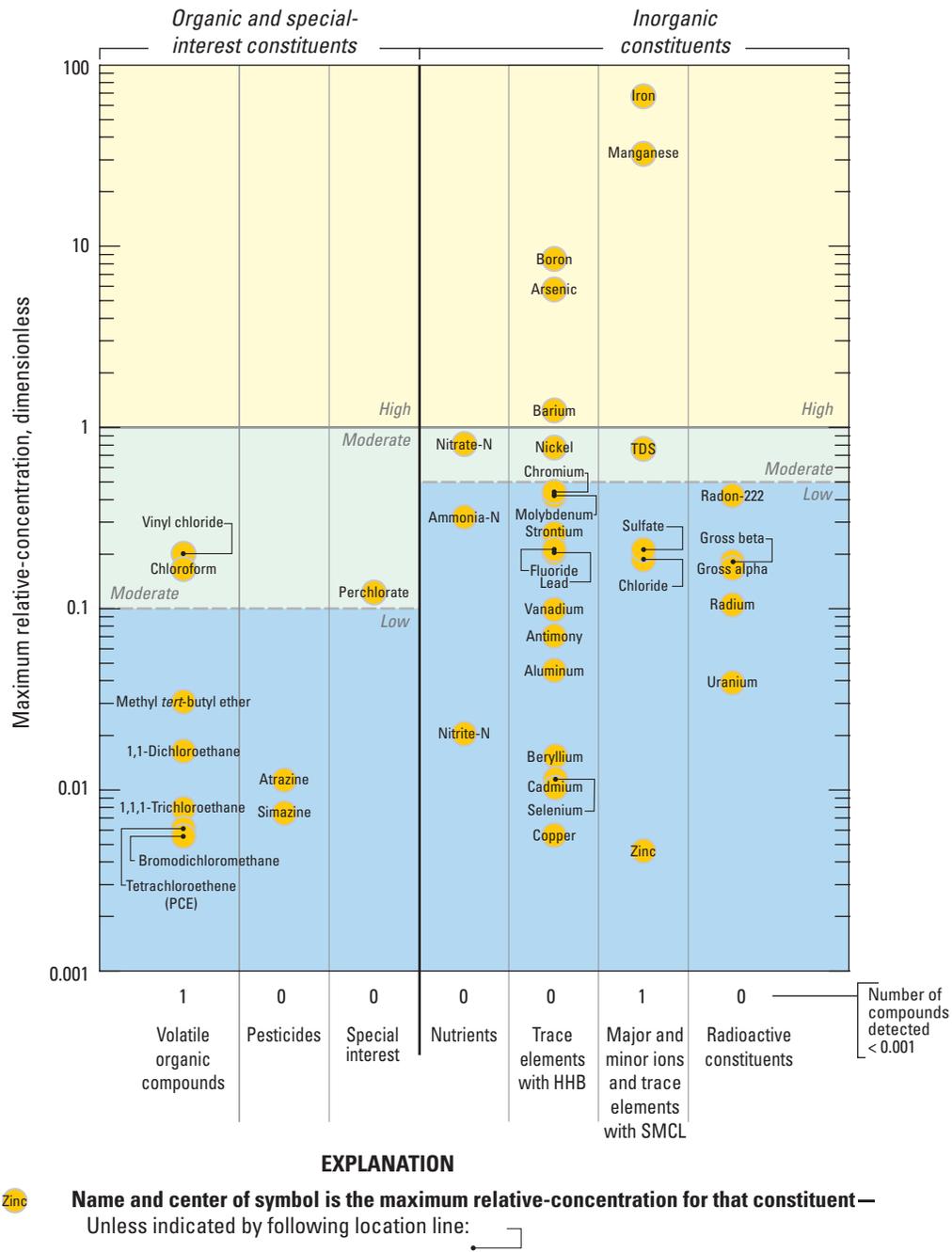


Figure 10. Maximum relative-concentration of constituents detected in USGS-GAMA grid sites, by constituent class, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

A

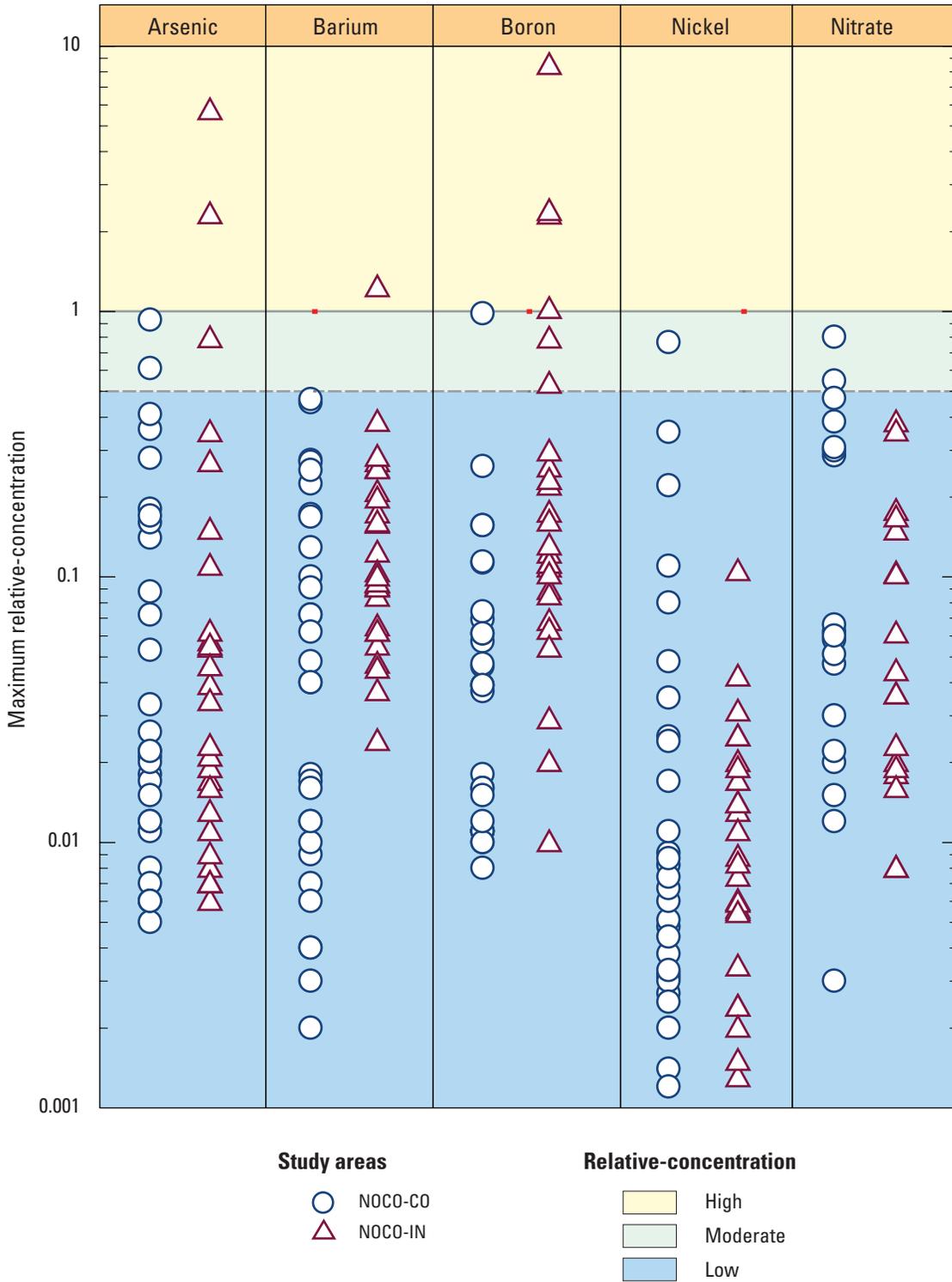


Figure 11. Relative-concentrations of (A) selected constituents with health-based benchmarks and (B) constituents with aesthetic-based benchmarks in USGS-GAMA grid sites, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area; TDS, total dissolved solids]

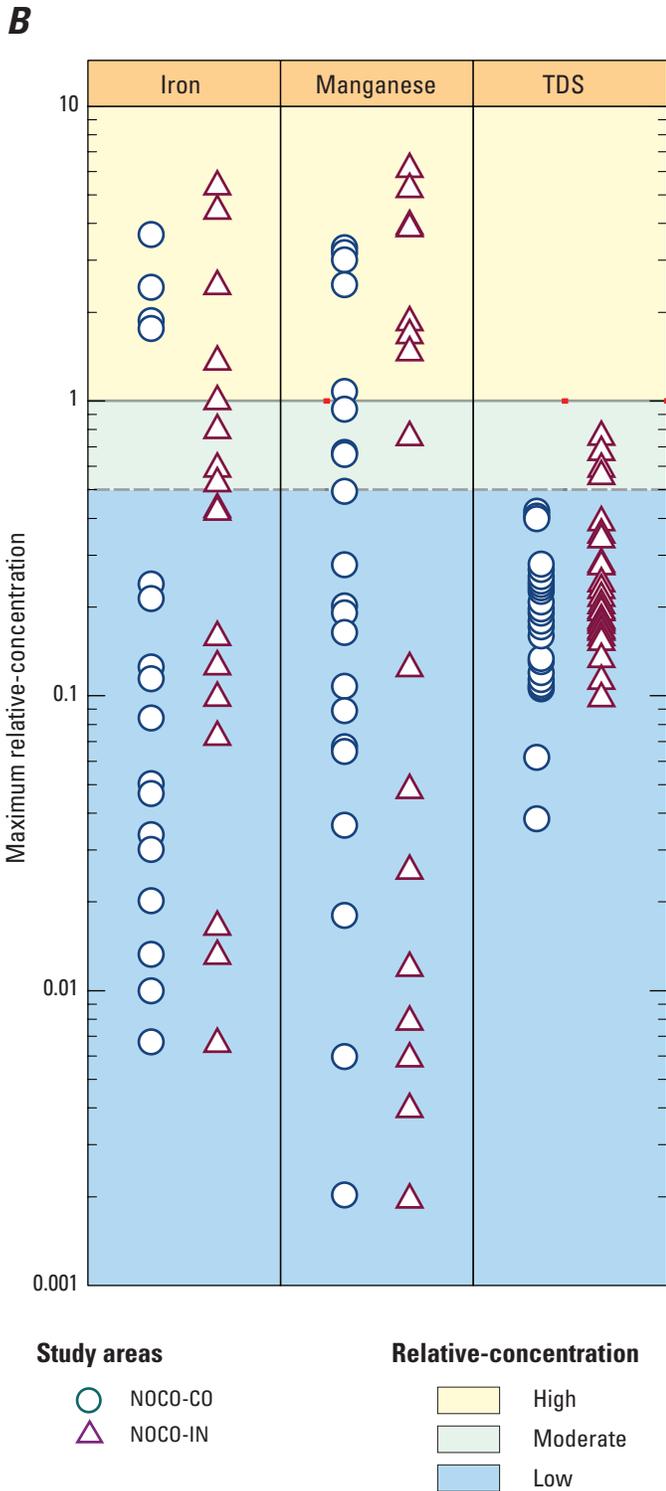


Figure 11. —Continued

Understanding Assessment for Boron

Boron is a naturally occurring semi-metallic element with high solubility in water. Natural sources of boron include igneous rocks, such as granite and pegmatite (as the mineral tourmaline), and evaporite minerals, such as borax, kernite, and colemanite (Hem, 1985; Reimann and de Caritat, 1998). Seawater contains 4,500 micrograms per liter ($\mu\text{g/L}$) of boron (Summerhayes and others, 1996), and groundwater from marine sediments and shales average about $0.1 \mu\text{g/L}$ boron (Ryan and Langmuir, 1993). Boron also is associated with shallow thermal springs, deep thermal waters, and volcanic activity (Hem, 1989; Webster and Nordstrom, 2003; Kulongoski and others, 2010; Forrest and others, 2013). Boron can occur in wastewater because borax is a component of many detergents. Other anthropogenic uses of boron compounds include semiconductors, insecticides, preservatives, chemical reagents, and fertilizers. Boron is an essential nutrient for plants, but is toxic to plants at high concentrations (Hem, 1989). The comparison benchmark used for boron in this study was the CDPH NL-CA of $1,000 \mu\text{g/L}$ (California Department of Public Health, 2010). At concentrations greater than the HAL-US of $6,000 \mu\text{g/L}$, boron may adversely affect fetal development (U.S. Environmental Protection Agency, 2008). Boron is highly mobile because no mineral has a low enough solubility to provide an upper limit to its concentration range (Ryan and Langmuir, 1993).

Boron was detected at high relative-concentrations in 8.6 percent of the primary aquifer system and at moderate relative-concentrations in 5.2 percent (table 7). Boron concentrations generally are higher in the NOCO-IN study area than in the NOCO-CO study area (table 6). All of the high relative-concentrations of boron were detected in the NOCO-IN study area, but moderate relative-concentrations were detected in both study areas (figs. 11A, 12; tables C1, C2). Boron concentrations were found to be negatively correlated with DO (table 9) and were significantly higher in groundwater with anoxic rather than oxic conditions (table 6; fig. 14D). All of the sites with high and moderate relative-concentrations of boron had anoxic groundwater conditions (table B5).

Multiple lines of evidence suggest that the elevated boron concentrations found in the NOCO study unit are most likely due to groundwater mixing with hydrothermal systems within the NOCO-IN study area, including (1) elevated groundwater temperatures, (2) groundwater chemistry characteristics similar to those of geothermal waters, and (3) proximity to known geothermal areas.

Samples with high and moderate relative-concentrations of boron have elevated groundwater temperatures. Boron concentrations were significantly positively correlated with field water temperature (table 9; fig. 14A). The median temperature ($18.0 \text{ }^\circ\text{C}$) for the 7 samples with high and moderate relative-concentrations of boron was significantly greater than the median temperature ($14.5 \text{ }^\circ\text{C}$) for the other 51 groundwater samples in the NOCO study unit (Wilcoxon rank-sum: $p = 0.006$). Specifically, the site with the highest

Table 8A. Summary of aquifer-scale proportions for inorganic constituent classes, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[Relative-concentration categories: *high*, concentration of at least one constituent in the class greater than its benchmark; *moderate*, concentration of at least one constituent in the class was greater than half of its benchmark and no constituents in the class had concentrations greater than their benchmarks; *low*, concentrations of all constituents in the class were less than or equal to half of the benchmark, including non-detections. **Abbreviation:** GAMA, Groundwater Ambient Monitoring and Assessment Program]

Constituent class	Aquifer-scale proportion, in percent		
	Low values	Moderate values	High values
Inorganic constituents with health-based benchmarks			
Trace elements ¹	81.1	8.6	10.3
Nutrients	96.5	3.5	0
Uranium and other radioactive constituents	100	0	0
<i>Any inorganic constituent with health-based benchmark</i>	75.9	13.8	10.3
Inorganic constituents with aesthetic-based benchmarks			
Iron, manganese, silver, and zinc	51.7	8.6	39.7
Total dissolved solids, specific conductance, sulfate, and chloride	93.1	6.9	0
<i>Any inorganic constituent with aesthetic-based benchmark</i>	50.0	10.3	39.7

¹Includes the minor element fluoride because it has a health-based benchmark.

Table 8B. Summary of aquifer-scale proportions for organic and special-interest constituent classes, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[Relative-concentration categories: *high*, concentration of at least one constituent in the class greater than its benchmark; *moderate*, concentration of at least one constituent in the class was greater than one-tenth of its benchmark and no constituents in the class had concentrations greater than their benchmarks; *low*, concentrations of all constituents in the class were less than or equal to one-tenth of the benchmark. **Abbreviation:** GAMA, Groundwater Ambient Monitoring and Assessment Program]

Constituent class	Aquifer-scale proportion, in percent			
	Not detected	Low values	Moderate values	High values
Organic constituents with health-based benchmarks				
Volatile organic compounds				
Solvents	91.4	8.6	0	0
Organic synthesis reagents	98.3	0	1.7	0
Gasoline components	92.9	6.9	0	¹ 0.2
Trihalomethanes	75.9	22.4	1.7	0
<i>Any volatile organic compound with health-based benchmark</i>	61.9	36.2	1.7	¹ 0.2
Pesticides and pesticide degradates ²	89.7	10.3	0	0
<i>Any organic constituent with health-based benchmark</i>	58.4	39.7	1.7	¹ 0.2
Special-interest constituent with health-based benchmark				
Perchlorate	75.9	22.4	1.7	0

¹Spatially weighted aquifer-scale proportion value.

²Only herbicides were detected.

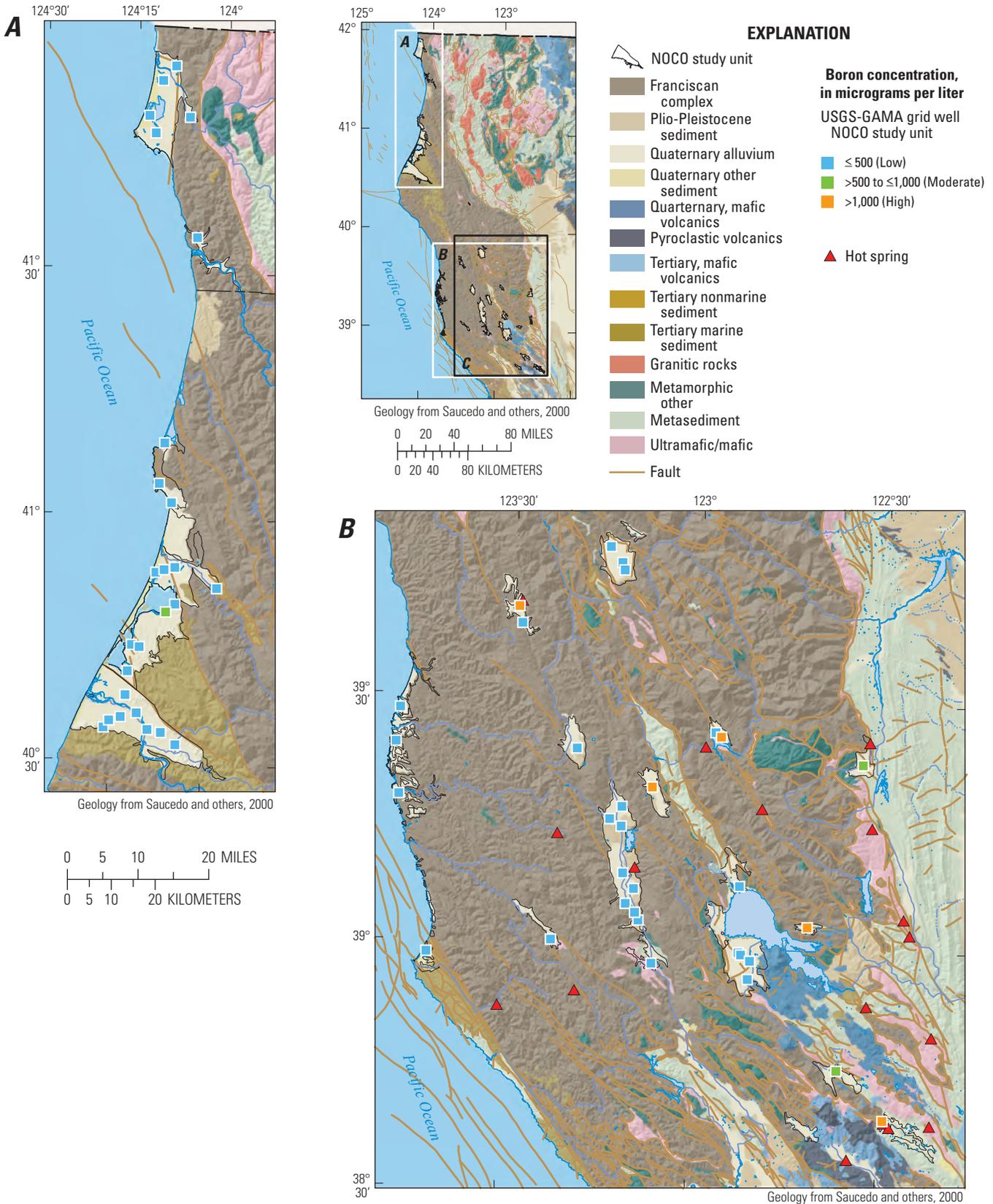


Figure 12. Relative-concentrations of boron in USGS-GAMA grid sites and geologic features in the (A) Coastal Basins study area and (B) Interior Basins study area, and (C) geologic formations, locations of hot springs and geothermal wells, and approximate boundaries of the Napa and Sonoma Valleys and the Geysers Steam Field, Northern Coast Ranges (NOCO) study unit, 2009, California GAMA Priority Basin Project.

C

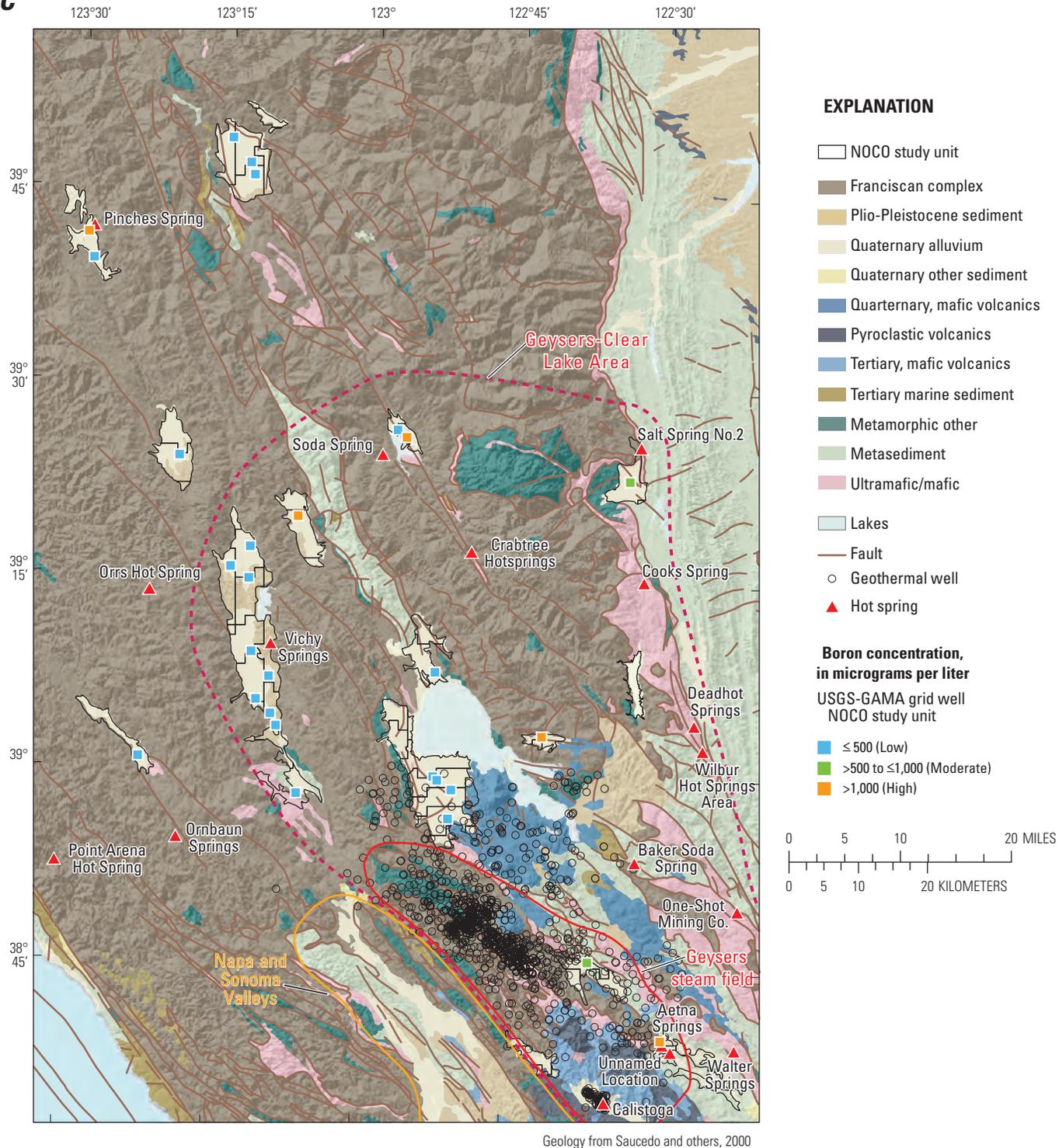


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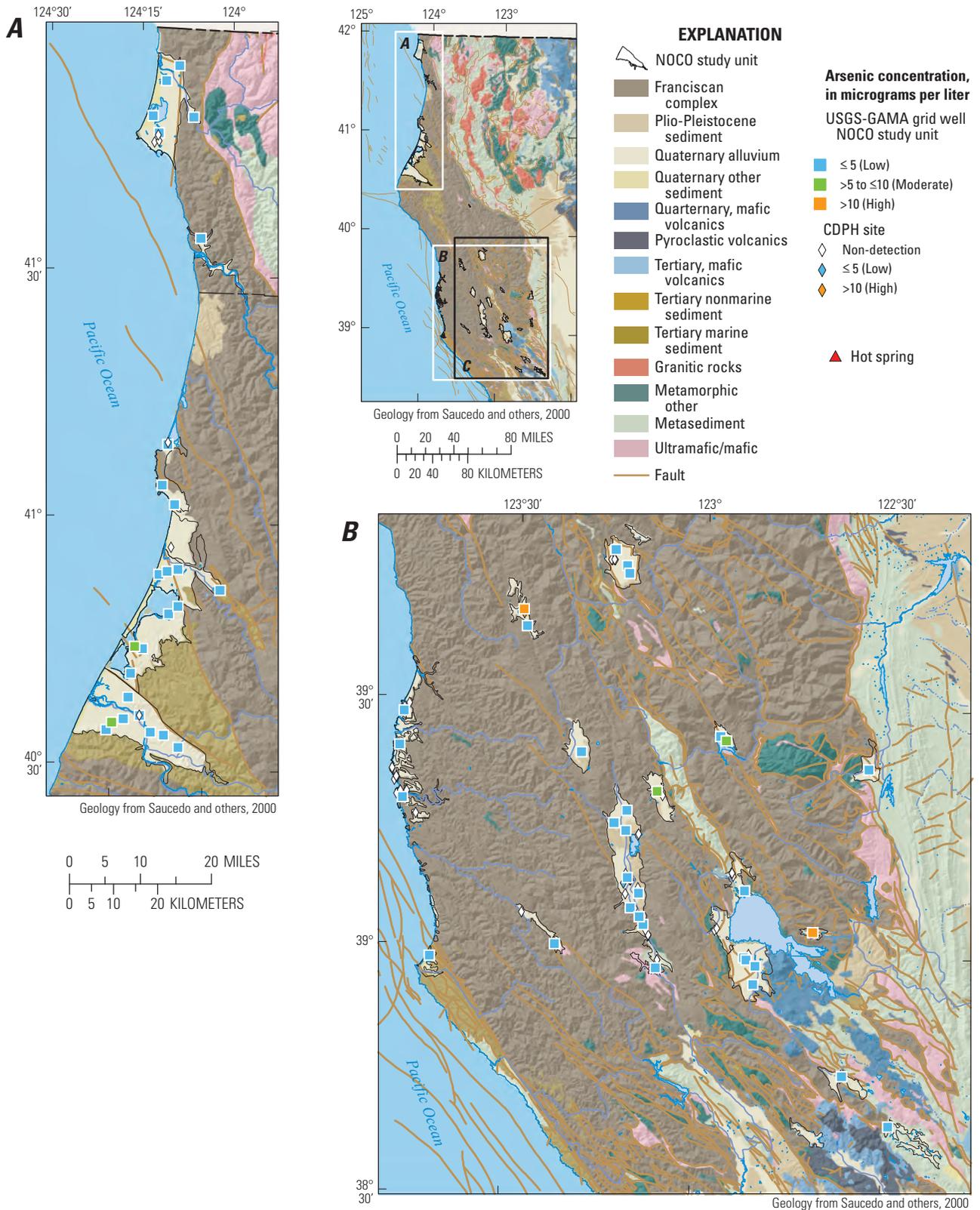


Figure 13. Relative-concentrations of arsenic in USGS-GAMA grid sites, CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database, and geologic features in the (A) Coastal Basins study area and (B) Interior Basins study area and (C) geologic formations, locations of hot springs and geothermal wells, and approximate boundaries of the Napa and Sonoma Valleys and the Geysers Steam Field Northern Coast Ranges (NOCO) study unit, 2009, California GAMA Priority Basin Project.

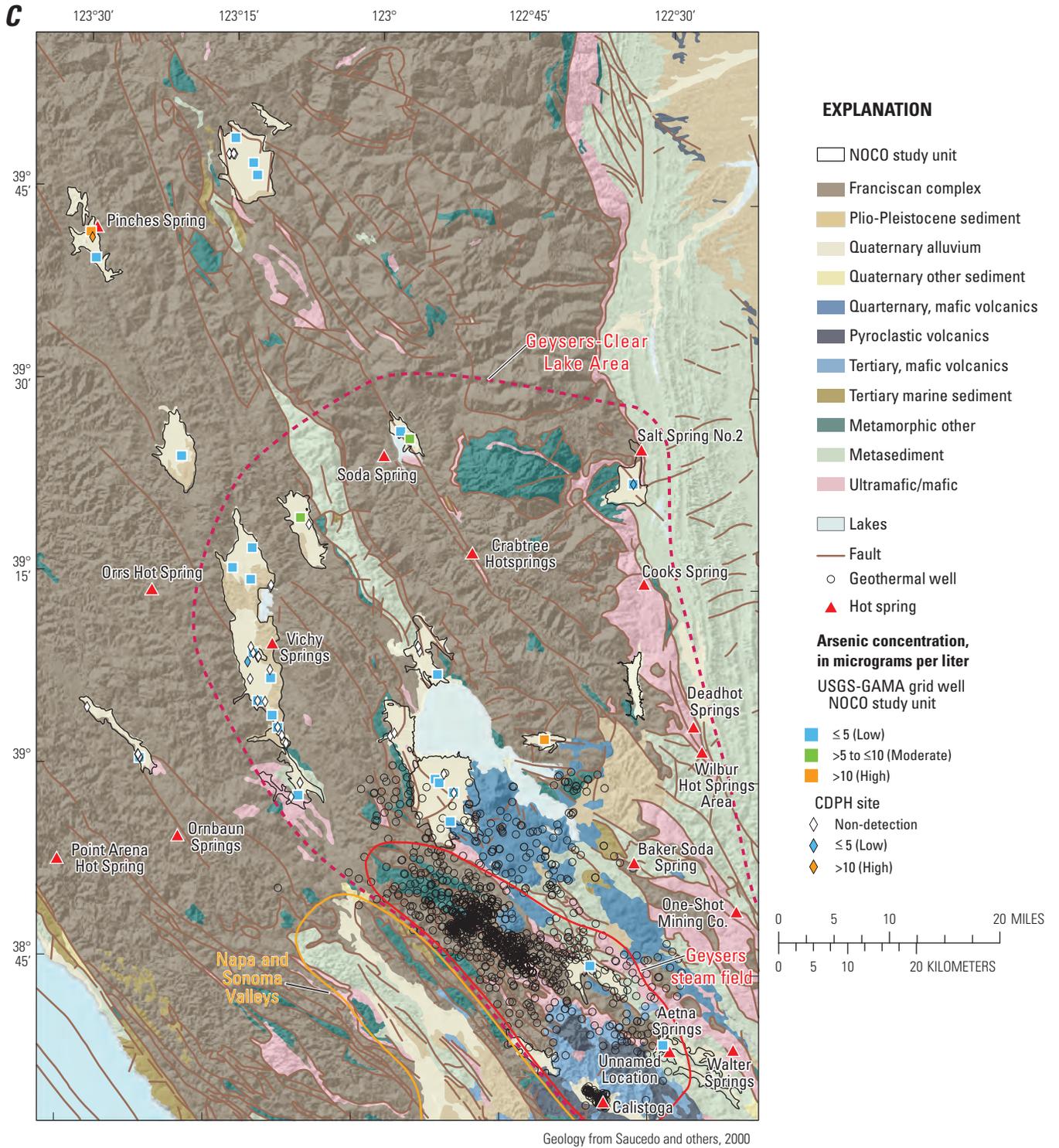
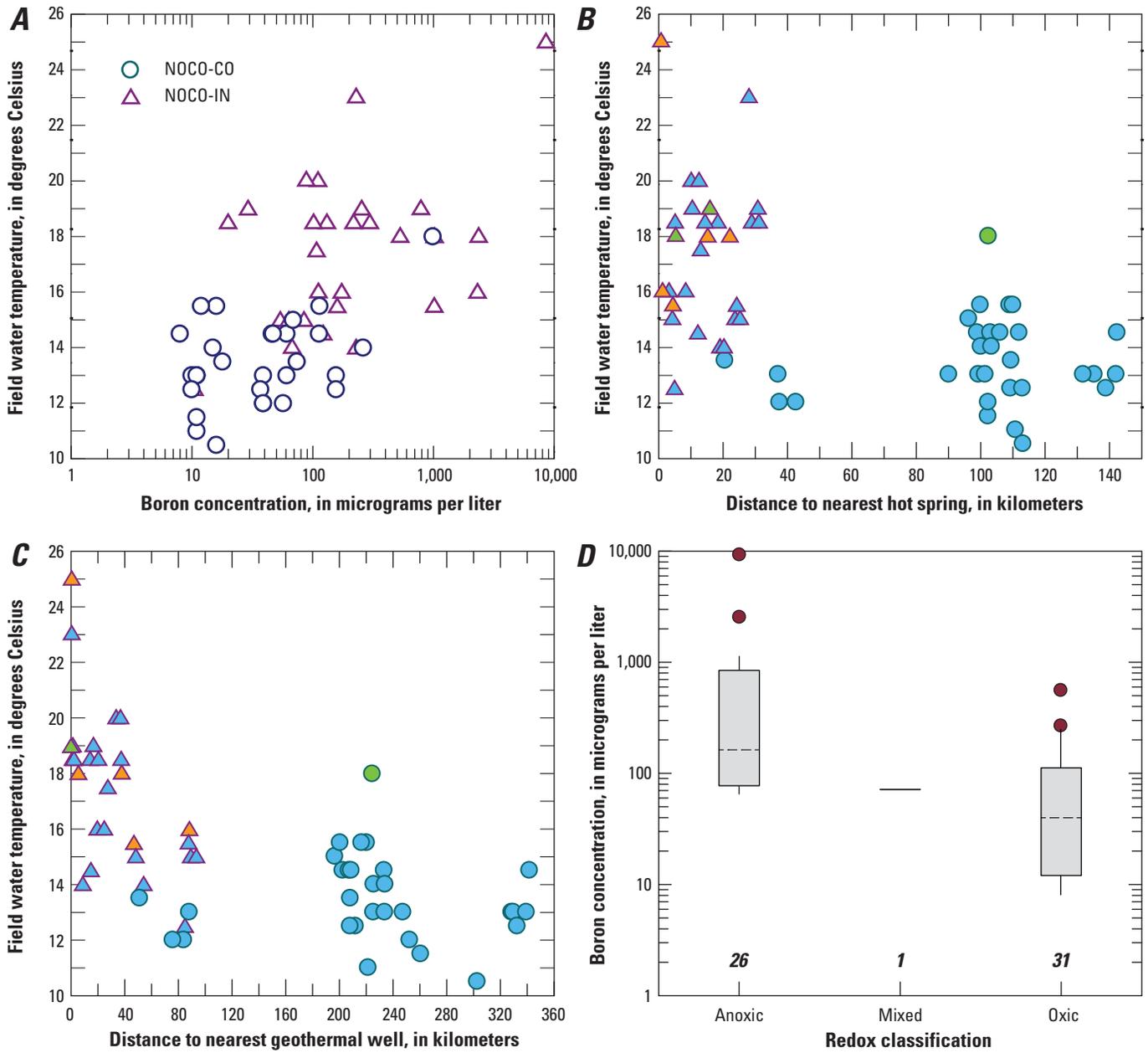


Figure 13. —Continued



EXPLANATION

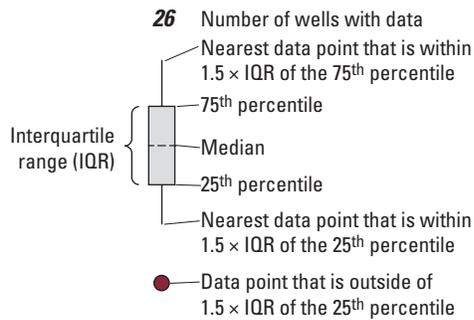
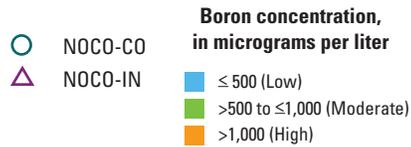


Figure 14. Boron concentrations relative to (A) field water temperature, (B) field water temperature and distance to nearest hot spring, (C) field water temperature and distance to nearest geothermal well, and (D) redox classification in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area]

boron concentration (NOCO-IN-22; 8,470 $\mu\text{g/L}$) had a field water temperature of 25.0 $^{\circ}\text{C}$ (table B3). These results follow the findings in previous studies in the Geysers–Clear Lake area, which classified groundwater as being under geothermal influence if temperatures were $>19.0^{\circ}\text{C}$ (Goff and others, 1993; Donnelly-Nolan and others, 1993).

Samples with high and moderate relative-concentrations of boron exhibit groundwater chemistry characteristics similar to those of geothermal waters. Groundwater data appear to show a similar pattern of occurrences of hydrothermal system indicators as that noted in the Geysers–Clear Lake area (Donnelly-Nolan and others, 1993; Goff and others, 1993; Peters, 1993) and in the Napa and Sonoma Valleys (Forrest and others, 2013). The shared characteristics of geothermal waters in these hydrothermal systems include raised water temperatures, near neutral pH, elevated concentrations of boron, arsenic, and lithium (in varied degrees), and high helium isotope ratios.

Samples with high and moderate relative-concentrations of boron all had near-neutral pH values, which ranged from 6.0 to 7.5 and had a median of 6.8 (table B5). Boron and arsenic concentrations were significantly correlated with each other (Spearman's *rho* test, $p = 0.002$, $rho = 0.56$). Two of the sites with high relative-concentrations of boron (NOCO-IN-05 and NOCO-IN-27) also had high relative-concentrations of arsenic, and two additional sites with high relative-concentrations of boron (NOCO-IN-11 and NOCO-IN-15) had moderate relative-concentrations of arsenic (figs. 12, 13). Boron and lithium concentrations were also significantly correlated with each other (Spearman's *rho* test, $p = 0.001$, $rho = 0.43$). Helium isotope ratios ($^3\text{He}/^4\text{He}$) are often discussed as R/Ra rather than as the ratio itself. R/Ra is the ratio in the sample (R) divided by the $^3\text{He}/^4\text{He}$ ratio in the atmosphere ($Ra = 1.4 \times 10^{-6}$). R/Ra >1.0 generally indicates contribution from the Earth's mantle (Kulongoski and others, 2003), and median mantle helium in the western United States has an R/Ra = 6.0 (Dodson and others, 1998). The sample from NOCO-IN-22 (the sample with the highest boron concentration) had an R/Ra = 5.0 (table E1).

Samples with high and moderate relative-concentrations of boron are located near known hydrothermal areas. Proximity of USGS-GAMA sites to known hydrothermal areas was assessed in three ways: (1) sample sites located within the approximate physical boundaries of geothermal areas, (2) distance from hot spring listed in the USGS MRDS, and (3) distance from geothermal well listed in the DOGGR database.

Four of the seven samples with high and moderate relative-concentrations of boron are from sites that are spatially located within the approximate boundaries of the Geysers–Clear Lake area (figs. 12B, C). Specifically, NOCO-IN-27 is located in the central portion of the study area, within the main mass of the Clear Lake volcanic field; NOCO-IN-22 and NOCO-IN-25 are located in the southern portion of the study area, near the Geysers steam field; and NOCO-IN-26 is located in the portion of the study area, in the eastern Clear Lake region (fig. 12B).

Boron concentrations were negatively correlated with the distance to nearest hot spring listed in the USGS MRDS and distance to nearest geothermal well listed in the DOGGR database (table 9; figs. 14B, C). The median distance from the nearest hot spring for the 7 samples with high and moderate relative-concentrations of boron was 5.3 km; the median distance for the other 51 groundwater samples was 66.1 km. Three of the sites with high relative-concentrations of boron were relatively near named hot springs (NOCO-IN-05: Pinches Spring; NOCO-IN-15: Soda Spring; NOCO-IN-22: Aetna Springs) as was one of the sites with a moderate relative-concentration (NOCO-IN-26: Salt Spring No. 2; U.S. Geological Survey, 2005; fig. 12; table B3). The median distance from nearest geothermal well for the 7 samples with high and moderate relative-concentrations of boron was 38.5 km; the median distance for the other 51 groundwater samples was 145 km. Specifically, NOCO-IN-22 was within 1 km of a geothermal well, and NOCO-IN-25 was 0.37 km from a geothermal well (California Department of Conservation, 2013).

There does not appear to be a causative relation between boron concentrations and aridity index, although they had a significant negative correlation (table 9). This result is most likely a function of aridity index being correlated with the geothermal explanatory factors (groundwater temperature and distance to nearest hot spring and geothermal well) rather than a causative relation that is explanatory of boron distribution in groundwater in the NOCO study unit.

The elevated groundwater temperatures, close relations among boron, arsenic, and lithium concentrations in the groundwater, helium isotope ratios, and proximity to known geothermal areas suggest that the primary source of boron in the NOCO study unit is geothermal waters.

Understanding Assessment for Arsenic

Arsenic is a naturally occurring semi-metallic trace element. Potential sources of arsenic to groundwater are natural or anthropogenic. Natural sources include the dissolution of arsenic-bearing minerals and desorption of arsenic from mineral surfaces. Potential anthropogenic sources of arsenic include mining of copper and gold ores, coal combustion, arsenical pesticides, arsenical veterinary pharmaceuticals, and wood preservatives (Welch and others, 2000). The MCL-US for arsenic was lowered from 50 to 10 $\mu\text{g/L}$ in 2002, and chronic exposure to arsenic concentrations between 10 and 50 $\mu\text{g/L}$ in drinking water has been linked to increased cancer risk and to non-cancerous effects including skin damage and circulatory problems (U.S. Environmental Protection Agency, 2012). Arsenic solubility increases with increasing water temperature, such that geothermal systems often exhibit high arsenic concentrations (Ballantyne and Moore, 1988; Goff and others, 1993; Peters, 1993; Webster and Nordstrom, 2003; Forrest and others, 2013).

Arsenic was detected at high relative-concentrations in 3.4 percent of the primary aquifer system and at moderate relative-concentrations in 6.9 percent (table 7). All of the

high relative-concentrations of arsenic were detected in the NOCO-IN study area (figs. 11, 13; tables C1, C2). Arsenic concentrations were significantly lower in modern groundwater than in groundwater with mixed or pre-modern ages (table 6; fig. 15A). Additionally, arsenic concentrations were positively correlated with well depth and depth to top-of-perforations (table 9). Arsenic concentrations were negatively correlated to DO (table 9), and arsenic concentrations were significantly higher in groundwater with anoxic conditions rather than oxic conditions (table 6).

Arsenic mobilization and distribution in groundwater is affected by the oxidation-reduction (redox) and pH conditions of the groundwater system. Arsenic concentrations are commonly found to be positively correlated with pH as a result of the desorption of arsenic from aquifer sediments with increasing pH (Belitz and others, 2003; Welch and others, 2006). Previous investigations and reviews of arsenic (for example, Welch and others, 2000, 2006; Frankenberger, 2002; Ravenscroft and others, 2009; Forrest and others, 2013) have attributed elevated arsenic concentrations in groundwater to three mechanisms: (1) the release of arsenic from dissolution of iron or manganese oxides under iron- or manganese-reducing (anoxic) conditions, (2) groundwater mixing with hydrothermal systems, and (3) arsenic desorption

from aquifer sediments or inhibition of arsenic sorption to aquifer sediments as a result of oxic and alkaline groundwater conditions (pH values ≥ 8.0).

Whether arsenic was released by dissolution under anoxic conditions, groundwater mixing with hydrothermal systems, or desorption under oxic, high-pH conditions, the accumulation of elevated concentrations of arsenic in groundwater also requires favorable hydrologic conditions. Arsenic accumulation is favored by longer contact times between groundwater and aquifer materials and minimizing the amount of flushing of the system (Smedley and Kinniburgh, 2002).

Evidence for the first mechanism, release of arsenic under reducing conditions, in the NOCO study unit includes the association of high and moderate concentrations of arsenic with groundwater having iron- or manganese-reducing conditions. Both of the sites having high relative-concentrations of arsenic (NOCO-IN-05 and NOCO-IN-27) and three of the four sites having moderate relative-concentrations (NOCO-IN-11, NOCO-IN-15, and NOCO-CO-21) had anoxic groundwater conditions (table B5). For these five sites with anoxic conditions and high or moderate relative-concentrations of arsenic, pH values ranged from 6.0 to 7.1. Four of these five sites also had high relative-concentrations of iron and manganese (all but NOCO-IN-15). The association of high iron and manganese

Table 9. Results of non-parametric (Spearman's *rho* method) analysis for correlations between selected water-quality constituents and potential explanatory factors, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[**Explanation:** Potential explanatory factors without significant correlations are not shown on this table. Spearman's *rho* test used and *rho* values shown. Significance level ($p \leq 0.05$); Correlations with $rho > 0.2$ and (or) p -values ≤ 0.05 are shown in **bold**. **Abbreviations:** GAMA, Groundwater Ambient Monitoring and Assessment Program; USGS, U.S. Geological Survey; \leq , less than or equal to; $>$, greater than; \geq , greater than or equal to; ns, no significant correlation between factors; black text, significant positive correlation; red text, significant negative correlation; DOGGR, Division of Oil, Gas, and Geothermal Resources (California Department of Conservation); %, percent]

Constituent	Percent agricultural land use ¹	Well depth	Depth to top of perforations	Dissolved oxygen (DO)	pH	Field water temperature	Aridity index	Distance to nearest hot spring listed in the USGS Mineral Resources Data System	Distance to nearest geothermal well listed in the DOGGR database
<i>rho</i> : Spearman's correlation statistic									
Inorganic constituents with health-based benchmarks									
Boron ²	ns	ns	ns	-0.54	ns	0.62	-0.68	-0.53	-0.64
Arsenic ²	ns	0.34	0.58	-0.41	0.51	ns	ns	ns	ns
Inorganic constituents with aesthetic-based benchmarks									
Iron ²	ns	ns	ns	-0.60	ns	ns	ns	ns	ns
Manganese ²	ns	ns	0.37	-0.73	ns	ns	ns	ns	ns
Organic constituents with health-based benchmarks									
Chloroform ³	ns	-0.31	-0.56	0.46	-0.31	ns	ns	ns	ns
Simazine ³	ns	-0.37	ns	ns	-0.30	ns	ns	ns	ns
Special-interest constituent with health-based benchmark									
Perchlorate ³	0.33	ns	-0.39	0.27	ns	ns	ns	ns	ns

¹Land-use percentages are within circles with buffers of 500 meters centered around each site included in analysis.

²Constituents with $\geq 2\%$ high aquifer-scale proportion.

³Organic constituents (or constituent of special interest) with detection frequencies at any concentration $\geq 10\%$.

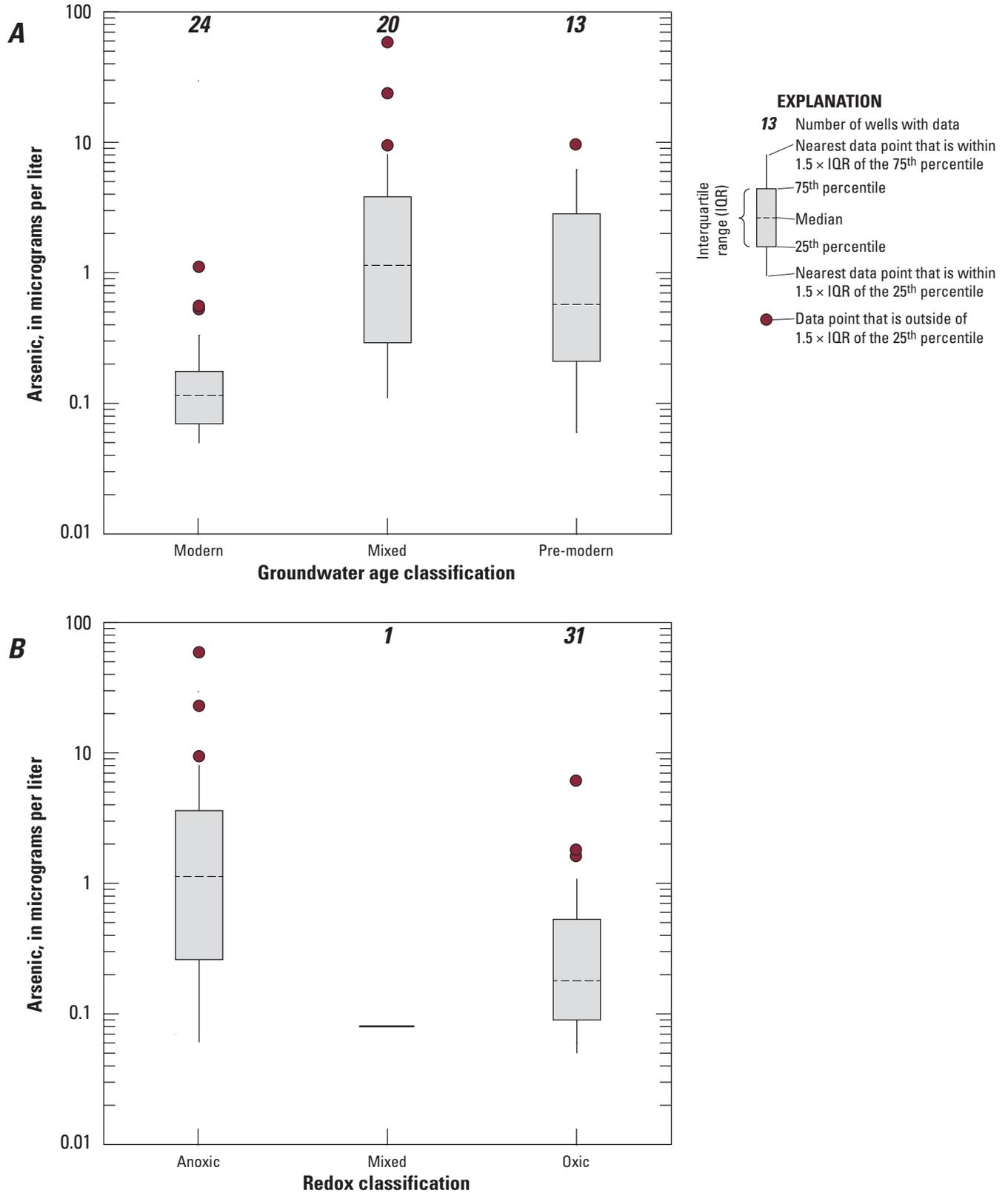


Figure 15. Arsenic concentrations relative to (A) groundwater age classification and (B) redox classification in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

relative-concentrations, low DO (anoxic conditions), and pH <7.0 at NOCO-IN-05, NOCO-IN-11, and NOCO-IN-27 suggests that reductive dissolution of iron and manganese oxides may account for portions of the high and moderate relative-concentrations of arsenic detected at these sites.

Evidence for the second mechanism, groundwater mixing with hydrothermal systems includes the association of high and moderate relative-concentrations of arsenic co-occurring with hydrothermal system indicators. As previously mentioned, arsenic and boron concentrations were significantly correlated with one other. Four of the sites in the NOCO-IN study area having high (NOCO-IN-05 and NOCO-IN-27) and moderate (NOCO-IN-11 and NOCO-IN-15) relative-concentrations of arsenic also had high relative-concentrations of boron (figs. 12, 13). Arsenic and lithium concentrations also were significantly correlated (Spearman's *rho* test, $p = 0.002$, $rho = 0.41$).

There is minimal evidence for the third mechanism, desorption of arsenic from aquifer sediments or inhibition of arsenic sorption to aquifer sediments with increasing pH. Arsenic concentrations were positively correlated with pH (table 9). Because only one site (NOCO-CO-03) had moderate relative-concentrations of arsenic, with oxic and relatively alkaline conditions (pH of 7.7; table B5), the relation of arsenic concentrations and pH is most likely a result of the relation of pH to anoxic conditions, well depth, and groundwater age.

In summary, high and moderate relative-concentrations of arsenic can result from release of arsenic from aquifer sediments under iron- and manganese-reducing conditions, contribution from hydrothermal systems (when present), and accumulation with groundwater age, or from other sources that are not fully understood.

Nutrients

Nutrients, as a class, were not detected at high relative-concentrations in the primary aquifer system of the NOCO study unit. Moderate relative-concentrations were detected in 3.5 percent of the primary aquifer system, and low relative-concentrations were detected in 94.8 percent of the primary aquifer system (table 8A). Moderate relative-concentrations of nitrate occurred in the NOCO-CO study area (table C2; Mathany and others, 2011).

Uranium and Other Radioactive Constituents

The relative-concentrations of uranium and other radioactive constituents are low in the primary aquifer system of the NOCO study unit (table 8A). No detections of uranium or other radioactive constituents were greater than their associated water-quality benchmark during the current period of study or in the historical data (table 4).

Inorganic Constituents with Aesthetic-Based Benchmarks

Inorganic constituents with aesthetic-based benchmarks, as a class, were detected at high relative-concentrations (for one or more constituents) in 39.7 percent, moderate relative-concentrations in 10.3 percent, and low relative-concentrations in 50.0 percent of the primary aquifer system (table 8A). High (39.7 percent) and moderate (8.6 percent) relative-concentrations of trace metals (iron and manganese) accounted for all of the high and most of the moderate aquifer-scale proportions. Inorganic constituents with aesthetic-based benchmarks that are indicators of salinity (TDS, field specific conductance, sulfate, and chloride) had a moderate aquifer-scale proportion of 6.9 percent of the primary aquifer system (table 8A). Among the salinity indicators, TDS was detected at moderate relative-concentrations in 6.9 percent of the primary aquifer system (table 7). Chloride had a concentration greater than its benchmark in at least one site in the CDPH water-quality database prior to June 1, 2006 (table 4), but not during the current period of study.

Understanding Assessment for Iron and Manganese

Potential natural sources of iron and manganese in groundwater include the dissolution of igneous and metamorphic rocks as well as the dissolution of various secondary minerals (Hem, 1989). Some rocks that contain significant amounts of iron and manganese have a relatively high composition of the minerals olivine, biotite, and hornblende. Potential anthropogenic sources of iron and manganese in groundwater include effluents associated with the steel and mining industries (Reimann and de Caritat, 1998) and soil amendments in the form of iron and manganese sulfates that are added to deficient soils to stimulate crop growth. Distributions of iron and manganese concentrations in groundwater are strongly influenced by redox conditions in the aquifer. In sediments, the oxyhydroxides of iron and manganese are common as coatings on mineral surfaces and as suspended particles (Sparks, 1995). Under anoxic conditions, however, the process of reductive dissolution can release these elements from mineral surfaces, and iron and manganese may remain in solution in aquifer systems (Sparks, 1995).

In the NOCO study unit, iron was detected at high relative-concentrations in 25.9 percent of the primary aquifer system and moderate relative-concentrations in 5.2 percent (table 7). Manganese was detected at high relative-concentrations in 39.7 percent and at moderate relative-concentrations in 8.6 percent of the primary aquifer system. High relative-concentrations of iron and manganese occurred in both study areas (figs. 11, 16, 17; tables C1, C2). Iron and manganese were significantly correlated with each other (Spearman's *rho* test, $p < 0.001$, $rho = 0.74$). Iron and manganese concentrations were significantly higher in pre-modern and mixed-age groundwater than in modern groundwater (table 6). In addition, manganese concentrations had a significant positive correlation with

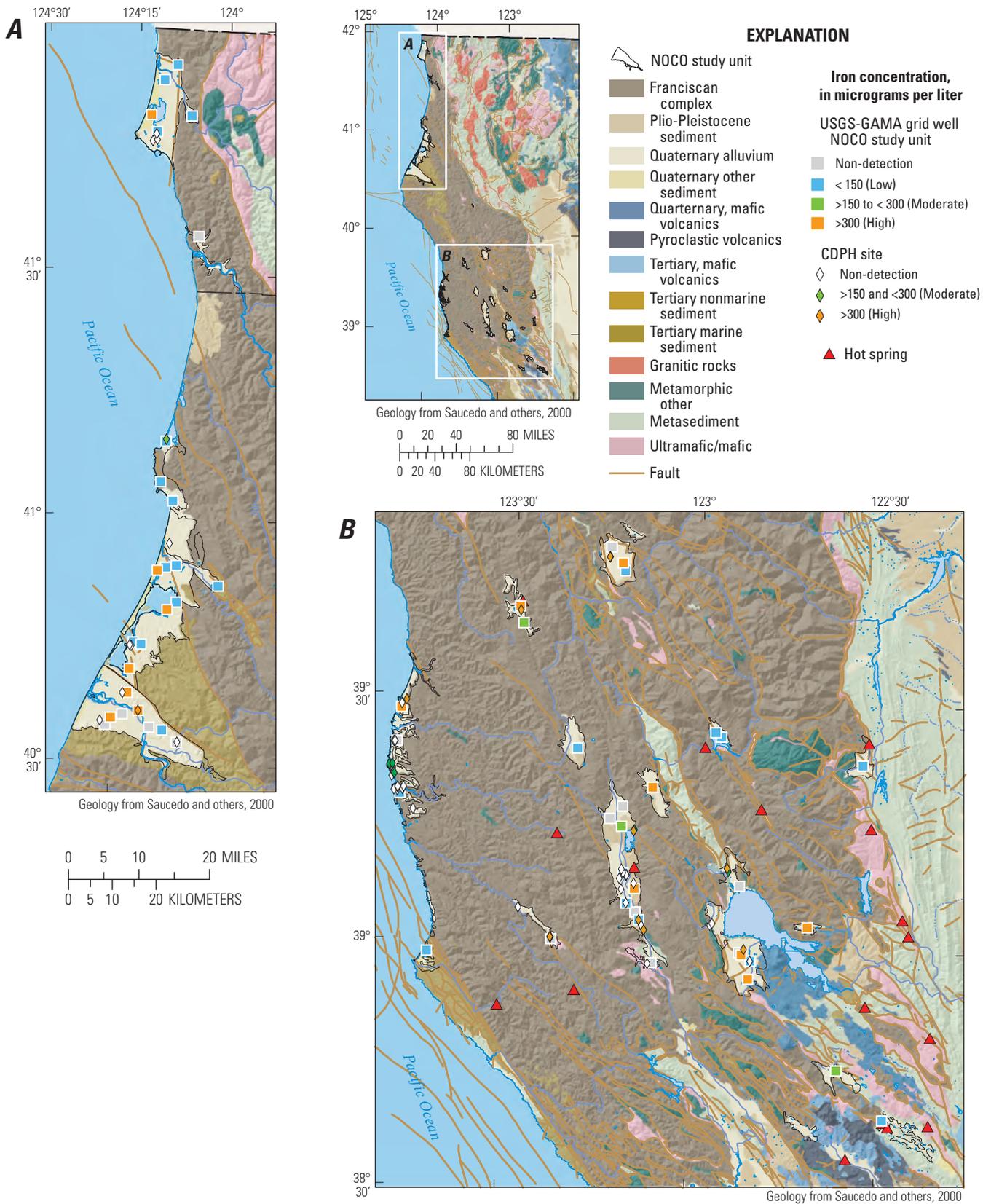


Figure 16. Relative-concentrations of iron in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database in the (A) Coastal Basins study area and (B) Interior Basins study area, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

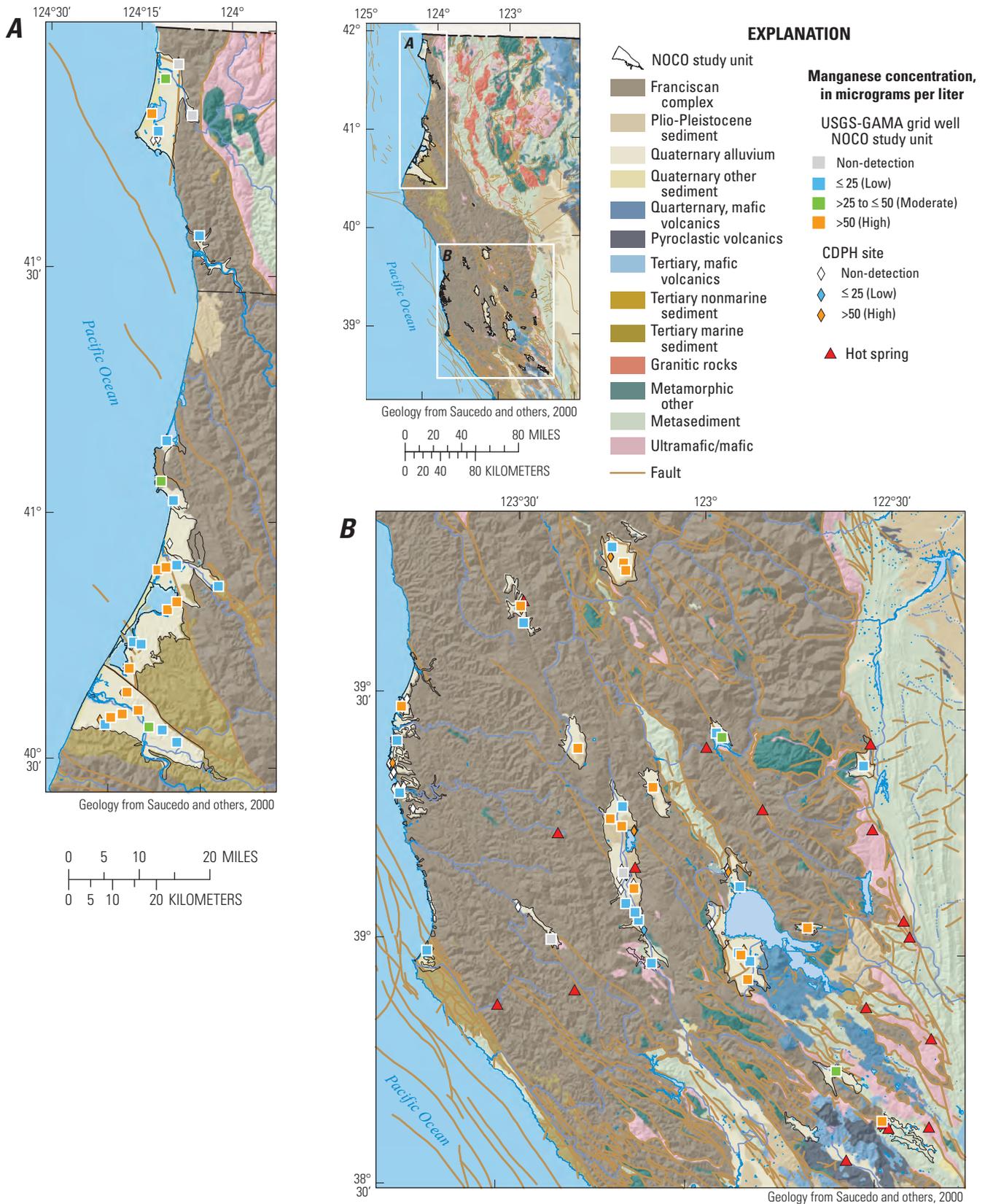


Figure 17. Relative-concentrations of manganese in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database in the (A) Coastal Basins study area and (B) Interior Basins study area, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

depth to top-of-perforations (table 9). Iron and manganese concentrations were negatively correlated with DO (table 9), with significantly higher iron and manganese concentrations in groundwater with anoxic compared with oxic conditions (table 6). This suggests reductive dissolution from aquifer sediments as the primary process for mobilizing iron and manganese into the groundwater.

Organic Constituents with Health-Based Benchmarks

The organic constituents with health-based benchmarks assessed in this study are volatile organic compounds (VOCs) and pesticides and pesticide degradates. VOCs may be present in paints, solvents, fuels, refrigerants, and fumigants, and may be formed as byproducts of water disinfection.

VOCs are characterized by a volatile nature, or tendency to evaporate, and they generally persist longer in groundwater than in surface water because groundwater is isolated from the atmosphere. Pesticides and pesticide degradates are used to control weeds, fungi, or insects in agricultural, urban, and suburban settings. Of the 168 organic constituents analyzed, 11 were detected at 1 or more USGS-GAMA grid sites. Of these 11 constituents, 10 have health-based benchmarks (table 2).

Four organic constituents with health-based benchmarks were identified for additional evaluation in the status assessment because they were detected at moderate or high relative-concentrations or detected in 10 percent or more of the USGS-GAMA grid sites: the VOCs chloroform, vinyl chloride, and methyl *tert*-butyl ether (MTBE) and the pesticide simazine (figs. 10, 18, 19; tables 3, 7).

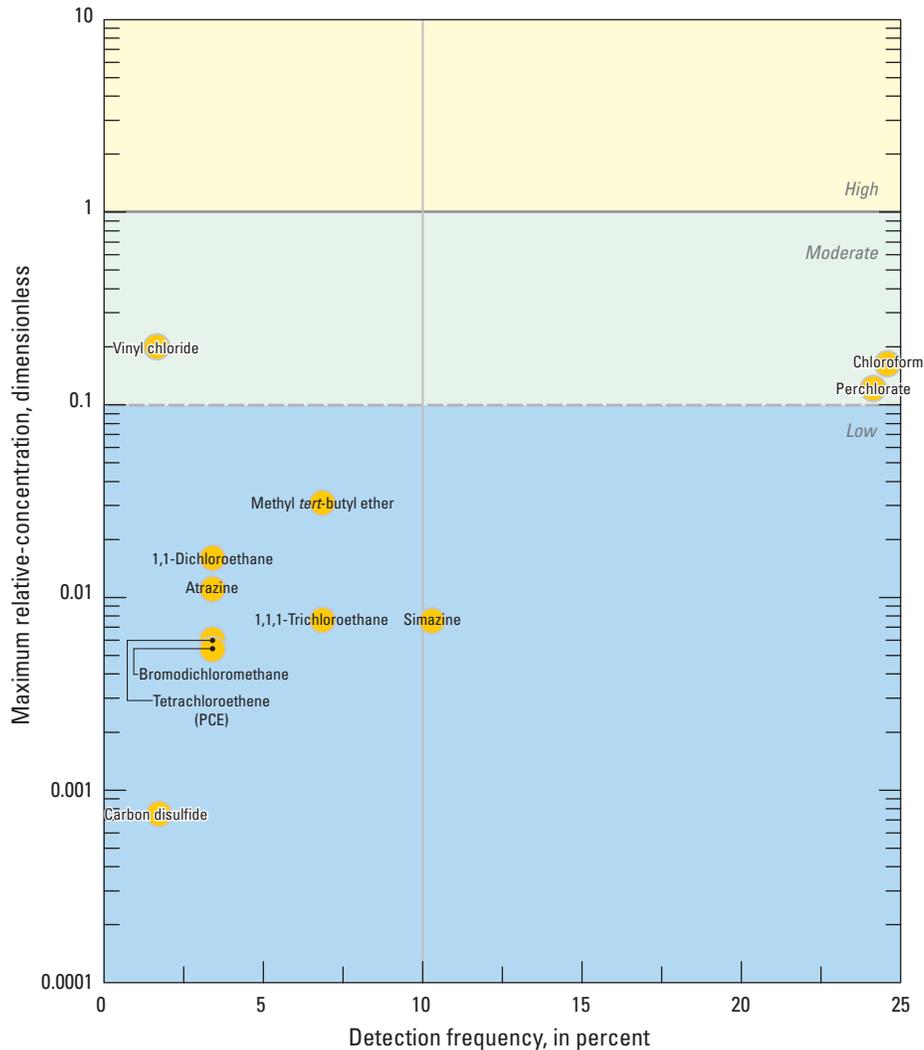


Figure 18. Maximum relative-concentration of organic and special-interest constituents detected in USGS-GAMA grid sites, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area]

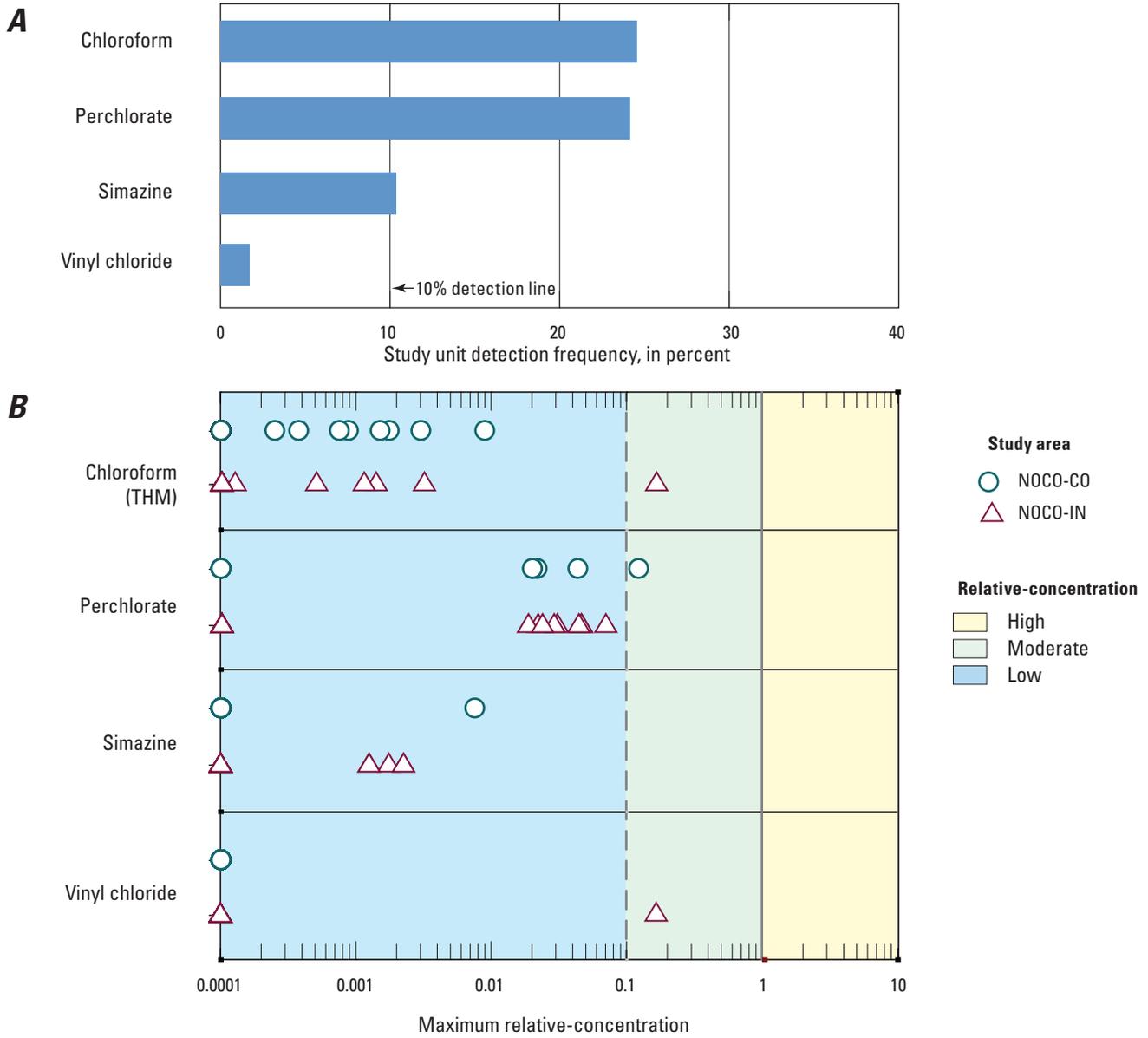


Figure 19. (A) Detection frequency and (B) maximum relative-concentration of organic and special-interest constituents detected in USGS-GAMA grid sites, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area]

Organic constituents with health-based benchmarks, as a group, were detected at high relative-concentrations in 0.2 percent (spatially weighted) of the primary aquifer system, at moderate relative-concentrations in 1.7 percent, and at low relative-concentrations in 39.7 percent. Organic constituents were not detected in 58.4 percent of the primary aquifer system ([table 8B](#)).

Volative Organic Compounds

VOCs, as a class, were detected at high relative-concentrations in 0.2 percent (spatially weighted) of the primary aquifer system, at moderate relative-concentrations in 1.7 percent, and at low relative-concentrations in 36.2 percent. VOCs were not detected in 61.9 percent of the primary aquifer system ([table 8B](#)). The gasoline component MTBE was detected at high relative-concentrations at one CDPH site during the current period of study. The trihalomethane chloroform and the organic synthesis reagent vinyl chloride were detected at moderate relative-concentrations in one grid site in the NOCO-IN study area ([figs. 10, 19B; table C1](#)).

Solvents are used for a variety of industrial, commercial, and domestic purposes (Zogorski and others, 2006). Historically, chlorinated solvents have had a wide variety of applications, including dry cleaning, vapor degreasing of metal parts, hand cleaning of metal parts, paint stripping, fire extinguishers, adhesives, lubricants, and silicones (Petrisor and Wells, 2008). No solvents were detected at high or moderate relative-concentrations in the primary aquifer system ([table 8B](#)). Solvents were detected at low relative-concentrations in 8.6 percent of the primary aquifer system. Solvents were not detected in 91.4 percent of the primary aquifer system.

Organic synthesis reagents are compounds used in the manufacture or preparation of organic compounds. Organic compounds (such as vinyl chloride) are commonly used to make polyvinyl chloride (PVC) pipes. PVC pipes are regularly used in drinking-water systems, and organic synthesis reagents

are commonly released from the PVC pipes into the water system (U.S. Environmental Protection Agency, 2013). No organic synthesis reagents were detected at high relative-concentrations in the primary aquifer system ([table 8B](#)). One organic synthesis reagent (vinyl chloride) was detected at moderate relative-concentrations in 1.7 percent of the primary aquifer system ([fig. 18](#)). Organic synthesis reagents were not detected in 98.3 percent of the primary aquifer system ([table 8B](#)).

Gasoline components include gasoline hydrocarbons and oxygenates. Gasoline hydrocarbons are straight, branched, or cyclic-structured VOCs containing only carbon and hydrogen atoms and are common ingredients in gasoline and other petroleum product formulations. Gasoline oxygenates are compounds that contain oxygen and are added to gasoline to increase the efficiency of combustion in order to meet the requirements of the 1990 Clean Air Act Amendments (Zogorski and others, 2006). Gasoline components were detected at high relative-concentrations in 0.2 percent (spatially weighted) of the primary aquifer system ([table 8B](#)). The gasoline component detected at high relative-concentration was the gasoline oxygenate MTBE. MTBE was also detected at low relative-concentrations in 6.9 percent of the primary aquifer system ([fig. 18](#)). Gasoline components were not detected in 92.9 percent of the primary aquifer system ([table 8B](#)).

Water used for drinking water and other household uses in domestic and municipal systems commonly is disinfected with chlorine solutions. As a side effect to disinfecting the water, the chlorine may react with natural organic matter to produce trihalomethanes (THMs) and other chlorinated and (or) brominated disinfection byproducts. No THMs were detected at high relative-concentrations in the primary aquifer system of the NOCO study unit. THMs were detected at moderate relative-concentrations in 1.7 percent of the primary aquifer system and at low relative-concentrations in 22.4 percent. THMs were not detected in 75.9 percent of the primary aquifer system ([table 8B](#)).

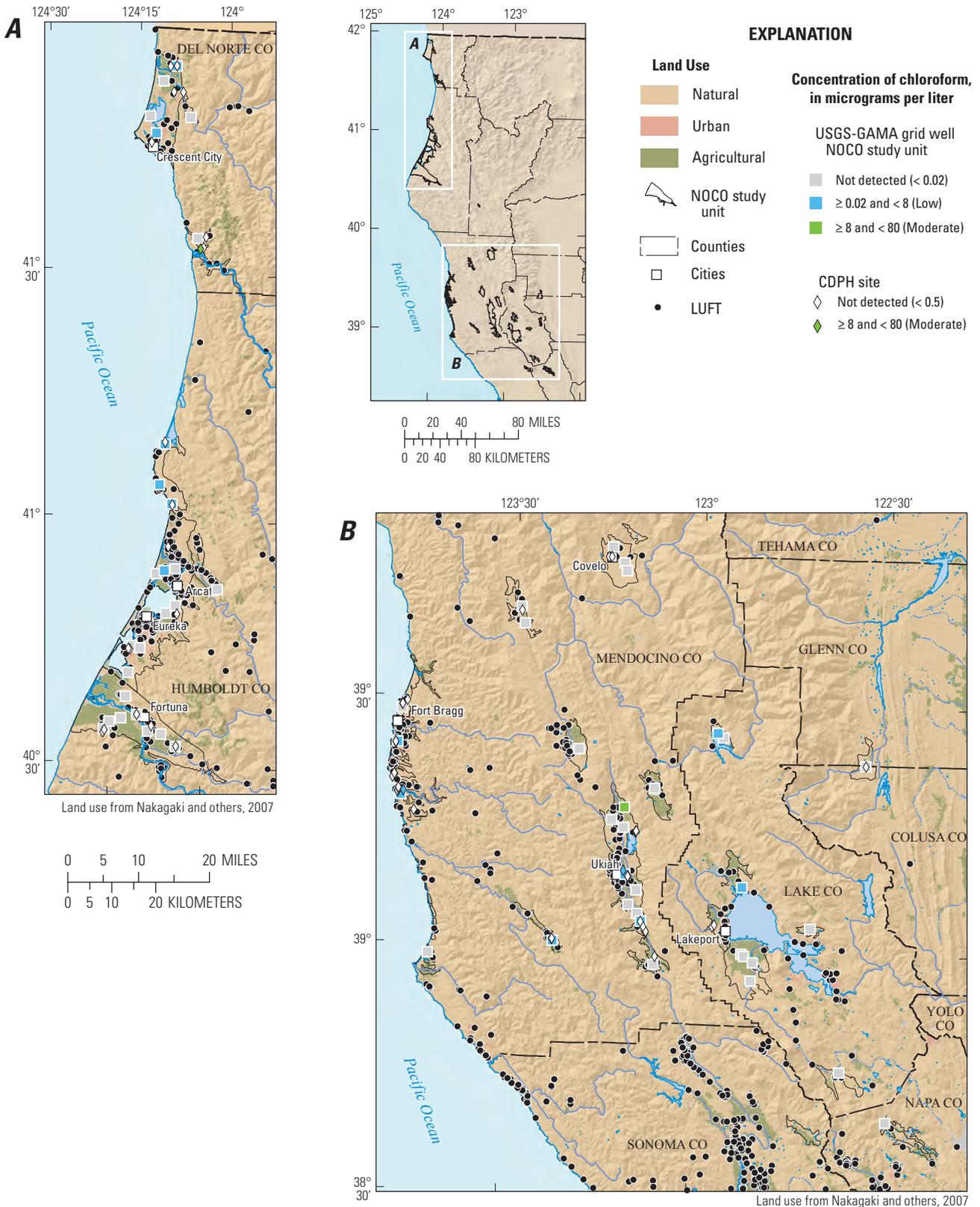


Figure 20. Relative-concentrations of chloroform detected in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database, and locations of leaking or formerly leaking underground fuel tanks (LUFTs), Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area]

Understanding Assessment for Chloroform

Chloroform has been widely detected in national, regional, and local studies of VOCs in ground, surface, source, and drinking waters (Ivahnenko and Barbash, 2004). Additionally, chloroform was the most frequently detected VOC in groundwater nationwide in studies conducted by the USGS NAWQA Program (Zogorski and others, 2006). Chloroform was detected in the NOCO study unit at a high relative-concentration in the CDPH water-quality database prior to June 1, 2006, but was not detected at high relative-concentrations during the current period of study (fig. 20). Chloroform was detected in 24 percent of the primary aquifer system (figs. 10, 18, 19).

Chloroform concentrations were significantly negatively correlated with well depth and depth to top-of-perforations (table 9). Seven of the 14 sites in which chloroform was detected had depths <70 ft bls. Thirteen of these sites had

depth to top-of-perforations information available, and all had depths to top-of-perforations ≤70 ft bls (fig. 21). Chloroform concentrations were positively correlated with DO (table 9) and were significantly higher in groundwater with oxic compared with anoxic conditions (table 6). Thirteen of the 14 sites in which chloroform was detected had oxic groundwater conditions. Chloroform concentrations also were significantly greater in sites with groundwater classified as modern age than in sites with mixed ages (table 6). This result was expected because chloroform concentrations and DO also were significantly correlated. Chloroform concentrations had a negative correlation with pH (table 9), and this relation is likely caused by the positive correlations between pH and well depth and depth to top-of-perforations (table 5). In summary, most sites containing chloroform have depth to top-of-perforations ≤70 ft bls with oxic, modern-age groundwater.

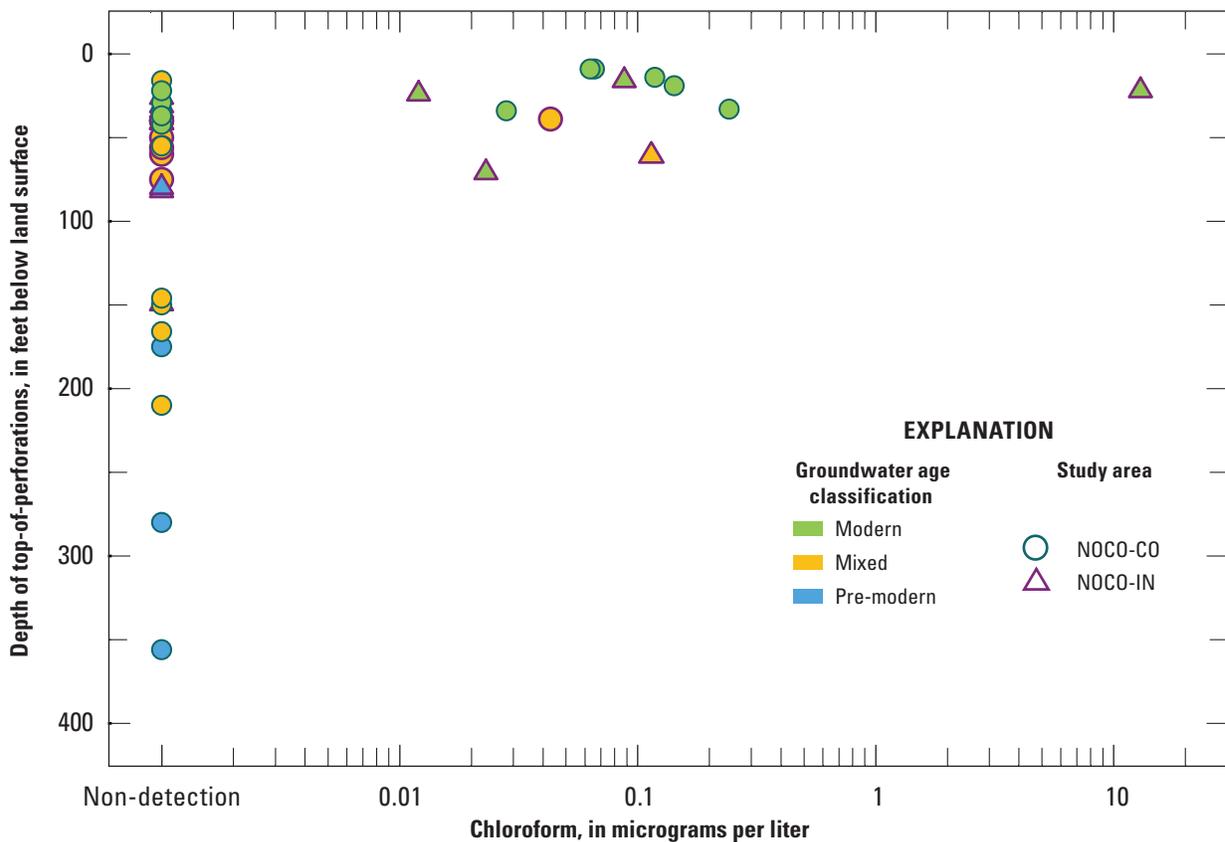


Figure 21. Trihalomethane chloroform concentrations relative to groundwater age classification and depth to top-of-perforations in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

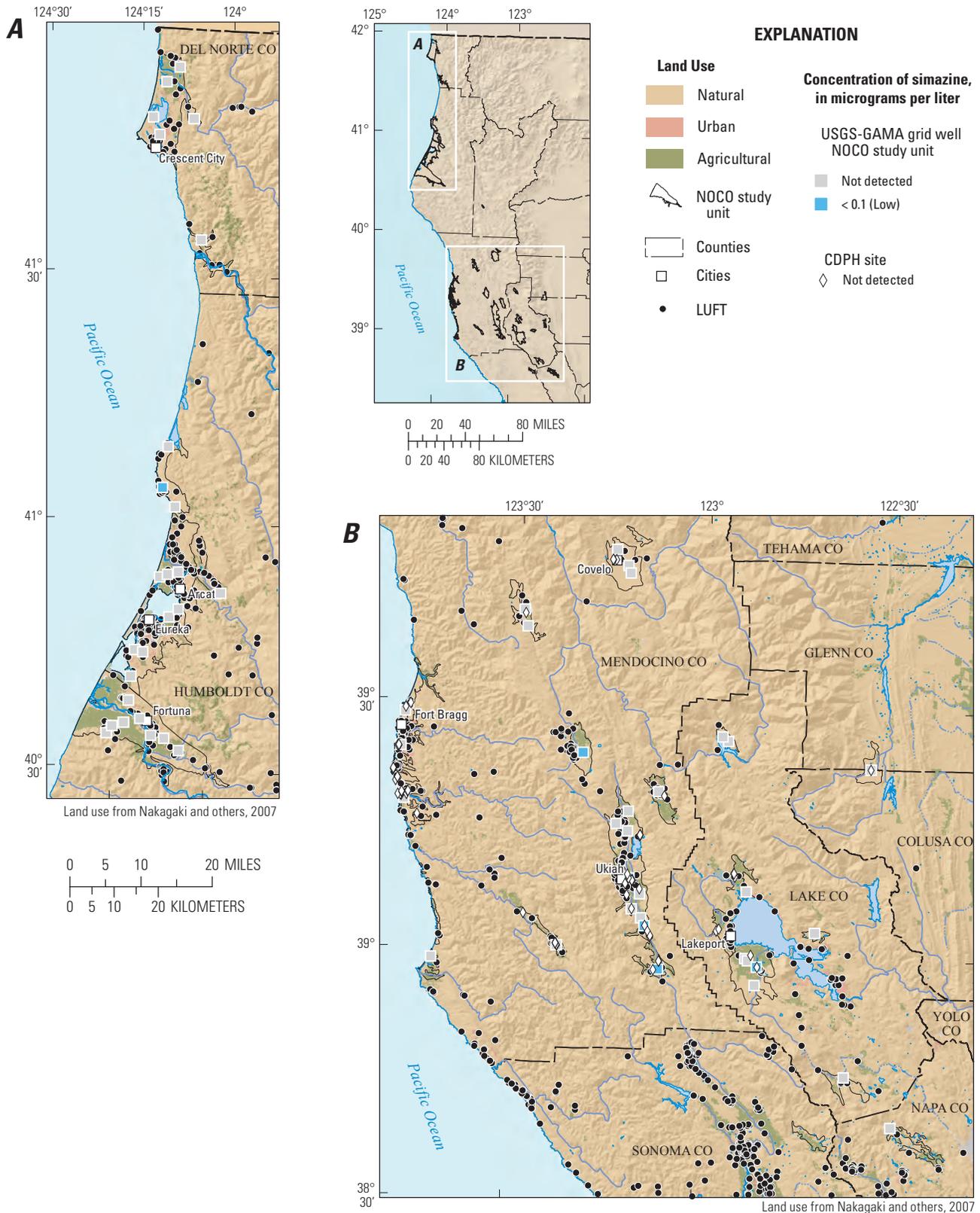


Figure 22. Relative-concentrations of simazine detected in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database, and locations of leaking or formerly leaking underground fuel tanks (LUFTs), Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project. [NOCO-IN, Interior Basins study area; NOCO-CO, Coastal Basins study area]

Pesticides and Pesticide Degradates

Pesticides and pesticide degradates include herbicides, insecticides, and fungicides and are used in agricultural applications on orchards and vineyards and in urban settings for weed control (Gilliom and others, 2006). Pesticides and pesticide degradates, as a class, were detected at low relative-concentrations in 10.3 percent of the primary aquifer system and not detected in 89.7 percent of the primary aquifer system (table 8B). No pesticides or pesticide degradates were detected at high or moderate relative-concentrations in the primary aquifer system.

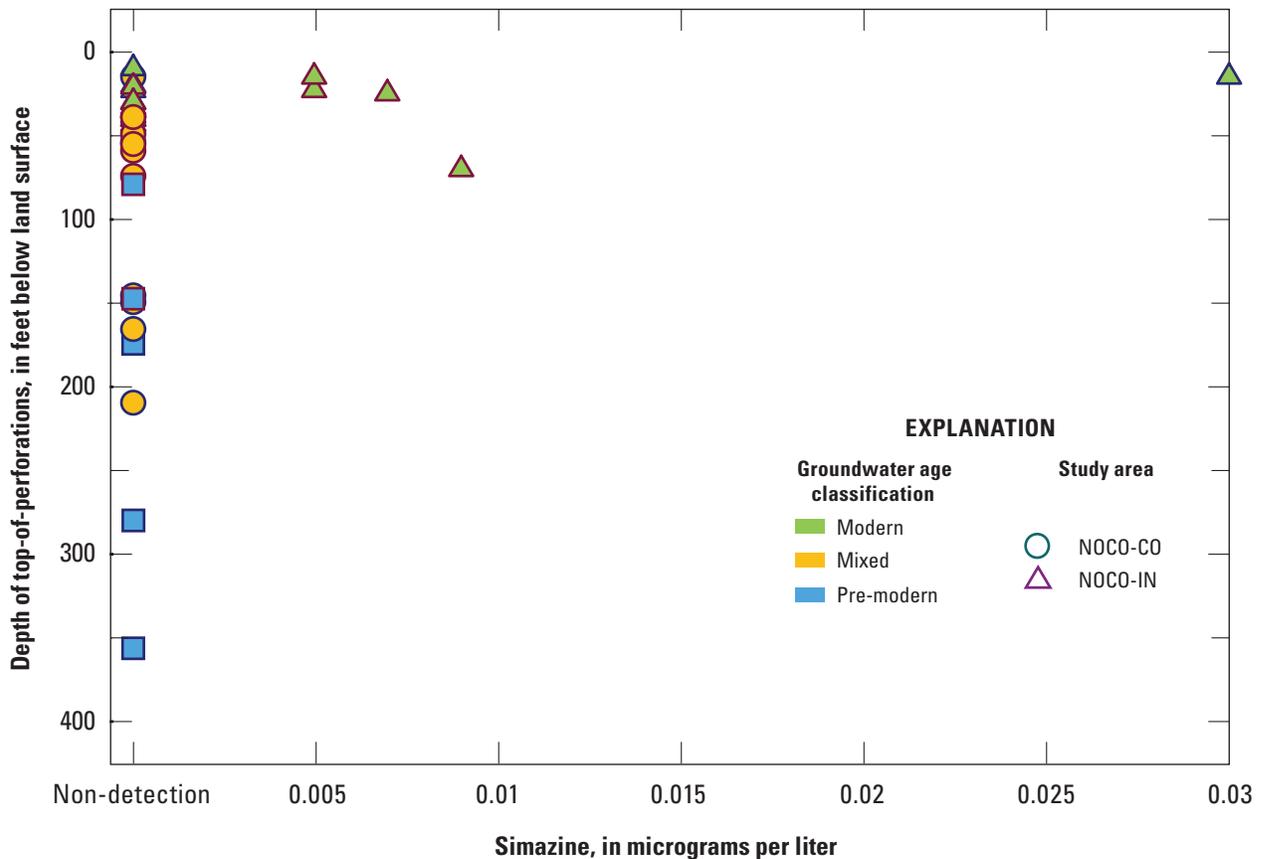
Understanding Assessment for Simazine

Simazine was the most frequently detected triazine herbicide in groundwater in California (Troiano and others, 2001). In addition, simazine was among the most commonly detected herbicides in groundwater in major aquifers across the United States (Gilliom and others, 2006). Simazine

was detected in 10.3 percent of the primary aquifer system (figs. 18, 19, 22).

Simazine was detected at six wells with depths and depths to the top-of-perforations of ≤ 70 ft bls (fig. 23; one of these wells was not plotted on figure 23, due to the lack of depth to the top-of-perforation information), and five of these wells also had depths of < 70 ft bls. All six simazine detections occurred in sites with groundwater classified as modern age (fig. 23), so it is not surprising that simazine concentrations were significantly higher at sites with modern groundwater than at sites with mixed-age groundwater (table 6).

Simazine concentrations were greater in groundwater with oxic redox conditions than in groundwater with anoxic conditions (table 6). Five simazine detections were from sites having groundwater classified as oxic, and one detection was from a site with groundwater with mixed redox conditions. In summary, most sites containing simazine may be characterized as shallow (< 70 ft bls), with depth to top-of-perforations ≤ 70 ft bls and oxic, modern-age groundwater.



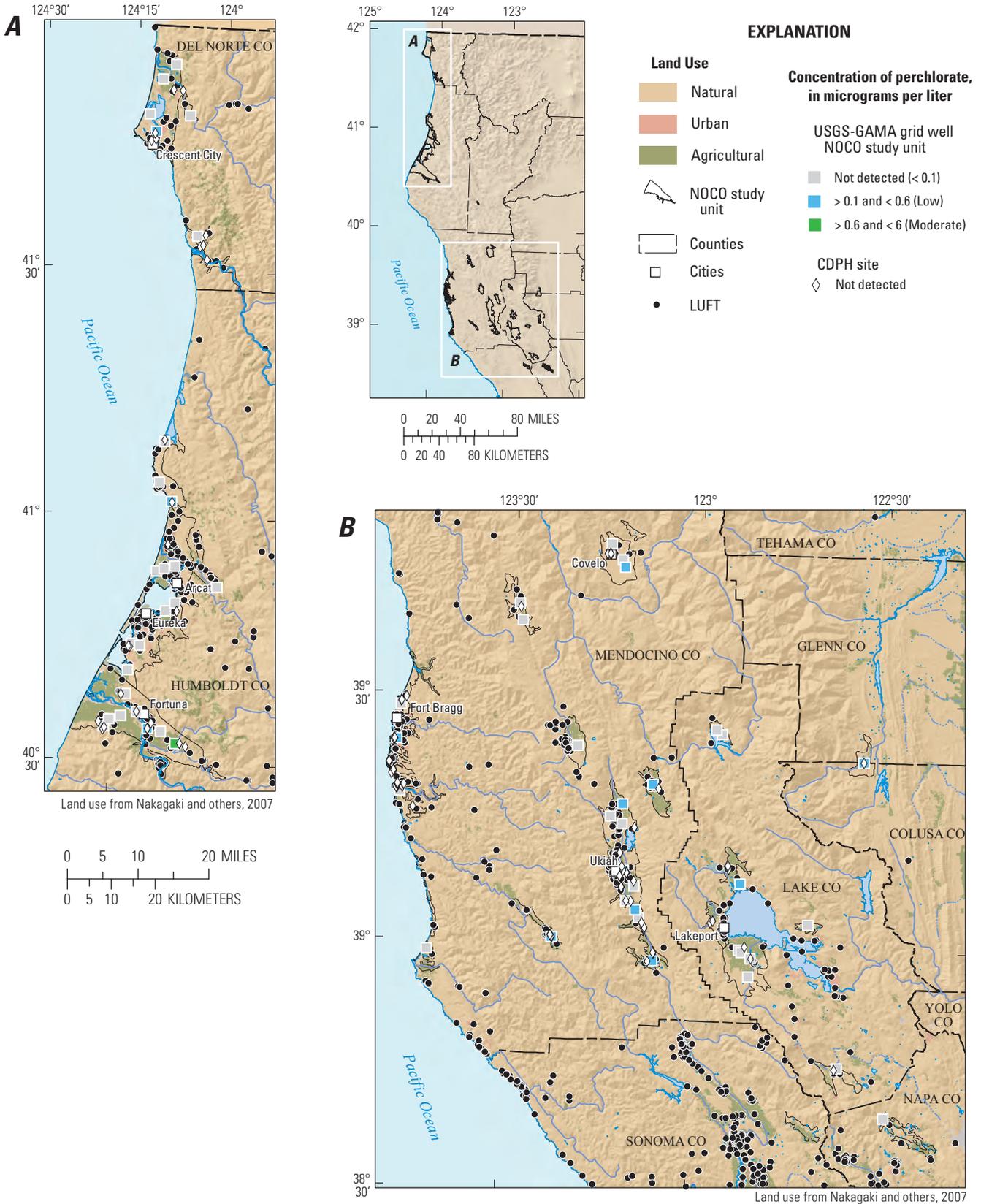


Figure 24. Relative-concentrations of perchlorate in USGS-GAMA grid sites and CDPH sites for the period June 1, 2006–June 1, 2009, from the CDPH water-quality database, and land use, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

Special-Interest Constituent with Health-Based Benchmark

Perchlorate was considered a special-interest constituent at the inception of the GAMA-PBP because it had recently been detected in some drinking-water supplies in California and the CDPH was evaluating whether or not an MCL-CA was warranted (California Department of Public Health, 2008). An MCL-CA of 6 µg/L was established in 2007. Perchlorate was not detected at high relative-concentrations in the primary aquifer system in the NOCO study unit (fig. 24). Perchlorate was detected at moderate relative-concentrations in 1.7 percent of the primary aquifer system and at low relative-concentrations in 22.4 percent (table 8B). Perchlorate was not detected in 75.9 percent of the primary aquifer system.

Understanding Assessment for Perchlorate

Potential sources of perchlorate to groundwater can be natural or anthropogenic. Perchlorate is formed naturally in the atmosphere, and very low concentrations are found in precipitation (Dasgupta and others, 2005; Parker and others, 2008; Rajagopalan and others, 2009). The distribution of perchlorate under natural conditions in groundwater is likely correlated with climate because the extent of evaporative

concentration of precipitation in the hydrologic cycle is likely to increase with increasing aridity (Fram and Belitz, 2011). Perchlorate salts accumulate in unsaturated zones and soils (Rao and others, 2007). Most of the known sites of anthropogenic contamination of groundwater with perchlorate are associated with facilities that manufacture or use solid rocket fuels; other anthropogenic sources include military munitions storage, use, and disposal, and manufacturing and disposal of pyrotechnics, safety flares, and explosives (Parker and others, 2008). Potential agricultural sources include the Chilean nitrate fertilizer applied extensively in the early to mid-1900s (Dasgupta and others, 2005). Irrigation can also be considered an agricultural source if it redistributes perchlorate salts that were originally naturally present in the unsaturated zone (Fram and Belitz, 2011).

Perchlorate concentrations were significantly positively correlated with percentage of agricultural land use (table 9). Agricultural land use (within 500-m-radius buffer areas around each of the sites) was the dominant land-use classification for 8 of the 14 sites in which perchlorate was detected.

Perchlorate concentrations had a significant negative correlation with depth to top-of-perforations (table 9). All 14 wells in which perchlorate was detected had depths to the top-of-perforations of ≤70 ft bls (fig. 25), and 7 of these wells had well depths of <70 ft bls.

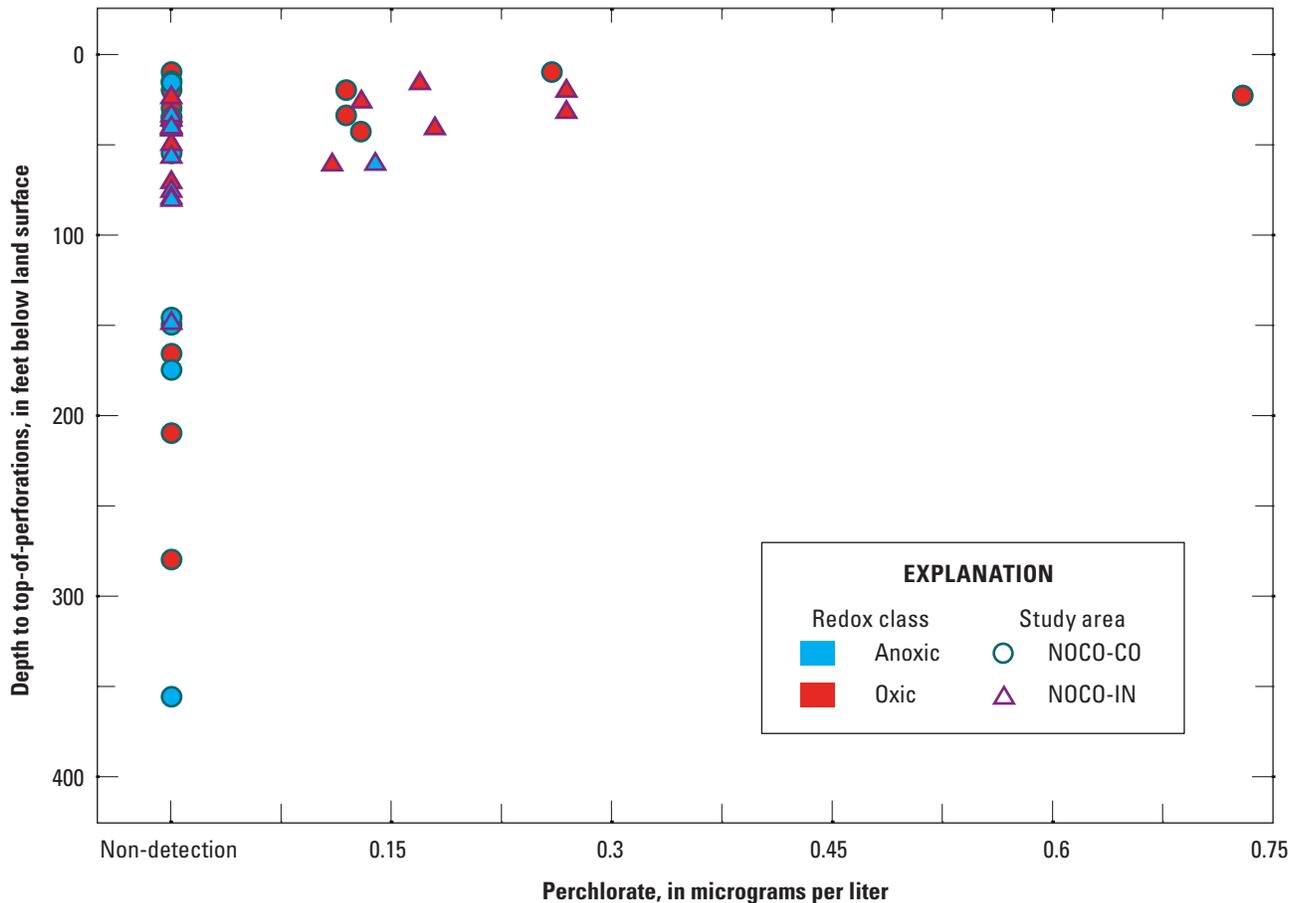


Figure 25. The special-interest constituent perchlorate concentrations relative to geochemical condition and depths to top-of-perforations in USGS-GAMA grid sites sampled for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

Perchlorate concentrations were positively correlated with DO (table 9). Perchlorate concentrations were significantly higher in sites with oxic groundwater than in anoxic groundwater (table 6; fig. 25). While perchlorate biodegrades under anoxic conditions in some aquifers (Sturchio and others, 2007), the apparent relation between perchlorate and DO in the NOCO study unit may be a result of its occurrence predominantly in relatively shallow groundwater.

Fram and Belitz (2011) developed a model for predicting the probability of detecting perchlorate in groundwater under natural conditions as a function of climate. Although perchlorate was often detected in sites having land use classified as agricultural, the observed detection frequencies of perchlorate in the NOCO-IN and NOCO-CO study areas are close to the predicted climate-based detection probabilities, thus, the occurrence of perchlorate in the NOCO study unit is consistent with natural sources of perchlorate.

Summary

Groundwater quality in the approximately 633-square-mile (1,639-square-kilometer) Northern Coast Ranges (NOCO) study unit was investigated as part of the Priority Basin Project of the Groundwater Ambient Monitoring and Assessment (GAMA) Program and the U.S. Geological Survey (USGS) National Water-Quality Assessment (NAWQA) Program. The GAMA Northern Coast Ranges study provides a spatially unbiased characterization of untreated groundwater quality in the primary aquifer system. The assessment is based on water-quality and ancillary data collected in 2009 by the USGS from 58 sites and on data from the California Department of Public Health (CDPH) water-quality database.

The first component of this study, the status of the current quality of the groundwater resource, was assessed by using data from samples analyzed for naturally occurring inorganic constituents (such as trace elements and major and minor ions), organic constituents (volatile organic compounds [VOCs] and pesticides and pesticide degradates), and the special-interest constituent perchlorate. The status assessment characterizes the quality of groundwater resources in the primary aquifer system of the NOCO study unit, not the treated drinking water delivered to consumers by water purveyors.

Relative-concentrations (sample concentration divided by the health- or aesthetic-based benchmark concentration) were used for evaluating groundwater quality for those constituents that have Federal and (or) California regulatory or nonregulatory benchmarks for drinking-water quality.

Aquifer-scale proportion was used as the primary metric for evaluating regional-scale groundwater quality. High aquifer-scale proportion was defined as the percentage of the primary aquifer system with relative-concentration greater than (>) 1.0 for a particular constituent or class of constituents;

proportion is based on an aerial rather than a volumetric basis. Moderate and low aquifer-scale proportions were defined as the percentage of the primary aquifer system with moderate and low relative-concentrations, respectively. Two statistical approaches, grid-based and spatially weighted, were used to evaluate aquifer-scale proportions for individual constituents and classes of constituents. Grid-based and spatially weighted estimates were comparable in the NOCO study unit (within 90 percent confidence intervals).

Inorganic constituents with health-based benchmarks were detected at high relative-concentrations in 10.3 percent of the primary aquifer system, moderate relative-concentrations in 13.8 percent, and low relative-concentrations in 75.9 percent. The high aquifer-scale proportion of inorganic constituents primarily reflected high aquifer-scale proportions of boron (8.6 percent), arsenic (3.4 percent), and barium (1.7 percent). The inorganic constituents with aesthetic-based benchmarks—iron and manganese—were detected at high relative-concentrations in 25.9 percent and 39.7 percent of the primary aquifer system, respectively.

Relative-concentrations of organic constituents with health-based benchmarks (one or more) were high in 0.2 percent of the primary aquifer system, moderate in 1.7 percent, and low in 39.7 percent. Organic constituents were not detected in 58.4 percent of the primary aquifer system. Of the 168 organic constituents analyzed, 11 constituents were detected. Nearly all detections were at low relative-concentrations. The VOC chloroform and the herbicide simazine were the only organic constituents detected in more than 10 percent of grid sites. The special-interest constituent perchlorate was detected at moderate relative-concentrations in 1.7 percent of the primary aquifer system and at low relative-concentrations in 22.4 percent.

The second component of this work, the understanding assessment, used statistical correlations between concentrations of constituents and values of selected potential explanatory factors to identify the factors potentially affecting the concentrations and occurrences of constituents detected at high relative-concentrations or, for organic constituents, with detection frequencies >10 percent. The potential explanatory factors evaluated were land use, density of septic tanks, density of leaking (or formerly leaking) underground fuel tanks, well depth and depth to the top of the perforated interval in the well, aridity index, field water temperature, distance to nearest hot spring and geothermal well, pH, dissolved oxygen concentration, study (geographic) area, groundwater age distribution, and redox condition.

The understanding assessment indicated that high and moderate relative-concentrations of boron primarily occurred in the Interior Basins study area and may be attributed to groundwater interacting with hydrothermal systems. High and moderate relative-concentrations of boron were associated with elevated groundwater temperatures, groundwater chemistry characteristics similar to those of geothermal waters, and proximity to known geothermal areas. Boron concentrations generally were higher where low dissolved oxygen concentrations or anoxic conditions

exist. The understanding assessment also indicated that high and moderate relative-concentrations of arsenic occur predominantly in the Interior Basins study area under reducing conditions and possibly are enhanced by hydrothermal systems (when present).

Chloroform, simazine, and perchlorate were detected predominantly in shallow sites with depths to top-of-perforations less than or equal to 70 feet below land surface, modern groundwater, and oxic groundwater conditions.

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Appendix A. Map of the Northern Coast Ranges (NOCO) Study Unit

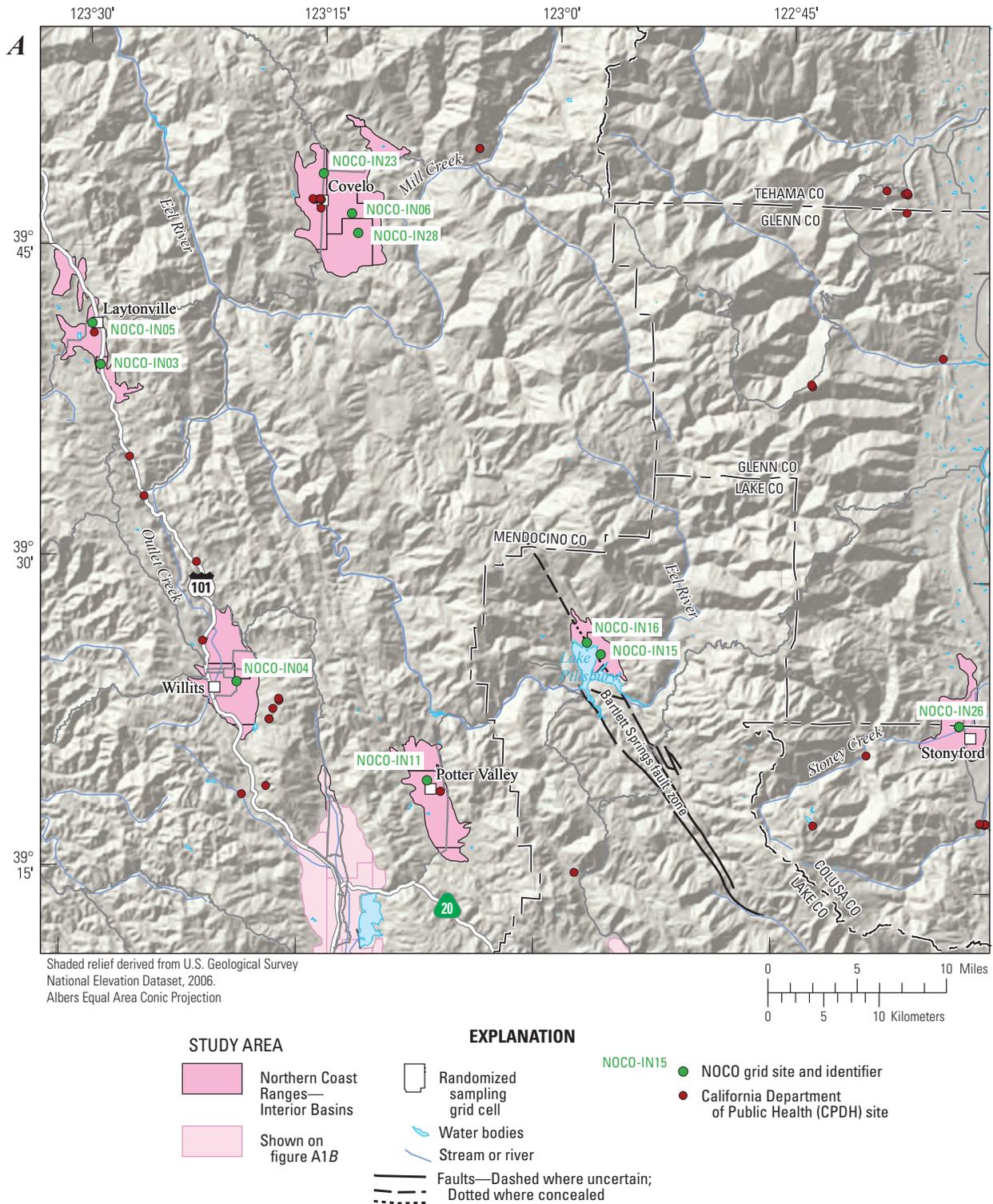
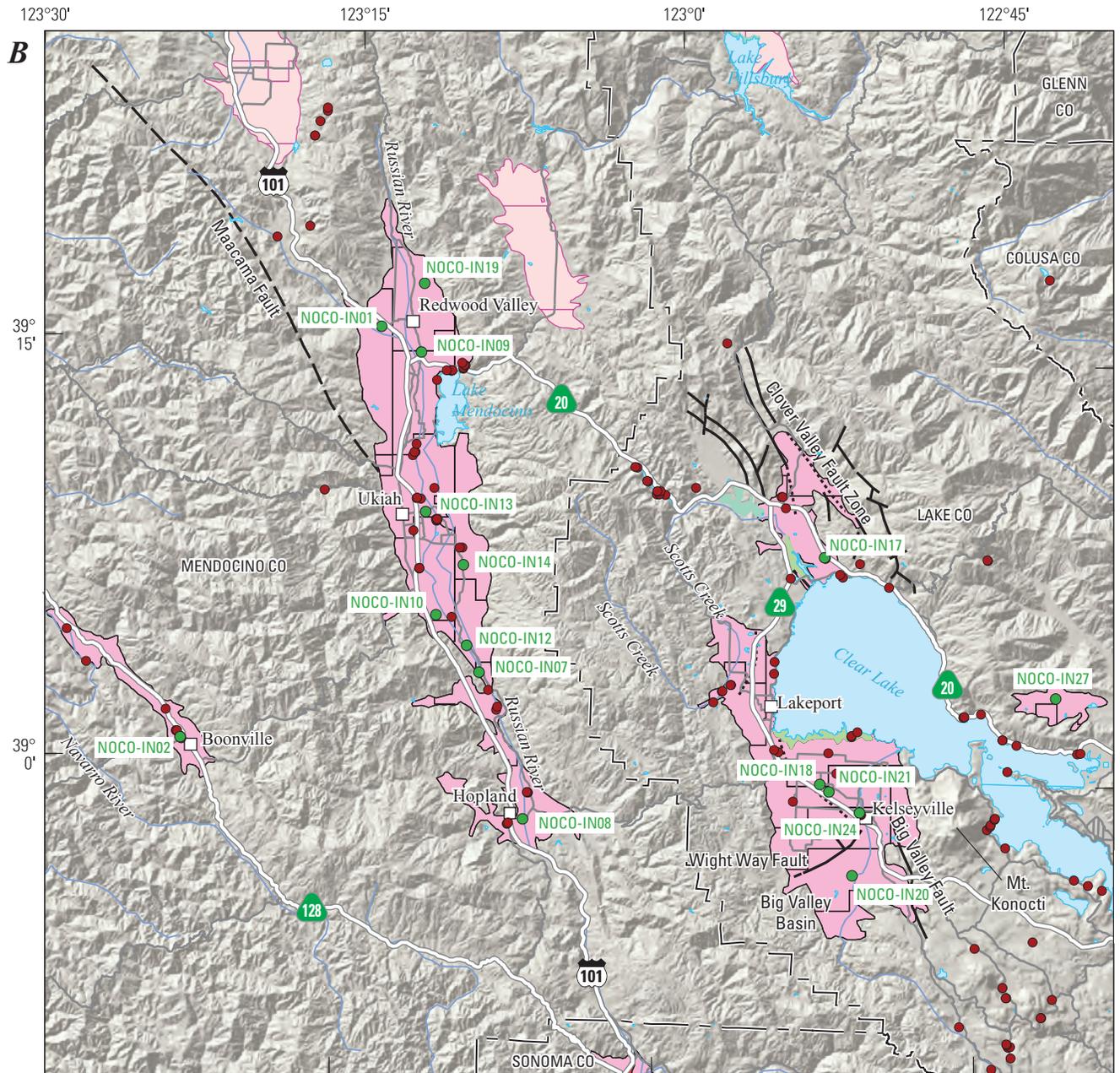
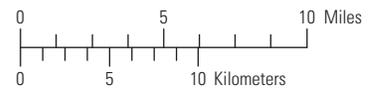


Figure A1. Northern Coast Ranges (NOCO) Groundwater Ambient Monitoring and Assessment (GAMA) study unit showing the location of (A) the northern part, (B) the central part, and (C) the southern part of the Interior Basins (NOCO-IN) study area and (D) the northern part, (E) the central part, (F) the southern part of the Coastal Basins (NOCO-CO) study area, the boundaries of the California Department of Water Resources (CDWR) defined groundwater basins, the distribution of the study area grid cells, the location of California Department of Public Health (CDPH) sites, and the location of sampled U.S. Geological Survey (USGS) grid sites, major cities, major roads, topographic features, and hydrologic features.



Shaded relief derived from U.S. Geological Survey National Elevation Dataset, 2006. Albers Equal Area Conic Projection



EXPLANATION

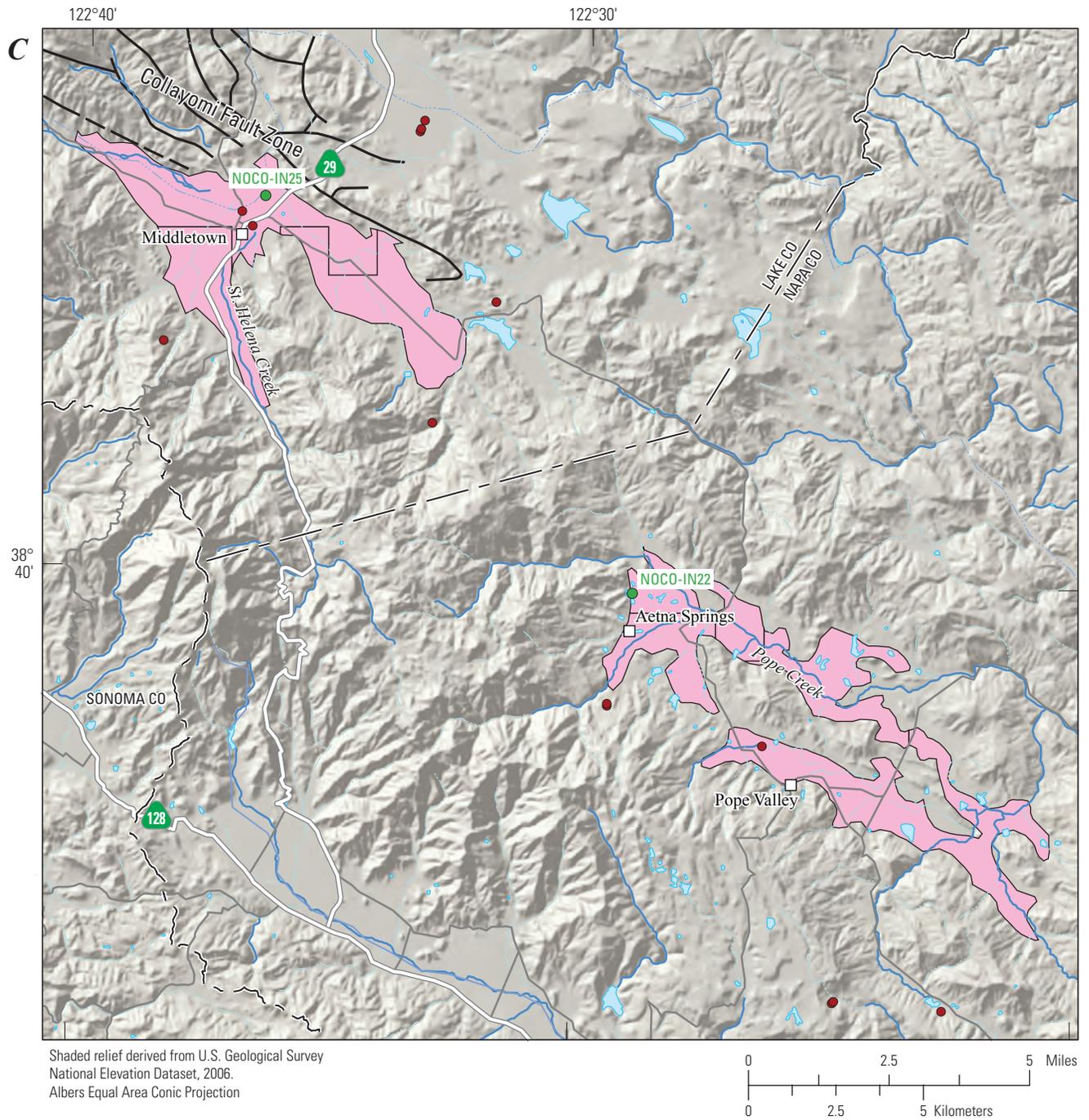
STUDY AREA

- Northern Coast Ranges—Interior Basins
- Shown on figure A1A

- Randomized sampling grid cell
- Water bodies
- Stream or river
- Faults—Dashed where uncertain;
- Dotted where concealed

- NOCO grid site and identifier
- California Department of Public Health (CPDH) site

Figure A1. —Continued



EXPLANATION

STUDY AREA

 Northern Coast Ranges—Interior Basins

 Randomized sampling grid cell

 **NOCO-IN22** NOCO grid site and identifier

 California Department of Public Health (CPDH) site

 Water bodies

 Stream or river

 Faults—Dashed where uncertain

Figure A1. —Continued

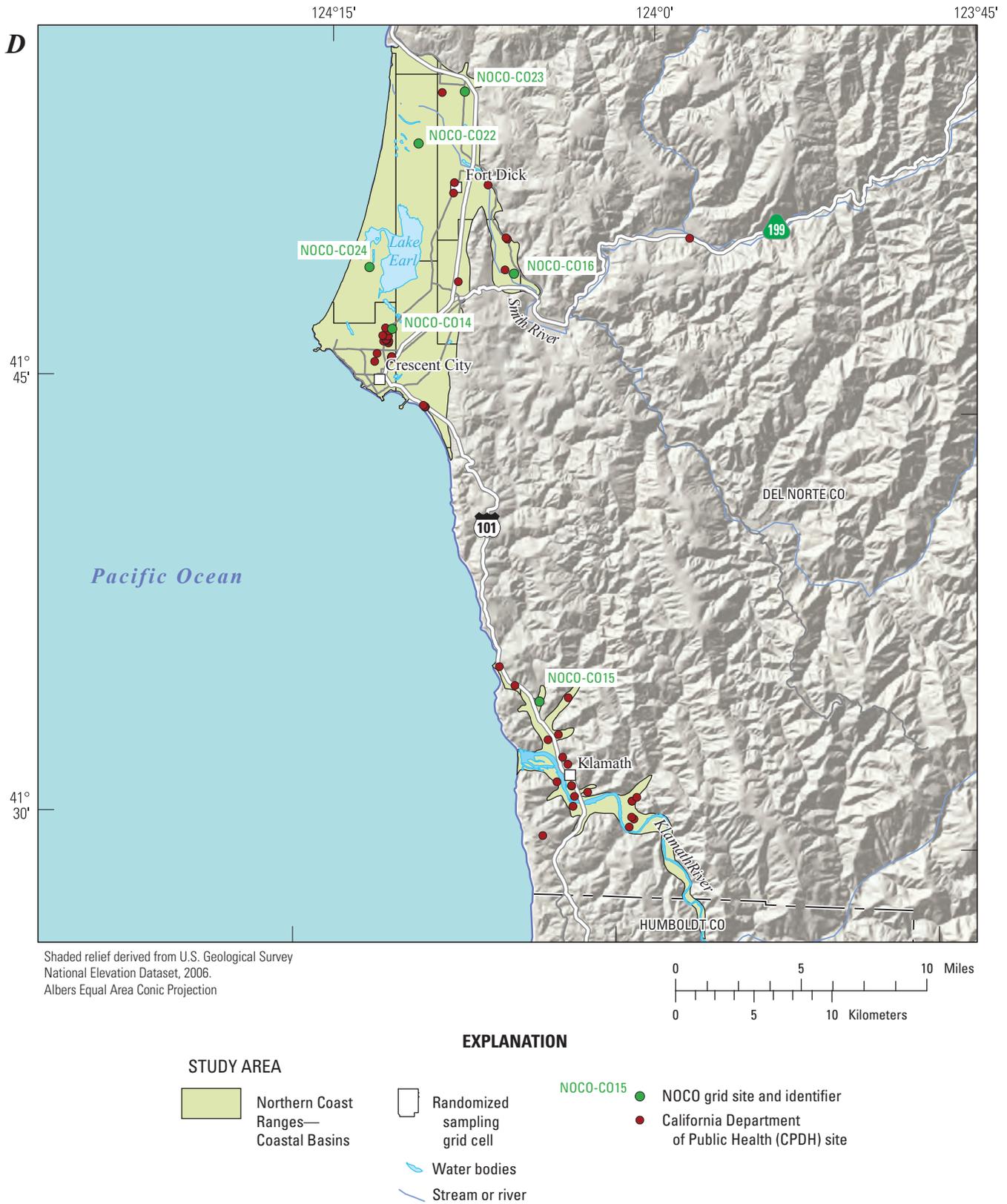
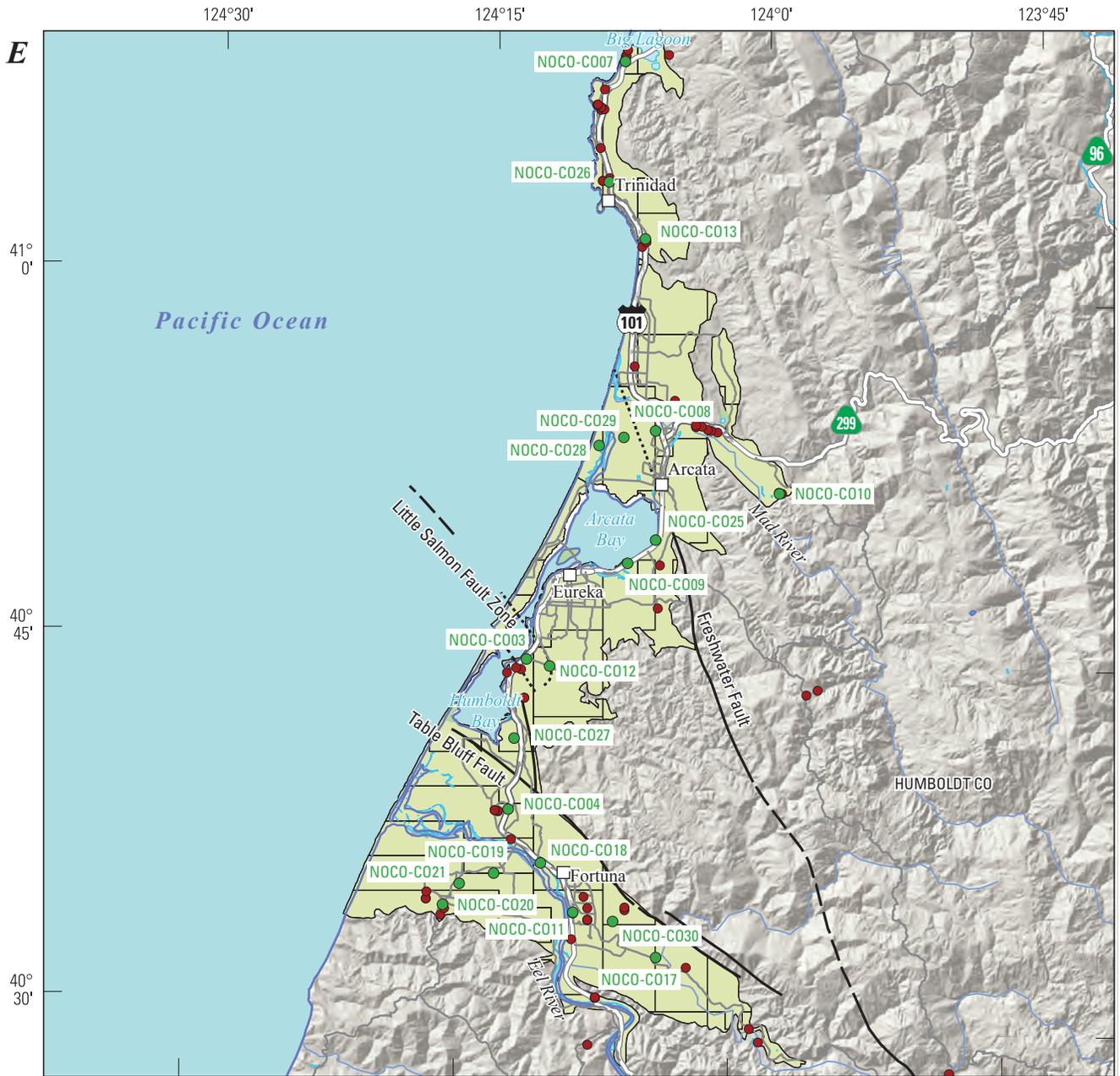
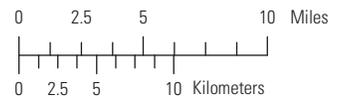


Figure A1. —Continued



Shaded relief derived from U.S. Geological Survey National Elevation Dataset, 2006. Albers Equal Area Conic Projection



STUDY AREA

 Northern Coast Ranges—Coastal Basins

EXPLANATION

 Randomized sampling grid cell

 Water bodies

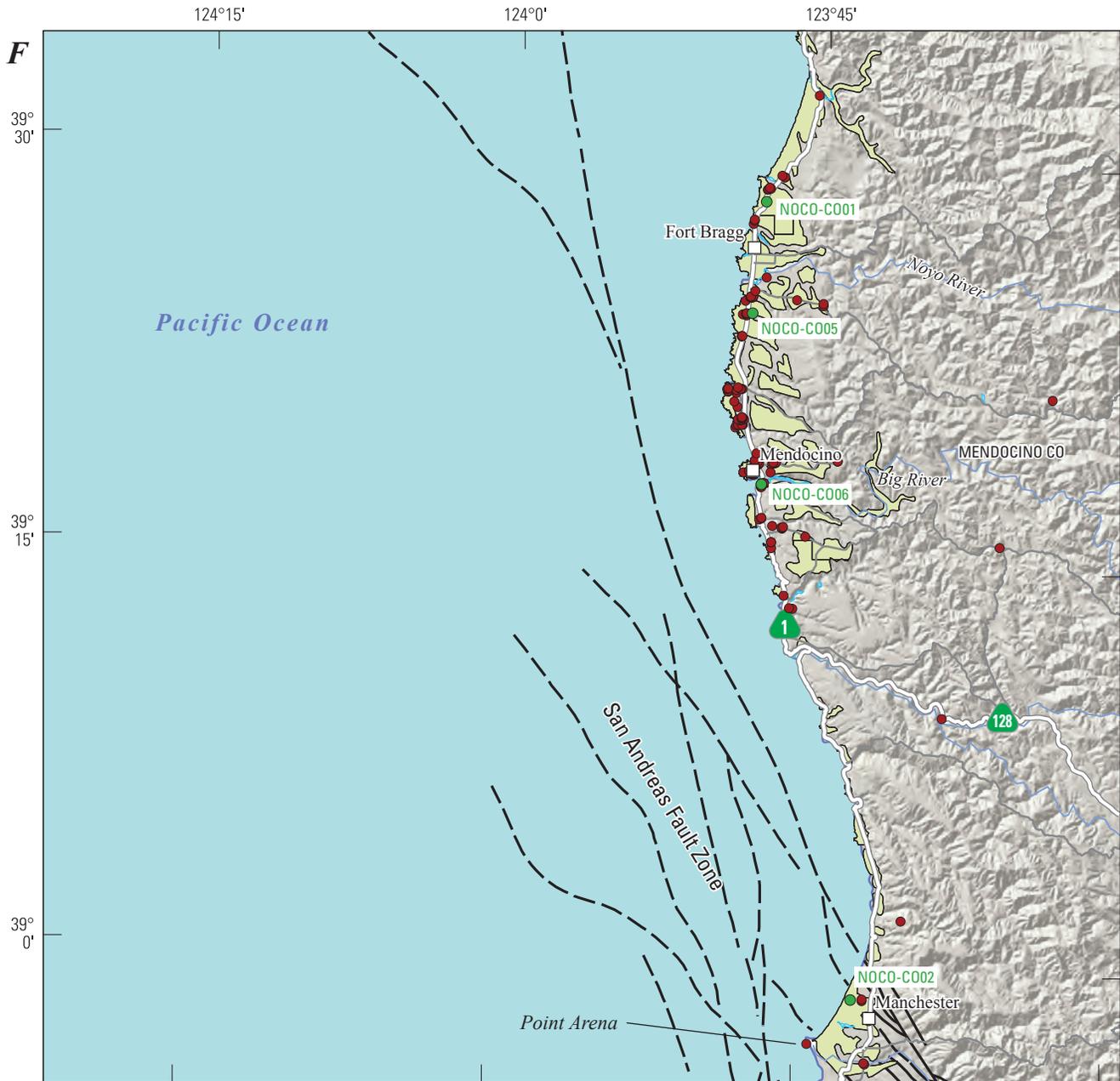
 Stream or river

 Faults—Dashed where uncertain; Dotted where concealed

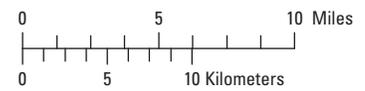
 NOCO-C011 NOCO grid site and identifier

 California Department of Public Health (CPDH) site

Figure A1. —Continued



Shaded relief derived from U.S. Geological Survey National Elevation Dataset, 2006. Albers Equal Area Conic Projection



STUDY AREA

Northern Coast Ranges—Coastal Basins

EXPLANATION

- Randomized sampling grid cell
- Water bodies
- Stream or river
- Faults—Dashed where uncertain
- NOCO-C005 NOCO grid site and identifier
- California Department of Public Health (CPDH) site

Figure A1. —Continued

Appendix B. Ancillary Datasets

Land-Use Classification

Land use was classified by using an enhanced version of the satellite-derived (30-m pixel resolution) nationwide USGS National Land Cover Dataset (NLCD) (Nakagaki and others, 2007). This dataset has been used in previous national and regional studies relating land use to water quality (Gilliom and others, 2006; Zogorski and others, 2006). The dataset characterizes land use during the early 1990s. One pixel in the dataset imagery represents a land area of 9,688 square feet (ft²) (900 square meters [m²]), calculated from the pixel of 98 ft (30 m). The imagery was classified into 25 land-cover classifications (Nakagaki and Wolock, 2005). These 25 land-cover classifications were aggregated into three principal land-use classes—urban, agricultural, and natural.

Percentages of the three land-use classes in the study unit and study areas, in areas within a radius of 500 m (500-m buffers) of each USGS-GAMA Program grid site, and in areas within 500-m buffers of each CDPH site were calculated by using Esri ArcGIS® (Johnson and Belitz, 2009; [tables B1](#), [B2](#)). A 500-m buffer surrounding a site has been shown to be effective at correlating urban land use with VOC occurrence for the purposes of statistical characterization (Johnson and Belitz, 2009).

Septic Tank and Leaking or Formerly Leaking Underground Fuel Tank Density

Septic tanks generally are associated with dispersed residences and may occur in areas classified as natural or agricultural in addition to areas classified as urban. Septic tank density was determined from the 1990 Census of Population and Housing: Summary Tape File 3A dataset (U.S. Department of Commerce, 1990). The density of septic tanks in each housing census block was calculated from the number of tanks and block area. The density of septic tanks around each USGS-GAMA grid site was then calculated from the area-weighted mean of the block densities within a 500-m buffer around the site (Tyler Johnson, U.S. Geological Survey, California Water Science Center, written commun., 2013; [table B2](#)).

LUFTs generally are associated with urban land use, although they may occur in areas classified as natural or agricultural. LUFT density was determined from the locations of tanks in the SWRCB's GeoTracker database of environmental cleanup sites (California State Water Resources Control Board, 2012). The density of LUFTs was calculated by using Theissen polygons (Tyler Johnson, U.S. Geological Survey, California Water Science Center, written commun., 2007). The boundaries of the Theissen polygon around a particular LUFT were created by bisecting the linear distances between the LUFT and all the surrounding LUFTs. The density of LUFTs in the polygon is the number of tanks in the polygon (generally one) divided by the area of the polygon in square kilometers. Each USGS-GAMA grid site was assigned the LUFT density of the Theissen polygon in which it is located ([table B2](#)).

Well Construction Information

Most well-construction data were from drillers' logs filed with the CDWR. For some wells, well construction data were obtained from ancillary records of site owners or the USGS National Water Information System (NWIS) database. Well depths and depths to the top and bottom of perforations for USGS-GAMA grid sites are listed in [table B3](#). The spring was assigned a value of "at LSD" (at land-surface datum) for the well depth and depths to the top and bottom of the perforations, equivalent to 0 ft bls. USGS-GAMA grid site verification procedures are described by Mathany and others (2011). Land-surface altitudes were obtained from USGS digital elevation GIS coverages and are reported in feet relative to the North American Vertical Datum of 1988 (NAVD 88) ([table B3](#)).

Aridity Index

The climate at each USGS-GAMA grid site was represented by an aridity index (United Nations Educational, Scientific, and Cultural Organization, 1979; Grove, 1999; [table B3](#)):

$$\text{aridity index} = \frac{\text{average annual precipitation}}{\text{average annual evapotranspiration}}. \quad (\text{B1})$$

Higher values of the index correspond to wetter conditions. Average annual precipitation for each site was extracted from the PRISM average annual precipitation for 1971–2000 GIS coverage (PRISM Group, Oregon State University, 2012). Average annual evapotranspiration for each site was extracted from an Esri ArcGIS® coverage modified from Flint and Flint (2007). The modification consisted of calibrating the evapotranspiration values to the measured California Irrigation Management Information System reference evapotranspiration values (California Irrigation Management Information System, 2005; Alan Flint, U.S. Geological Survey, California Water Science Center, oral commun., 2009).

Distance to Nearest Hot Spring and Geothermal Well

Distances to nearest hot spring and geothermal well were used as potential direct indicators of groundwater interaction with the localized geothermal systems located within the Geysers–Clear Lake area. Hot spring locations were obtained from the USGS MRDS, which is a collection of reports describing metallic and nonmetallic mineral resources throughout the world. Included in the MRDS are deposit name, location, commodity, deposit description, geologic characteristics, production, reserves, resources, and references (U.S. Geological Survey, 2005). Geothermal well locations were obtained from the California Department of Conservation DOGGR database. DOGGR oversees the drilling, operation, maintenance, and plugging and abandonment of oil, natural gas, and geothermal wells. Each USGS-GAMA grid site was assigned the Euclidean distance, in kilometers, to the nearest hot spring and geothermal well using the "Near" tool in the Esri ArcGIS® software package ([table B3](#)).

Groundwater Age Classification

Groundwater dating techniques indicate the time since the groundwater was recharged into the primary aquifer system. The techniques used in this report to estimate groundwater residence times or ‘age’ were those based on tritium (for example: Tolstikhin and Kamenskiy, 1969; Torgersen and others, 1979) and carbon-14 activities (for example: Vogel and Ehhalt, 1963; Plummer and others, 1993).

Tritium is a short-lived radioactive isotope of hydrogen with a half-life of 12.32 years (Lucas and Unterweger, 2000). Tritium is produced naturally in the atmosphere from the interaction of cosmogenic radiation with nitrogen (Craig and Lal, 1961), by aboveground nuclear explosions, and by the operation of nuclear reactors. Tritium enters the hydrologic cycle following oxidation to tritiated water. Aboveground nuclear explosions resulted in a large increase in tritium values in precipitation, beginning in about 1952 and peaking in 1963 at values over 1,000 TU in the northern hemisphere (Michel, 1989). Radioactive decay over a period of 50 years would decrease tritium values of 10 TU to 0.6 TU. Tritium values in precipitation under natural conditions in the NOCO study unit in 2009 would be about 2 to 5 TU (Robert Michel, U.S. Geological Survey, California Water Science Center, Menlo Park Stable Isotope and Tritium Laboratories, written commun., 2013).

Previous investigations have used a range of tritium values from 0.3 to 1.0 TU as thresholds for indicating presence of water that has exchanged with the atmosphere since 1952 (Michel, 1989; Plummer and others, 1993; Michel and Schroeder, 1994; Clark and Fritz, 1997; Manning and others, 2005). For groundwater samples collected for the NOCO study unit in 2009 (Mathany and others, 2011), tritium values greater than a threshold of 0.5 TU were defined as indicating presence of groundwater recharged since 1952.

Carbon-14 (^{14}C) is a widely used chronometer based on the radiocarbon content of organic and inorganic carbon. Dissolved inorganic carbon species, carbonic acid, bicarbonate, and carbonate typically are used for ^{14}C dating of groundwater. ^{14}C is formed in the atmosphere by the interaction of cosmic-ray neutrons with nitrogen and, to a lesser degree, with oxygen and carbon. ^{14}C is incorporated into carbon dioxide and mixed throughout the atmosphere. The carbon dioxide enters the hydrologic cycle because it dissolves in precipitation and surface water in contact with the atmosphere. ^{14}C activity in groundwater, expressed as percent modern carbon (pmc), reflects the time since groundwater was last exposed to the atmospheric ^{14}C source. ^{14}C has a half-life of 5,730 years and can be used to estimate groundwater ages ranging from 1,000 to approximately 30,000 years before present.

The ^{14}C age (residence time, presented in years) is calculated on the basis of the decrease in ^{14}C activity as a result of radioactive decay since groundwater recharge, relative to an assumed initial ^{14}C concentration (Clark and Fritz, 1997). An average initial ^{14}C activity of 100 pmc is assumed for this study, with estimated errors on calculated groundwater ages up to ± 20 percent. Calculated ^{14}C ages in this study are referred to as “uncorrected” because they have not been adjusted to consider exchanges with sedimentary

sources of carbon (Fontes and Garnier, 1979). Groundwater with a ^{14}C activity of >88 pmc is reported as having an age of $<1,000$ years; no attempt has been made to refine ^{14}C ages $<1,000$ years. Measured values of pmc can be >100 pmc because the definition of the ^{14}C activity in “modern” carbon does not include the excess ^{14}C produced in the atmosphere by aboveground nuclear weapons testing. For the NOCO study unit, ^{14}C activity <88 pmc was defined as indicative of the presence of groundwater recharged before 1952.

In this study, the age distributions of samples are classified as pre-modern, modern, or mixed ages (table B4). Groundwater with tritium activity <0.5 TU and ^{14}C <88 pmc is designated as pre-modern, defined as having been recharged before 1952. Groundwater with tritium activities >0.5 TU and ^{14}C >88 pmc is designated as modern, defined as having been recharged after 1952. Samples with pre-modern and modern components are designated as mixed-age groundwater, which includes substantial fractions of old and young waters. In reality, pre-modern groundwater could contain very small fractions of modern groundwater, and modern groundwater could contain small fractions of pre-modern groundwater. Although more sophisticated lumped parameter models that incorporate mixing are available for analyzing age distributions (for example, Cook and Böhlke, 2000), use of these alternative models to characterize age mixtures was beyond the scope of this report. Rather, classification into modern (recharged after 1952), mixed-age, and pre-modern (recharged before 1952) categories was sufficient to provide an appropriate and useful characterization for the purposes of examining groundwater quality.

Classification of Geochemical Condition

Geochemical conditions investigated as potential explanatory variables in this report include oxidation-reduction (redox) characteristics, pH, and DO (table B5). Redox conditions influence the mobility of many organic and inorganic constituents (McMahon and Chapelle, 2008). Along groundwater flow paths, redox conditions commonly proceed along a well-documented sequence of Terminal Electron Acceptor Processes (TEAPs); one TEAP typically dominates at a particular time and aquifer location (Chapelle and others, 1995; Chapelle, 2001). The predominant TEAPs (in order of least to most reducing) are oxygen-reducing, nitrate-reducing, manganese-reducing, iron-reducing, sulfate-reducing, and methanogenesis. The presence of redox-sensitive chemical species suggesting more than one TEAP may indicate mixed waters from different redox zones upgradient of the site, a well screened across more than one redox zone, or spatial heterogeneity in microbial activity in the aquifer. Different redox elements (for example, iron, manganese, and sulfur) tend not to reach overall equilibrium in most natural water systems (Lindberg and Runnels, 1984); therefore, a single redox measurement usually cannot represent the system, further complicating the assessment of redox conditions. pH is the measure of hydrogen-ion activity in a water sample and is sensitive to a number of geochemical reactions in addition to redox conditions. DO is a measure of how much oxygen is dissolved in a water sample.

Table B1. Land-use classification for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[GAMA, Groundwater Ambient Monitoring and Assessment Program]

	Land-use category			Classification
	Agricultural (percent)	Natural (percent)	Urban (percent)	
Northern Coast Ranges (NOCO) study unit				
Study unit as a whole	29	60	11	Natural
Study unit based on buffer areas surrounding grid sites	37	46	17	Mixed
Study unit based on buffer areas surrounding California Department of Public Health (CDPH) sites	19	56	25	Natural
Interior Basins (NOCO-IN) study area				
Study area as a whole	33	60	7	Natural
Study area based on buffer areas surrounding grid sites	43	47	10	Mixed
Study area based on buffer areas surrounding CDPH sites	36	45	19	Mixed
Coastal Basins (NOCO-CO) study area				
Study area as a whole	25	60	14	Natural
Study area based on buffer areas surrounding grid sites	31	46	23	Mixed
Study area based on buffer areas surrounding CDPH sites	11	61	28	Natural

Table B2. Land-use percentages, septic tank density, and leaking (or formerly leaking) underground fuel tank (LUFT) density for areas within a 500-meter radius of the U.S. Geological Survey (USGS) GAMA grid sites in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[USGS-GAMA site identification numbers: NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Abbreviation:** tanks/km², tanks per square kilometer]

USGS-GAMA site identification number	Land-use percentages ¹			Septic tank density (tanks/km ²)	LUFT density (tanks/km ²)
	Agricultural	Natural	Urban		
Interior Basins study area					
NOCO-IN-01	52	27	21	30.8	0.05
NOCO-IN-02	42	6	52	8.0	1.71
NOCO-IN-03	6	93	1	12.5	0.01
NOCO-IN-04	49	14	37	10.7	0.01
NOCO-IN-05	28	9	63	21.6	0.81
NOCO-IN-06	0	100	0	4.1	0.02
NOCO-IN-07	22	78	0	3.5	0.01
NOCO-IN-08	80	20	0	2.3	0.79
NOCO-IN-09	56	44	0	24.9	0.22
NOCO-IN-10	82	18	0	4.3	0.02
NOCO-IN-11	60	2	38	14.7	1.77
NOCO-IN-12	80	19	1	3.8	0.01
NOCO-IN-13	69	22	9	21.8	0.80
NOCO-IN-14	70	25	5	4.7	0.02
NOCO-IN-15	0	99	1	0.1	0
NOCO-IN-16	1	99	0	0.1	0
NOCO-IN-17	88	10	2	4.9	0.08
NOCO-IN-18	93	4	3	12.1	0.01
NOCO-IN-19	78	22	0	5.4	0.05
NOCO-IN-20	0	100	0	3.3	0.02

Table B2. Land-use percentages, septic tank density, and leaking (or formerly leaking) underground fuel tank (LUFT) density for areas within a 500-meter radius of the U.S. Geological Survey (USGS) GAMA grid sites in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.—Continued[USGS-GAMA site identification numbers: NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Abbreviation:** tanks/km², tanks per square kilometer]

USGS-GAMA site identification number	Land-use percentages ¹			Septic tank density (tanks/km ²)	LUFT density (tanks/km ²)
	Agricultural	Natural	Urban		
Interior Basins study area—Continued					
NOCO-IN-21	84	2	14	11.4	0.01
NOCO-IN-22	23	77	0	1.7	0.01
NOCO-IN-23	0	100	0	12.4	0
NOCO-IN-24	48	29	23	7.2	0.06
NOCO-IN-25	56	44	0	4.8	0.08
NOCO-IN-26	0	100	0	0.3	0
NOCO-IN-27	46	54	0	1.3	0.07
NOCO-IN-28	0	100	0	4.1	0.02
Coastal Basins study area					
NOCO-CO-01	0	23	77	21.4	0.40
NOCO-CO-02	78	22	0	0.4	0.02
NOCO-CO-03	34	19	47	7.6	0.57
NOCO-CO-04	0	89	11	3.2	0.14
NOCO-CO-05	0	40	60	39.3	0.50
NOCO-CO-06	0	97	3	7.7	0.43
NOCO-CO-07	0	78	22	1.2	0.02
NOCO-CO-08	58	1	41	31.1	0.79
NOCO-CO-09	37	46	17	7.8	0.45
NOCO-CO-10	5	95	0	3.5	0.02
NOCO-CO-11	61	9	30	8.1	1.49
NOCO-CO-12	49	31	21	20.3	0.03
NOCO-CO-13	0	14	86	23.1	0.04
NOCO-CO-14	0	47	53	44.9	0.10
NOCO-CO-15	0	83	17	0.6	0.01
NOCO-CO-16	0	100	0	0.4	0.02
NOCO-CO-17	90	10	0	17.3	0.02
NOCO-CO-18	2	63	35	9.5	0.11
NOCO-CO-19	99	1	0	2.1	0.46
NOCO-CO-20	1	80	19	1.7	0.18
NOCO-CO-21	80	1	19	2.0	1.17
NOCO-CO-22	88	11	1	5.8	0.05
NOCO-CO-23	44	22	34	24.9	0.37
NOCO-CO-24	1	99	0	12.4	0.11
NOCO-CO-25	9	50	41	7.3	0.63
NOCO-CO-26	0	82	18	1.3	0.33
NOCO-CO-27	93	5	2	1.1	0.04
NOCO-CO-28	0	100	0	1.4	0.19
NOCO-CO-29	95	5	0	3.4	0.10
NOCO-CO-30	1	58	41	4.8	0.19

¹Land-use data from Nakagaki and others (2007). Land-use percentages within 500-meter radius of each USGS-GAMA grid site (Johnson and Belitz, 2009).

Table B3. Well and construction information and hydrologic conditions for U.S. Geological Survey (USGS) GAMA grid sites in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[USGS-GAMA site identification numbers: NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Abbreviations:** GAMA, Groundwater Ambient Monitoring and Assessment Program; ft bls, feet below land surface; DOGGR, Division of Oil, Gas, and Geothermal Resources (California Department of Conservation); km, kilometer; LSD, land-surface datum; na, not available; DOM, domestic well; PSW, public-supply well (California Department of Health); IRR, irrigation well; IND, industrial well; INS, institutional well; SPR, spring site; STK, stock well; UNS, unused site; °C, degrees Celsius]

General information			Construction information				Hydrologic conditions			
USGS-GAMA site identification number	Altitude of LSD (ft above NAVD 88) ¹	Site type	Well depth (ft bls)	Depth to top of perforations (ft bls)	Depth to bottom of perforations (ft bls)	Length from top of uppermost perforated interval to bottom of well (ft bls)	Aridity index ²	Field water temperature (°C)	Distance to nearest hot spring listed in the USGS Mineral Resources Data System (km)	Distance to nearest geothermal well listed in the DOGGR database (km)
Interior Basins study area										
NOCO-IN-01	748	IRR	165	79	157	78	0.85	20.0	12.3	36.8
NOCO-IN-02	363	PSW	120	60	120	60	0.88	17.5	12.8	27.8
NOCO-IN-03	1,673	DOM	58	33	58	25	1.71	12.5	4.9	85.3
NOCO-IN-04	1,385	DOM	55	na	na	na	1.16	14.0	19.8	54.4
NOCO-IN-05	1,633	UNS	60	na	na	na	1.73	16.0	1.3	88.9
NOCO-IN-06	1,342	DOM	136	na	na	na	0.97	15.0	24.4	89.6
NOCO-IN-07	523	PSW	40	23	38	15	0.83	14.5	12.0	15.3
NOCO-IN-08	483	PSW	40	25	40	15	0.82	14.0	18.6	9.2
NOCO-IN-09	713	PSW	400	148	385	237	0.84	20.0	9.8	33.9
NOCO-IN-10	553	PSW	101	35	94	59	0.80	16.0	8.4	19.9
NOCO-IN-11	958	DOM	100	60	100	40	0.92	18.0	15.2	38.5
NOCO-IN-12	558	IRR	215	40	215	175	0.82	19.0	10.2	17.0
NOCO-IN-13	599	PSW	36	15	36	21	0.82	16.0	3.0	25.1
NOCO-IN-14	658	IRR	135	75	135	60	0.81	18.5	4.9	20.8
NOCO-IN-15	1,856	PSW	300	80	220	140	1.07	15.5	4.2	47.5
NOCO-IN-16	1,845	PSW	152	39	152	113	1.03	15.0	4.0	48.5
NOCO-IN-17	1,368	DOM	75	na	na	na	0.78	18.5	18.3	14.6
NOCO-IN-18	1,360	IRR	170	50	160	110	0.65	19.0	30.6	1.6
NOCO-IN-19	823	DOM	80	21	80	59	0.88	18.5	14.2	37.7
NOCO-IN-20	1,470	DOM	75	na	na	na	0.75	23.0	27.7	0.82

Table B3. Well and construction information and hydrologic conditions for U.S. Geological Survey (USGS) GAMA grid sites in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.—Continued

[USGS-GAMA site identification numbers: NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Abbreviations:** GAMA, Groundwater Ambient Monitoring and Assessment Program; ft bls, feet below land surface; DOGGR, Division of Oil, Gas, and Geothermal Resources (California Department of Conservation); km, kilometer; LSD, land-surface datum; na, not available; DOM, domestic well; PSW, public-supply well (California Department of Health); IRR, irrigation well; IND, industrial well; INS, institutional well; SPR, spring site; STK, stock well; UNS, unused site; °C, degrees Celsius]

General information			Construction information				Hydrologic conditions			
USGS-GAMA site identification number	Altitude of LSD (ft above NAVD 88) ¹	Site type	Well depth (ft bls)	Depth to top of perforations (ft bls)	Depth to bottom of perforations (ft bls)	Length from top of uppermost perforated interval to bottom of well (ft bls)	Aridity index ²	Field water temperature (°C)	Distance to nearest hot spring listed in the USGS Mineral Resources Data System (km)	Distance to nearest geothermal well listed in the DOGGR database (km)
Interior Basins study area—Continued										
NOCO-IN-21	361	IRR	52	40	48	8	0.64	18.5	30.9	1.4
NOCO-IN-22	801	IRR	180	na	na	na	0.80	25.0	0.61	1.0
NOCO-IN-23	1,399	DOM	142	40	142	102	0.98	15.0	23.7	93.7
NOCO-IN-24	1,380	PSW	116	70	115	45	0.61	18.5	28.5	2.3
NOCO-IN-25	1,072	PSW	283	56	217	161	0.83	19.0	16.0	0.37
NOCO-IN-26	1,154	PSW	126	30	114	84	0.48	18.0	5.3	42.5
NOCO-IN-27	1,776	PSW	155	40	115	75	0.71	18.0	22.0	6.0
NOCO-IN-28	1,345	DOM	41	na	na	na	0.98	15.5	24.3	87.8
Coastal Basins study area										
NOCO-CO-01	71	DOM	25	16	18	2	0.93	13.0	36.8	88.0
NOCO-CO-02	48	PSW	55	35	55	20	0.86	13.5	20.2	51.1
NOCO-CO-03	38	PSW	380	280	370	90	1.19	11.0	111	221
NOCO-CO-04	243	PSW	325	na	na	na	1.28	12.5	109	212
NOCO-CO-05	126	PSW	15	10	10	0	0.99	12.0	42.3	83.9
NOCO-CO-06	95	PSW	37	20	37	17	1.01	12.0	37.2	75.8
NOCO-CO-07	124	PSW	130	10	130	120	1.54	11.5	102	260
NOCO-CO-08	33	PSW	215	150	210	60	1.29	14.5	98.6	233
NOCO-CO-09	5	IND	397	356	376	20	na	18.0	102	225
NOCO-CO-10	115	IRR	225	210	220	10	1.49	13.0	89.8	225
NOCO-CO-11	50	PSW	115	43	103	60	1.23	14.5	103	202
NOCO-CO-12	56	DOM	62	na	na	na	1.25	15.5	109	220
NOCO-CO-13	318	PSW	100	34	100	66	1.48	13.0	99.1	247
NOCO-CO-14	42	INS	32	20	32	12	1.73	13.0	135	328
NOCO-CO-15	27	PSW	85	30	85	55	2.15	10.5	113	302

Table B3. Well and construction information and hydrologic conditions for U.S. Geological Survey (USGS) GAMA grid sites in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.—Continued

[USGS-GAMA site identification numbers: NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Abbreviations:** GAMA, Groundwater Ambient Monitoring and Assessment Program; ft bls, feet below land surface; DOGGR, Division of Oil, Gas, and Geothermal Resources (California Department of Conservation); km, kilometer; LSD, land-surface datum; na, not available; DOM, domestic well; PSW, public-supply well (California Department of Health); IRR, irrigation well; IND, industrial well; INS, institutional well; SPR, spring site; STK, stock well; UNS, unused site; °C, degrees Celsius]

General information			Construction information				Hydrologic conditions			
USGS-GAMA site identification number	Altitude of LSD (ft above NAVD 88) ¹	Site type	Well depth (ft bls)	Depth to top of perforations (ft bls)	Depth to bottom of perforations (ft bls)	Length from top of uppermost perforated interval to bottom of well (ft bls)	Aridity index ²	Field water temperature (°C)	Distance to nearest hot spring listed in the USGS Mineral Resources Data System (km)	Distance to nearest geothermal well listed in the DOGGR database (km)
Coastal Basins study area—Continued										
NOCO-CO-16	56	PSW	75	38	58	20	2.17	13.0	132	329
NOCO-CO-17	153	PSW	50	23	50	27	1.28	15.0	96.0	196
NOCO-CO-18	39	PSW	65	55	60	5	1.26	14.5	106	207
NOCO-CO-19	37	DOM	50	na	na	na	1.17	13.5	109	208
NOCO-CO-20	93	SPR	at LSD	at LSD	at LSD	at LSD	1.24	12.5	113	208
NOCO-CO-21	24	PSW	168	146	166	20	1.18	14.5	112	209
NOCO-CO-22	15	IRR	35	na	na	na	1.87	13.0	142	339
NOCO-CO-23	39	PSW	45	35	45	10	2.00	14.5	142	341
NOCO-CO-24	24	STK	15	na	na	na	1.70	12.5	139	332
NOCO-CO-25	13	IND	193	175	193	18	1.20	14.0	99.8	225
NOCO-CO-26	267	PSW	15	15	15	0	1.55	12.0	102	252
NOCO-CO-27	8	DOM	400	na	na	na	1.20	15.5	110	216
NOCO-CO-28	18	DOM	120	na	na	na	1.21	14.0	103	234
NOCO-CO-29	10	IRR	200	na	na	na	1.26	13.0	101	234
NOCO-CO-30	390	PSW	196	166	196	30	1.38	15.5	99.6	200

¹LSD is a datum plane that is approximately at land surface at each site. The altitude of the LSD is described in feet above the North American Vertical Datum of 1988.

²Aridity index is average annual precipitation (PRISM Group, Oregon State University, 2012) divided by average annual evapotranspiration (Flint and Flint, 2007).

Table B4. Groundwater age classification and the associated data for U.S. Geological Survey (USGS) GAMA grid sites in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[Groundwater age classification: Modern, groundwater recharged since 1952; Pre-modern, groundwater recharged before 1952; Mixed-age, groundwater recharged before and after 1952. Well depth classifications: Shallow, wells with depth above 70 ft bls; Mixed-depth, wells with perforations beginning above 70 ft bls and ending below 70 ft bls; Deep, wells with perforations entirely below 70 ft bls; Mixed-depth or Deep, wells with total well depth below 70 ft bls and unknown perforation information. USGS-GAMA site identification numbers: NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. Abbreviations: ft bls, feet below land surface; ¹⁴C, carbon-14; TU, tritium units; nc, not collected; <, less than; LSD, land-surface datum]

USGS-GAMA site identification number	Tritium (TU)	Percent modern carbon	Uncorrected ¹⁴ C age (years)	Groundwater age classification	Well depth classifications
Interior Basins study area					
NOCO-IN-01	0.09	40	7,260	Pre-modern	Deep
NOCO-IN-02	1.41	83	1,460	Mixed-age	Mixed-depth
NOCO-IN-03	1.31	106	<1,000	Modern	Shallow
NOCO-IN-04	1.81	99	<1,000	Modern	Shallow
NOCO-IN-05	1.47	58	4,310	Mixed-age	Shallow
NOCO-IN-06	-0.06	32	8,970	Pre-modern	Mixed-depth or Deep
NOCO-IN-07	2.82	105	<1,000	Modern	Shallow
NOCO-IN-08	2.16	88	<1,000	Modern	Shallow
NOCO-IN-09	0.09	22	12,030	Pre-modern	Deep
NOCO-IN-10	2.16	103	<1,000	Modern	Mixed-depth
NOCO-IN-11	1.37	84	1,380	Mixed-age	Mixed-depth
NOCO-IN-12	1.06	86	1,130	Mixed-age	Mixed-depth
NOCO-IN-13	3.57	105	<1,000	Modern	Shallow
NOCO-IN-14	0.81	57	4,520	Mixed-age	Deep
NOCO-IN-15	0.37	28	10,290	Pre-modern	Deep
NOCO-IN-16	1.78	79	1,850	Mixed-age	Mixed-depth
NOCO-IN-17	0.47	77	2,060	Pre-modern	Mixed-depth or Deep
NOCO-IN-18	1.31	78	1,900	Mixed-age	Mixed-depth
NOCO-IN-19	2.06	105	<1,000	Modern	Mixed-depth
NOCO-IN-20	1.41	29	9,930	Mixed-age	Mixed-depth or Deep
NOCO-IN-21	1.81	106	<1,000	Modern	Shallow
NOCO-IN-22	1.00	25	11,200	Mixed-age	Mixed-depth or Deep
NOCO-IN-23	2.03	93	<1,000	Modern	Mixed-depth
NOCO-IN-24	1.97	89	<1,000	Modern	Mixed-depth
NOCO-IN-25	1.78	59	4,120	Mixed-age	Mixed-depth
NOCO-IN-26	2.53	99	<1,000	Modern	Mixed-depth
NOCO-IN-27	1.06	5	23,780	Mixed-age	Mixed-depth
NOCO-IN-28	0.21	10	17,980	Pre-modern	Shallow

Table B4. Groundwater age classification and the associated data for U.S. Geological Survey (USGS) GAMA grid sites in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.—Continued

[**Groundwater age classification:** Modern, groundwater recharged since 1952; Pre-modern, groundwater recharged before 1952; Mixed-age, groundwater recharged before and after 1952. **Well depth classifications:** Shallow, wells with depth above 70 ft bls; Mixed-depth, wells with perforations beginning above 70 ft bls and ending below 70 ft bls; Deep, wells with perforations entirely below 70 ft bls; Mixed-depth or Deep, wells with total well depth below 70 ft bls and unknown perforation information. **USGS-GAMA site identification numbers:** NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Abbreviations:** ft bls, feet below land surface; ¹⁴C, carbon-14; TU, tritium units; nc, not collected; <, less than; LSD, land-surface datum]

USGS-GAMA site identification number	Tritium (TU)	Percent modern carbon	Uncorrected ¹⁴ C age (years)	Groundwater age classification	Well depth classifications
Coastal Basins study area					
NOCO-CO-01	1.00	57	4,410	Mixed-age	Shallow
NOCO-CO-02	1.50	107	<1,000	Modern	Shallow
NOCO-CO-03	0.09	75	2,270	Pre-modern	Deep
NOCO-CO-04	0.00	14	15,540	Pre-modern	Mixed-depth or Deep
NOCO-CO-05	1.94	108	<1,000	Modern	Shallow
NOCO-CO-06	2.25	92	<1,000	Modern	Shallow
NOCO-CO-07	3.91	103	<1,000	Modern	Mixed-depth
NOCO-CO-08	0.68	70	2,760	Mixed-age	Deep
NOCO-CO-09	-0.03	9	19,060	Pre-modern	Deep
NOCO-CO-10	1.15	75	2,290	Mixed-age	Deep
NOCO-CO-11	1.69	95	<1,000	Modern	Mixed-depth
NOCO-CO-12	0.37	79	1,830	Pre-modern	Shallow
NOCO-CO-13	1.72	90	<1,000	Modern	Mixed-depth
NOCO-CO-14	1.47	nc	nc	Modern or Mixed-age	Shallow
NOCO-CO-15	1.88	101	<1,000	Modern	Mixed-depth
NOCO-CO-16	2.31	104	<1,000	Modern	Mixed-depth
NOCO-CO-17	2.38	93	<1,000	Modern	Shallow
NOCO-CO-18	1.78	86	1,110	Mixed-age	Shallow
NOCO-CO-19	1.63	104	<1,000	Modern	Shallow
NOCO-CO-20	1.66	67	3,210	Mixed-age	at LSD
NOCO-CO-21	0.75	58	4,280	Mixed-age	Deep
NOCO-CO-22	1.47	107	<1,000	Modern	Shallow
NOCO-CO-23	1.88	102	<1,000	Modern	Shallow
NOCO-CO-24	1.25	103	<1,000	Mixed-age	Shallow
NOCO-CO-25	-0.03	10	18,510	Pre-modern	Deep
NOCO-CO-26	1.94	107	<1,000	Modern	Shallow
NOCO-CO-27	-0.06	5	23,550	Pre-modern	Mixed-depth or Deep
NOCO-CO-28	-0.06	53	15,010	Pre-modern	Mixed-depth or Deep
NOCO-CO-29	2.88	77	1,990	Mixed-age	Mixed-depth or Deep
NOCO-CO-30	3.22	60	4,100	Mixed-age	Deep

¹Sample oxidation-reduction state was anoxic (table B5), and stable carbon isotope ratio was greater than -10 per mil, suggesting that the carbon isotope composition of the sample may have been altered by methanogenesis. The uncorrected carbon-14 apparent age thus may be too old.

Table B5. Oxidation-reduction classification based on McMahon and Chapelle (2008), dissolved oxygen concentration, and pH for U.S. Geological Survey (USGS) GAMA grid sites in the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[**Oxidation-reduction classification:** Mn-red, manganese-reducing; Fe-red, iron-reducing; MnFe-red, manganese- and iron-reducing; NO₃MnFe-red, nitrate-, manganese-, and iron-reducing; NO₃Mn-red, nitrate- and manganese-reducing. **USGS-GAMA site identification numbers:** NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Abbreviations:** GAMA, Groundwater Ambient Monitoring and Assessment Program; mg/L, milligrams per liter; <, less than]

USGS-GAMA site identification number	Dissolved oxygen (mg/L)	pH (standard units)	Oxidation-reduction classification ¹	USGS-GAMA site identification number	Dissolved oxygen (mg/L)	pH (standard units)	Oxidation-reduction classification ¹
Interior Basins study area				Coastal Basins study area			
NOCO-IN-01	0.7	7.5	Anoxic Mn-red	NOCO-CO-01	<0.2	7.2	Anoxic MnFe-red
NOCO-IN-02	7.0	7.2	Oxic	NOCO-CO-02	11.8	6.5	Oxic
NOCO-IN-03	0.7	6.1	Anoxic Fe-red	NOCO-CO-03	1.3	7.7	Oxic
NOCO-IN-04	10.9	6.3	Mixed (Oxic/Anoxic Fe-red)	NOCO-CO-04	<0.2	7.5	Anoxic MnFe-red
NOCO-IN-05	0.3	6.7	Anoxic MnFe-red	NOCO-CO-05	5.2	5.6	Oxic
NOCO-IN-06	0.4	7.2	Anoxic MnFe-red	NOCO-CO-06	4.1	6.5	Oxic
NOCO-IN-07	1.3	6.0	Oxic	NOCO-CO-07	10.4	6.5	Oxic
NOCO-IN-08	5.5	6.6	Oxic	NOCO-CO-08	<0.2	7.5	Anoxic suboxic
NOCO-IN-09	0.1	7.5	Anoxic MnFe-red	NOCO-CO-09	<0.2	7.5	Anoxic MnFe-red
NOCO-IN-10	5.0	6.4	Oxic	NOCO-CO-10	3.3	7.8	Oxic
NOCO-IN-11	0.3	6.8	Anoxic NO ₃ MnFe-red	NOCO-CO-11	1.3	6.7	Oxic
NOCO-IN-12	2.2	6.9	Oxic	NOCO-CO-12	1.1	6.8	Oxic
NOCO-IN-13	3.8	6.1	Oxic	NOCO-CO-13	7.1	6.3	Oxic
NOCO-IN-14	<0.2	7.3	Anoxic MnFe-red	NOCO-CO-14	9.7	7.7	Oxic
NOCO-IN-15	<0.2	7.1	Anoxic suboxic	NOCO-CO-15	6.9	6.4	Oxic
NOCO-IN-16	1.9	6.4	Oxic	NOCO-CO-16	7.9	6.7	Oxic
NOCO-IN-17	6.2	6.9	Oxic	NOCO-CO-17	3.1	6.5	Oxic
NOCO-IN-18	1.3	6.7	Oxic	NOCO-CO-18	0.3	6.5	Anoxic NO ₃ MnFe-red
NOCO-IN-19	2.6	6.0	Oxic	NOCO-CO-19	0.2	6.6	Anoxic NO ₃ Mn-red
NOCO-IN-20	0.3	6.3	Anoxic MnFe-red	NOCO-CO-20	6.1	7.5	Oxic
NOCO-IN-21	<0.2	6.8	Anoxic MnFe-red	NOCO-CO-21	0.4	7.0	Anoxic NO ₃ MnFe-red
NOCO-IN-22	0.5	7.4	Anoxic MnFe-red	NOCO-CO-22	7.5	6.7	Oxic
NOCO-IN-23	1.9	7.2	Oxic	NOCO-CO-23	6.3	6.2	Oxic
NOCO-IN-24	3.8	7.3	Oxic	NOCO-CO-24	0.8	6.8	Anoxic MnFe-red
NOCO-IN-25	0.7	7.5	Anoxic Fe-red	NOCO-CO-25	0.3	6.7	Anoxic Mn-red
NOCO-IN-26	4.8	6.7	Oxic	NOCO-CO-26	4.2	5.5	Oxic
NOCO-IN-27	<0.2	6.0	Anoxic MnFe-red	NOCO-CO-27	0.2	6.8	Anoxic MnFe-red
NOCO-IN-28	<0.2	7.2	Anoxic Mn-red	NOCO-CO-28	0.4	7.2	Anoxic MnFe-red
				NOCO-CO-29	0.4	7.3	Anoxic NO ₃ Mn-red
				NOCO-CO-30	4.6	7.1	Oxic

¹Data from Mathany and others (2011) were used for the classification, in addition to the dissolved oxygen and pH data presented in this table.

Appendix C. Calculation of Aquifer-Scale Proportions

Table C1. Aquifer-scale proportions calculated for the Interior Basins study area, using grid-based and spatially weighted methods for those constituents that met criteria for additional evaluation in the status assessment, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[**Relative-concentration categories:** *high*, concentration greater than water-quality benchmark; *moderate*, concentration is less than the benchmark and is greater than or equal to 0.1 of the benchmark (for organic and special-interest constituents) or 0.5 of the benchmark (for inorganic constituents); *low*, concentration less than 0.1 of benchmark (for organic constituents and special-interest constituents) or 0.5 of the benchmark (for inorganic constituents). **Abbreviations:** Groundwater Ambient Monitoring and Assessment Program; CDPH, California Department of Public Health; TDS, total dissolved solids; MTBE, methyl *tert*-butyl ether; µg/L, micrograms per liter]

Constituent	Raw detection frequency, in percent ¹			Spatially weighted aquifer-scale proportion, in percent ¹			Grid-based aquifer-scale proportion, in percent			90 percent confidence interval for grid-based high proportion, in percent ²	
	Number of sites	Moderate	High	Number of cells	Moderate aquifer proportion	High aquifer proportion	Number of sites	Moderate aquifer proportion	High aquifer proportion	Lower limit	Upper limit
Inorganic constituents with health-based benchmarks											
Trace Elements											
Arsenic	71	2.8	4.2	28	5.4	7.1	28	7.1	7.1	2.1	18.4
Barium	71	0	1.4	28	0	3.6	28	0	3.6	0.6	13.1
Boron ³	28	7.1	17.9	28	7.1	17.9	28	7.1	17.9	8.5	31.9
Nickel	71	0	0	28	0	0	28	0	0	0	6.6
Nutrients											
Nitrate	92	0	0	28	0	0	28	0	0	0	6.6
Inorganic constituents with aesthetic-based benchmarks											
Iron	60	8.3	26.7	28	12.2	27.7	28	10.7	25.0	13.7	39.9
Manganese	62	3.2	37.1	28	7.1	46.1	28	7.1	42.9	28.5	58.3
TDS	58	8.6	0	28	13.7	0	28	14.3	0	0	6.6
Organic constituents with health-based benchmarks											
Volatile organic compounds											
Chloroform (Trichloromethane)	53	1.9	0	28	2.3	0	28	3.7	0	0	6.6
Vinyl chloride (Chloroethene)	52	1.9	0	28	3.6	0	28	3.6	0	0	6.6
MTBE ⁴	67	0	1.5	28	0	0.5	28	0	0	0	6.6
Pesticides and pesticide degradates											
Simazine	70	0	0	28	0	0	28	0	0	0	6.6
Special-interest constituent with health-based benchmark											
Perchlorate	78	0	0	28	0	0	28	0	0	0	6.6

¹Based on most recent analyses for each CDPH site during June 1, 2006–June 1, 2009, combined with data from USGS-GAMA grid sites.

²Based on the Jeffrey's interval for the binomial distribution (Brown and others, 2001).

³Boron was not sampled by the CDPH during June 1, 2006–June 1, 2009.

⁴High relative-concentrations of MTBE were reported multiple times in the CDPH water-quality database from one site in the NOCO-IN study area during the current period of study. Sixteen groundwater samples were collected by the CDPH between August 16, 2006, and May 20, 2009, from this site for analysis of MTBE; 15 had concentrations of MTBE between 13 and 21 µg/L (MCL-CA = 13 µg/L), and 1 was a non-detection (most recent value, May 20, 2009). On that basis, it was determined that the non-detection of MTBE reported as the most recent value in the CDPH water-quality database does not represent present-day conditions. The MTBE sample collected on April 30, 2009 (17 µg/L) was substituted in order to more accurately reflect present-day conditions.

Table C2. Aquifer-scale proportions calculated for the Coastal Basins study area, using grid-based and spatially weighted methods for those constituents that met criteria for additional evaluation in the status assessment, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[**Relative-concentration categories:** *high*, concentration greater than water-quality benchmark; *moderate*, concentration is less than the benchmark and is greater than or equal to 0.1 of the benchmark (for organic and special-interest constituents) or 0.5 of the benchmark (for inorganic constituents); *low*, concentration less than 0.1 of benchmark (for organic constituents and special-interest constituents) or 0.5 of the benchmark (for inorganic constituents). **Abbreviations:** GAMA, Groundwater Ambient Monitoring and Assessment Program; CDPH, California Department of Public Health; %, percent; TDS, total dissolved solids; MTBE, methyl *tert*-butyl ether]

Constituent	Raw detection frequency, in percent ¹			Spatially weighted aquifer-scale proportion, in percent ¹			Grid-based aquifer-scale proportion, in percent			90 percent confidence interval for grid-based high proportion, in percent ²	
	Number of sites	Moderate	High	Number of cells	Moderate aquifer proportion	High aquifer proportion	Number of sites	Moderate aquifer proportion	High aquifer proportion	Lower limit	Upper limit
Inorganic constituents with health-based benchmarks											
Trace Elements											
Arsenic	59	3.4	1.7	30	6.7	0.2	30	6.7	0	0	6.2
Barium	60	0	0	30	0	0	30	0	0	0	6.2
Boron ³	30	3.3	0	30	3.3	0	30	3.3	0	0	6.2
Nickel	62	1.7	0	30	3.3	0	30	3.3	0	0	6.2
Nutrients											
Nitrate	128	3.1	0	30	2.7	0	29	6.9	0	0	6.2
Inorganic constituents with aesthetic-based benchmarks											
Iron	61	4.9	16.4	30	1.0	23.9	30	0	26.7	15.3	41.2
Manganese	61	4.9	24.6	30	8.3	36.7	30	10.0	36.7	23.5	51.6
TDS	60	1.7	0	30	0.6	0	30	0	0	0	6.2
Organic constituents with health-based benchmarks											
Volatile organic compounds											
Chloroform (Trichloromethane)	65	1.5	0	30	0.8	0	30	0	0	0	6.2
Vinyl chloride (Chloroethene)	63	0	0	30	0	0	30	0	0	0	6.2
MTBE	67	0	0	30	0	0	30	0	0	0	6.2
Pesticides and pesticide degradates											
Simazine	53	0	0	30	0	0	30	0	0	0	6.2
Special-interest constituent with health-based benchmark											
Perchlorate	84	1.2	0	30	1.1	0	30	3.3	0	0	6.2

¹Based on most recent analyses for each CDPH site during June 1, 2006–June 1, 2009, combined with data from USGS-GAMA grid sites.

²Based on the Jeffrey's interval for the binomial distribution (Brown and others, 2001).

³Boron was not sampled by the CDPH during June 1, 2006–June 1, 2009.

Appendix D. Comparison of CDPH and GAMA Priority Basin Data

CDPH and USGS-GAMA data were compared to assess the validity of combining data for inorganic constituents from these different sources. Concentrations of inorganic constituents (calcium, magnesium, sodium, alkalinity, chloride, sulfate, TDS, and nitrate), which generally are prevalent at concentrations substantially greater than reporting levels, were compared for each site by using data from both sources. The USGS and CDPH water-quality databases contained data for major ions, trace elements, TDS, or nitrate (as nitrogen) for 17 of 58 sites. Wilcoxon rank-sum tests of paired analyses for these eight constituents indicated no significant differences between USGS-GAMA and CDPH data for these constituents. Although differences between the paired datasets occurred for some sites (most notably nitrate, as nitrogen), most sample pairs plotted close to a 1:1 line (fig. D1). These direct

comparisons indicated that the USGS-GAMA and CDPH inorganic data were not significantly different.

Major-ion data from USGS-GAMA grid sites were plotted on a trilinear diagram (Piper, 1944) along with all CDPH major-ion data to determine whether the groundwater types in the USGS-GAMA grid sites were similar to groundwater types observed historically in the study unit. Trilinear diagrams show the relative abundance of major cations and anions (on a charge equivalent basis) as a percentage of the total ion content of the water (fig. D2). Trilinear diagrams often are used to define groundwater type (Hem, 1989). All cation and anion data in the CDPH water-quality database with a cation and anion imbalance less than 10 percent were retrieved and plotted on the trilinear diagram for comparison with USGS-GAMA grid site data.

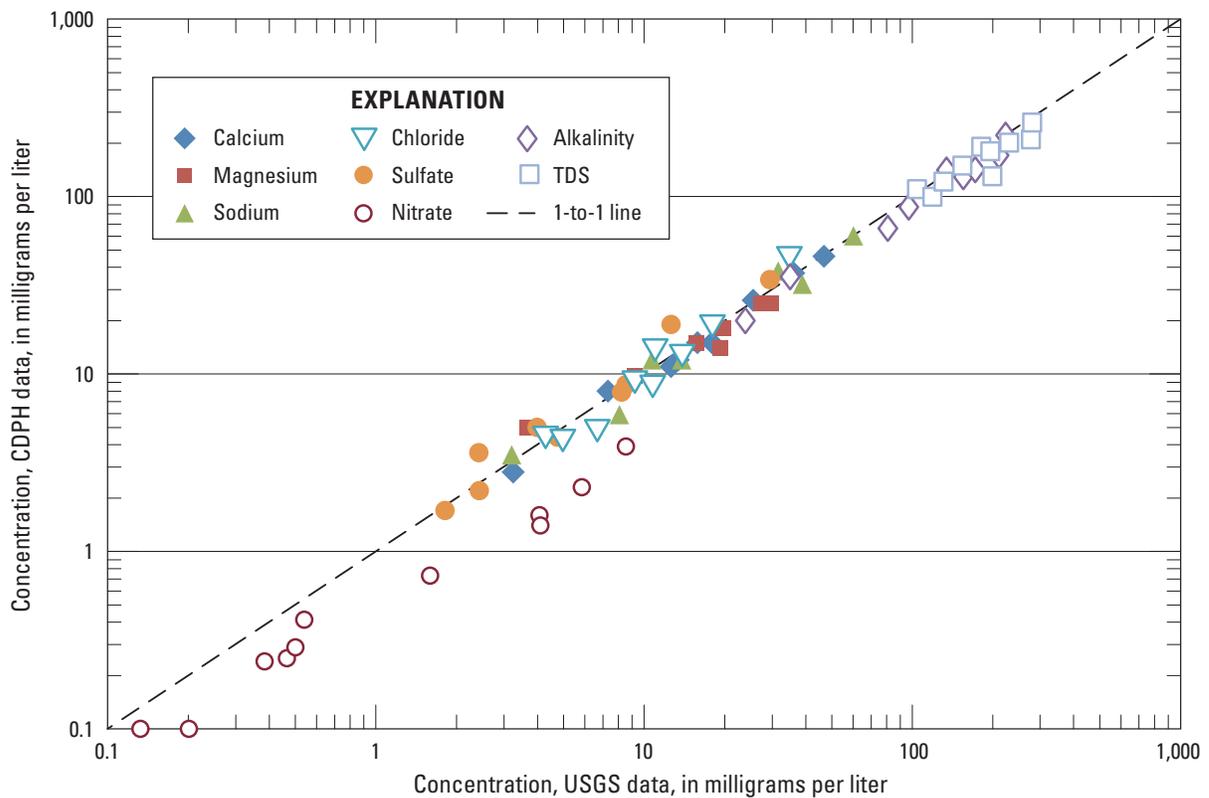


Figure D1. Paired inorganic constituent concentrations from sites sampled by the USGS-GAMA Program from June to November 2009 and from the CDPH database for the same sites during the period June 1, 2006–June 1, 2009, Northern Coast Ranges study unit, California GAMA Priority Basin Project.

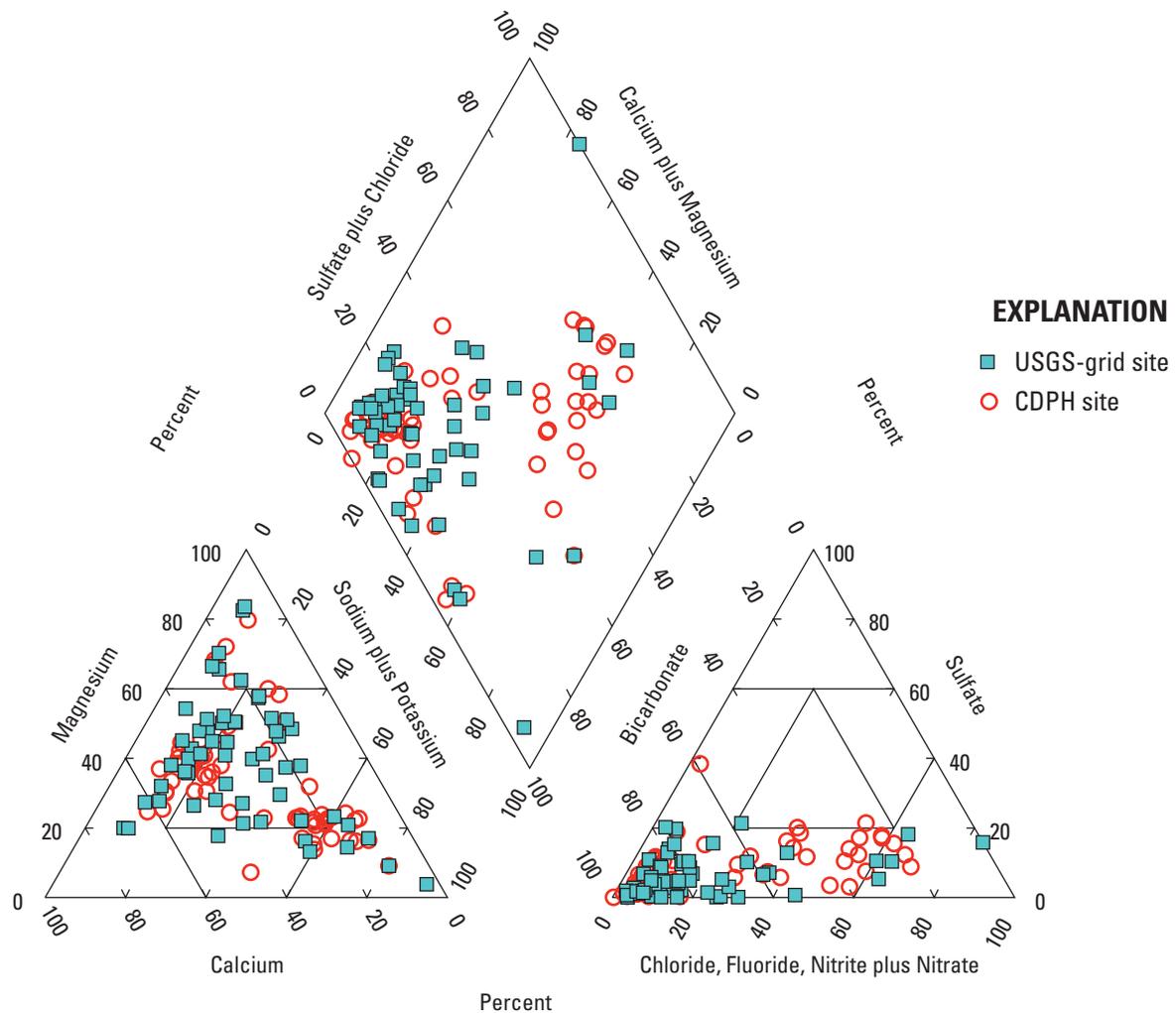


Figure D2. Selected inorganic data from USGS-GAMA grid sites and from all sites in the CDPH database that have a charge imbalance of less than 10 percent, Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

The ranges of water types for USGS-GAMA grid sites and other sites from the CDPH water-quality database were similar (fig. D2). In most water samples from the sites, no single cation accounted for more than 60 percent of the total cations, and bicarbonate accounted for more than 60 percent of the total anions. Waters in these sites are described as *mixed cation–bicarbonate* type waters. Some sites contained *calcium/magnesium–bicarbonate* type waters, for which calcium plus magnesium and bicarbonate accounted for more than

60 percent of the cations and anions, respectively. Waters in a minority of sites were classified as *sodium-chloride* type waters, indicating that sodium and chloride accounted for more than 60 percent of the total cations and anions, respectively.

The determination that the range of relative abundance of major cations and anions in USGS-GAMA grid sites (58 sites) is similar to the range of those in all CDPH sites (60 sites) indicates that the grid sites represent the types of water present in the NOCO study unit.

Appendix E. Additional Water-Quality Data

[Table E1](#) presents the data generated at Lawrence Livermore National Laboratory (LLNL) for dissolved noble gases (argon, helium-4, krypton, neon, and xenon) and helium isotope ratios. These results were not completed in time for inclusion in the U.S. Geological Survey (USGS) Data Series Report for the Northern Coast Ranges (NOCO) study unit (U.S. Geological Survey Data Series 609, Mathany and others, 2011) and are included in this report for completeness.

Table E1. Results from the Lawrence Livermore National Laboratory (LLNL) analyses of dissolved noble gases and helium isotope ratios in samples from U.S. Geological Survey (USGS) GAMA grid sites collected for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.

[The five-digit number in parentheses below the constituent name is the USGS parameter code used to uniquely identify a specific constituent or property. **USGS-GAMA site identification numbers:** NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Measurement errors:** Helium-3/Helium-4 ratios: 1% error; helium-4 and argon concentrations: 2% error; krypton concentrations: 3% error; neon and xenon concentrations: 4% error. **Abbreviations:** g⁻¹ H₂O, pressure per gram of water; R, helium-3/helium-4 ratio of the groundwater sample; Ra, helium-3/helium-4 ratio in the atmosphere; na, not available]

USGS-GAMA site identification number	Helium-4 (g ⁻¹ H ₂ O) (85561) x 10 ⁻⁸	Neon (g ⁻¹ H ₂ O) (61046) x 10 ⁻⁷	Argon (g ⁻¹ H ₂ O) (85563) x 10 ⁻⁴	Krypton (g ⁻¹ H ₂ O) (85565) x 10 ⁻⁸	Xenon (g ⁻¹ H ₂ O) (85567) x 10 ⁻⁸	Helium-3/Helium-4 (atom ratio) (61040) x 10 ⁻⁶	Helium-3/Helium-4 (R/Ra)
Interior Basins study area							
NOCO-IN-01	9.38	2.83	4.18	8.48	1.13	1.10	0.8
NOCO-IN-02	19.75	7.27	7.20	0.13	1.59	1.44	1.0
NOCO-IN-03	5.38	3.49	4.05	8.28	1.16	1.43	1.0
NOCO-IN-04	20.28	7.25	8.46	18.22	2.37	1.37	1.0
NOCO-IN-05	38.48	1.73	8.42	7.34	1.27	0.32	0.2
NOCO-IN-06	13.23	3.58	4.56	9.53	1.25	1.01	0.7
NOCO-IN-07	5.71	2.12	3.60	8.13	1.25	1.24	0.9
NOCO-IN-08	4.88	2.15	3.51	7.38	1.12	1.37	1.0
NOCO-IN-09	18.85	2.51	3.79	8.10	1.14	0.52	0.4
NOCO-IN-10	4.56	2.85	3.61	7.54	1.01	1.39	1.0
NOCO-IN-11	55.56	3.71	3.54	6.52	0.77	0.16	0.1
NOCO-IN-12	4.37	1.80	2.98	6.63	0.95	1.37	1.0
NOCO-IN-13	5.40	2.31	3.78	8.44	1.27	1.39	1.0
NOCO-IN-14	10.84	3.63	4.71	9.53	1.48	1.18	0.8
NOCO-IN-15	13.27	2.75	4.13	8.61	1.33	0.71	0.5
NOCO-IN-16	5.23	2.26	3.41	7.30	1.03	1.40	1.0
NOCO-IN-17	137.4	38.39	9.14	17.19	1.74	1.45	1.0
NOCO-IN-18	8.00	2.72	3.89	8.22	1.14	1.62	1.2
NOCO-IN-19	5.49	2.33	3.74	8.17	1.23	1.39	1.0
NOCO-IN-20	na	na	na	na	na	na	na
NOCO-IN-21	5.54	2.11	3.41	7.73	1.07	1.74	1.2
NOCO-IN-22	65.36	2.65	4.19	7.74	1.04	6.95	5.0
NOCO-IN-23	8.44	3.79	4.37	9.01	1.17	1.44	1.0
NOCO-IN-24	6.86	2.67	3.22	6.65	0.97	1.39	1.0
NOCO-IN-25	25.40	2.37	3.64	7.19	0.99	0.53	0.4
NOCO-IN-26	4.69	2.05	3.25	7.15	0.99	1.29	0.9
NOCO-IN-27	3.87	1.74	1.16	2.78	4.27	1.15	0.8
NOCO-IN-28	49.62	3.34	4.45	9.21	1.29	0.40	0.3

Table E1. Results from the Lawrence Livermore National Laboratory (LLNL) analyses of dissolved noble gases and helium isotope ratios in samples from U.S. Geological Survey (USGS) GAMA grid sites collected for the Northern Coast Ranges study unit, 2009, California GAMA Priority Basin Project.—Continued

[The five-digit number in parentheses below the constituent name is the USGS parameter code used to uniquely identify a specific constituent or property. **USGS-GAMA site identification numbers:** NOCO-IN, Northern Coast Ranges study unit, Interior Basins study area; NOCO-CO, Northern Coast Ranges study unit, Coastal Basins study area. **Measurement errors:** Helium-3/Helium-4 ratios: 1% error; helium-4 and argon concentrations: 2% error; krypton concentrations: 3% error; neon and xenon concentrations: 4% error. **Abbreviations:** $\text{g}^{-1} \text{H}_2\text{O}$, pressure per gram of water; R, helium-3/helium-4 ratio of the groundwater sample; Ra, helium-3/helium-4 ratio in the atmosphere; na, not available]

USGS-GAMA site identification number	Helium-4 ($\text{g}^{-1}\text{H}_2\text{O}$) (85561) $\times 10^{-8}$	Neon ($\text{g}^{-1}\text{H}_2\text{O}$) (61046) $\times 10^{-7}$	Argon ($\text{g}^{-1}\text{H}_2\text{O}$) (85563) $\times 10^{-4}$	Krypton ($\text{g}^{-1}\text{H}_2\text{O}$) (85565) $\times 10^{-8}$	Xenon ($\text{g}^{-1}\text{H}_2\text{O}$) (85567) $\times 10^{-8}$	Helium-3/ Helium-4 (atom ratio) (61040) $\times 10^{-6}$	Helium-3/ Helium-4 (R/Ra)
Coastal Basins study area							
NOCO-CO-01	3.70	1.45	3.30	8.24	1.43	1.21	0.9
NOCO-CO-02	5.15	2.20	3.58	7.93	1.09	1.30	0.9
NOCO-CO-03	5.17	2.22	3.81	8.70	1.22	1.37	1.0
NOCO-CO-04	6.59	2.84	4.46	10.09	1.70	1.26	0.9
NOCO-CO-05	3.70	1.55	3.19	7.34	1.22	1.45	1.0
NOCO-CO-06	24.91	8.06	6.68	11.73	1.44	1.46	1.0
NOCO-CO-07	5.91	2.55	3.96	8.92	1.24	2.13	1.5
NOCO-CO-08	5.95	2.24	3.87	8.42	1.25	1.59	1.1
NOCO-CO-09	76.99	2.41	8.30	7.80	1.44	0.24	0.2
NOCO-CO-10	6.37	2.93	4.36	9.42	1.27	1.63	1.2
NOCO-CO-11	10.53	4.25	4.72	9.76	1.30	1.42	1.0
NOCO-CO-12	4.80	2.15	3.67	8.67	1.20	1.37	1.0
NOCO-CO-13	5.76	2.09	3.62	8.43	1.09	1.20	0.9
NOCO-CO-14	na	na	na	na	na	na	na
NOCO-CO-15	5.17	2.30	3.72	8.46	1.18	1.30	0.9
NOCO-CO-16	5.63	2.86	4.00	9.24	1.23	1.40	1.0
NOCO-CO-17	4.77	2.08	3.52	7.92	1.28	1.36	1.0
NOCO-CO-18	5.61	2.35	3.93	8.76	1.35	1.39	1.0
NOCO-CO-19	6.16	2.62	4.32	9.42	1.54	1.40	1.0
NOCO-CO-20	4.42	1.99	3.67	8.37	1.21	1.39	1.0
NOCO-CO-21	6.81	2.58	4.85	7.36	1.31	1.31	0.9
NOCO-CO-22	6.66	2.60	4.28	9.33	1.41	1.37	1.0
NOCO-CO-23	4.54	1.97	3.43	7.41	1.04	1.41	1.0
NOCO-CO-24	72.59	2.70	4.66	9.15	1.45	0.35	0.3
NOCO-CO-25	40.69	2.38	4.20	9.27	1.37	0.41	0.3
							0.0
NOCO-CO-26	5.05	2.11	3.87	9.03	1.38	1.36	1.0
NOCO-CO-27	23.28	2.36	8.75	8.47	1.66	0.43	0.3
NOCO-CO-28	na	na	na	na	na	na	na
NOCO-CO-29	4.27	4.06	4.62	2.97	0.37	2.64	1.9
NOCO-CO-30	6.19	3.24	4.14	7.78	0.88	1.40	1.0

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